## Upper Columbia River

# FINAL Soil Study Data Summary Report 

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## CONTENTS

LIST OF FIGURES ..... vi
LIST OF MAPS ..... ix
LIST OF TABLES ..... x
ACRONYMS AND ABBREVIATIONS ..... xiii
UNITS OF MEASURE ..... xv
1 INTRODUCTION ..... 1-1
1.1 BACKGROUND ..... 1-1
1.2 REPORT ORGANIZATION ..... 1-1
22014 SOIL STUDY DESIGN ..... 2-1
2.1 PURPOSE OF STUDY ..... 2-1
2.2 DATA QUALITY OBJECTIVES ..... 2-1
2.2.1 Step 1 - State the Problem ..... 2-1
2.2.2 Step 2 - Identify the Goals of the Study ..... 2-2
2.2.3 Step 3 - Identify Information Inputs ..... 2-3
2.2.4 Step 4 - Define the Boundaries of the Study ..... 2-4
2.2.5 Step 5 - Develop the Analytical Approach ..... 2-5
2.2.6 Step 6 - Specify Performance or Acceptance Criteria ..... 2-7
2.2.7 Step 7 - Develop the Plan for Obtaining the Data ..... 2-9
2.3 STUDY DESIGN ..... 2-9
2.3.1 Overall Design ..... 2-9
2.3.2 Selection of Sampling Areas ..... 2-10
2.3.3 Identification of Target DUs and Sampling Locations ..... 2-11
3 METHODS ..... 3-1
3.1 FIELD METHODS ..... 3-1
3.1.1 Sampling Locations ..... 3-1
3.1.2 Methods for Sample Collection ..... 3-2
3.1.3 Field Changes and Deviations ..... 3-4
3.2 LABORATORY METHODS ..... 3-5
3.2.1 Methods for Chemical Analysis ..... 3-6
3.2.2 Laboratory Deviations ..... 3-6
3.3 DATA EVALUATION APPROACH ..... 3-7
3.3.1 Methods ..... 3-7
3.3.2 Deviations from Planned Data Evaluation Approach ..... 3-8
4 VALIDATION ASSESSMENT ..... 4-1
4.1 OVERALL DATA QUALITY ..... 4-1
4.2 SAMPLE TRANSPORT AND HOLDING TIMES ..... 4-2
4.3 METALS ..... 4-3
4.3.1 Calibration ..... 4-3
4.3.2 Blanks ..... 4-3
4.3.3 Matrix Spikes ..... 4-3
4.3.4 Laboratory Control Samples ..... 4-3
4.3.5 Laboratory Duplicates, Field Split Samples, and Triplicate Samples ..... 4-4
4.3.6 Interference Check Samples ..... 4-4
4.3.7 Serial Dilutions ..... 4-4
4.3.8 Internal Standards ..... 4-4
4.4 CONVENTIONAL PARAMETERS ..... 4-4
4.4.1 Laboratory and Field Duplicates and Triplicates ..... 4-4
4.5 IVBA ..... 4-5
4.5.1 Calibration ..... 4-5
4.5.2 Blanks ..... 4-5
4.5.3 Matrix Spikes ..... 4-5
4.5.4 Laboratory Control Samples and Standard Reference Material ..... 4-6
4.5.5 Laboratory Duplicates and Field Split Samples ..... 4-6
4.5.6 Interference Check Samples ..... 4-6
4.5.7 Serial Dilutions ..... 4-6
4.5.8 Internal Standards ..... 4-6
5 RESULTS ..... 5-1
5.1 AERIAL DEPOSITION AREAS ..... 5-2
5.1.1 Metals and Conventional Parameters ..... 5-2
5.1.2 IVBA ..... 5-2
5.2 RELICT FLOODPLAIN DEPOSITION AREAS ..... 5-2
5.2.1 Metals and Conventional Parameters ..... 5-3
5.2.2 IVBA ..... 5-3
5.3 WINDBLOWN SEDIMENT DEPOSITION AREAS ..... 5-3
5.3.1 Metals and Conventional Parameters ..... 5-3
5.3.2 IVBA ..... 5-3
5.4 FIELD QC SUMMARY ..... 5-3
5.4.1 Aerial Deposition Areas ..... 5-4
5.4.2 Relict Floodplain Deposition Areas ..... 5-4
5.4.3 Windblown Sediment Deposition Areas ..... 5-5
5.5 EVALUATION OF REPORTING LIMITS FOR NONDETECTED SAMPLES ..... 5-5
5.6 COMPARISON WITH SCREENING LEVELS ..... 5-5
5.6.1 Ecological Screening Levels ..... 5-6
5.6.2 Human Health Screening Levels. ..... 5-7
6 SUMMARY AND RECOMMENDATIONS ..... 6-1
7 REFERENCES ..... 7-1
Appendix A Field Activity ReportAppendix B. Decision Memorandum for IVBA Samples

## LIST OF FIGURES

Figure 5-1a pH in Bulk Soil Samples by Deposition Area
Figure 5-1b Percent Solids in Bulk Soil Samples by Deposition Area
Figure 5-1c Total Fines in Bulk Soil Samples by Deposition Area
Figure 5-2a Percent Solids in < 2-mm Soil Fractions by Deposition Area
Figure 5-2b Organic Carbon in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area
Figure 5-2c Cation Exchange Capacity in < 2-mm Soil Fractions by Deposition Area

Figure 5-3 Percent solids in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area
Figure 5-4a Aluminum Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4b Antimony Concentrations in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area

Figure 5-4c Arsenic Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4d Barium Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4e Beryllium Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4f Cadmium Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure $5-4 \mathrm{~g} \quad$ Calcium Concentrations in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area

Figure 5-4h Chromium Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4i Cobalt Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4j Copper Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4k Iron Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4l Lead Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4m Magnesium Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4n Manganese Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4o Mercury Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4p Molybdenum Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure $5-4 \mathrm{q} \quad$ Nickel Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4r Potassium Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4s Selenium Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4t Silver Concentrations in < 2-mm Soil Fractions by
Deposition Area
Figure 5-4u Sodium Concentrations in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area

Figure 5-4v Thallium Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4w Vanadium Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-4x Zinc Concentrations in < 2-mm Soil Fractions by Deposition Area

Figure 5-5a Aluminum Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5b Antimony Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5c Arsenic Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5d Barium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5e Beryllium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5f Cadmium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure $5-5 \mathrm{~g} \quad$ Calcium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5h Chromium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5i Cobalt Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5j Copper Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5k Iron Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5l Lead Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5m Magnesium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5n Manganese Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5o Mercury Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5p Molybdenum Concentrations in $<149-\mu m$ Soil Fractions by Deposition Area

Figure 5-5q Nickel Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by
Deposition Area
Figure 5-5r Potassium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5s Selenium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by
Deposition Area

Figure 5-5t Silver Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5u Sodium Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5v Thallium Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5w Vanadium Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-5x Zinc Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6a Percent Bioaccessible Aluminum in $<149-\mu m$ Soil Fractions by Deposition Area

Figure 5-6b Percent Bioaccessible Antimony in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6c Percent Bioaccessible Arsenic in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6d Percent Bioaccessible Barium in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6e Percent Bioaccessible Beryllium in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6f Percent Bioaccessible Cadmium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6g Percent Bioaccessible Calcium in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6h Percent Bioaccessible Chromium in < $149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6i Percent Bioaccessible Cobalt in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6j Percent Bioaccessible Copper in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6k Percent Bioaccessible Iron in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-61 Percent Bioaccessible Lead in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6m Percent Bioaccessible Magnesium in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6n Percent Bioaccessible Manganese in < 149- $\mu$ m Soil Fractions by Deposition Area

Figure 5-6o Percent Bioaccessible Mercury in < 149- $\mu$ m Soil Fractions by Deposition Area

Figure 5-6p Percent Bioaccessible Molybdenum in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6q Percent Bioaccessible Nickel in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6r Percent Bioaccessible Potassium in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6s Percent Bioaccessible Selenium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6t Percent Bioaccessible Silver in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6u Percent Bioaccessible Sodium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6v Percent Bioaccessible Thallium in < 149- $\mu$ m Soil Fractions by Deposition Area

Figure 5-6w Percent Bioaccessible Vanadium in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

Figure 5-6x Percent Bioaccessible Zinc in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area

## LIST OF MAPS

## Map 2-1 Soil Study Areas

Map 2-2 Decision Unit Locations in the Aerial Deposition Areas
Map 2-3 Decision Unit Locations in the Relict Floodplain Deposition Areas
Map 2-4 Decision Unit Locations in the Windblown Sediment Deposition Areas
Map 5-1 Spatial Distribution of Arsenic Concentrations in the < 2-mm Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-2 Spatial Distribution of Barium Concentrations in the < 2-mm Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-3 Spatial Distribution of Chromium Concentrations in the $<2-\mathrm{mm}$ Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-4 Spatial Distribution of Cobalt Concentrations in the $<2-\mathrm{mm}$ Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-5 Spatial Distribution of Copper Concentrations in the < 2-mm Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-6 Spatial Distribution of Manganese Concentrations in the < 2-mm Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-7 Spatial Distribution of Nickel Concentrations in the < 2-mm Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-8 Spatial Distribution of Selenium Concentrations in the $<2-\mathrm{mm}$ Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-9 Spatial Distribution of Arsenic Concentrations in the < 2-mm Fraction of Soil Collected from the Relict Floodplain Deposition Areas

Map 5-10 Spatial Distribution of Barium Concentrations in the $<2-\mathrm{mm}$ Fraction of Soil Collected from the Relict Floodplain Deposition Areas

Map 5-11 Spatial Distribution of Chromium Concentrations in the $<2-\mathrm{mm}$ Fraction of Soil Collected from the Relict Floodplain Deposition Areas

Map 5-12 Spatial Distribution of Cobalt Concentrations in the < 2-mm Fraction of Soil Collected from the Relict Floodplain Deposition Areas

Map 5-13 Spatial Distribution of Copper Concentrations in the < 2-mm Fraction of Soil Collected from the Relict Floodplain Deposition Areas

Map 5-14 Spatial Distribution of Selenium Concentrations in the $<2-\mathrm{mm}$ Fraction of Soil Collected from the Relict Floodplain Deposition Areas

Map 5-15 Spatial Distribution of Antimony Concentrations in the $<2$-mm Fraction of Soil Collected from the Windblown Sediment Deposition Areas

Map 5-16 Spatial Distribution of Cadmium Concentrations in the $<2-\mathrm{mm}$ Fraction of Soil Collected from the Windblown Sediment Deposition Areas

Map 5-17 Spatial Distribution of Lead Concentrations in the < 2-mm Fraction of Soil Collected from the Windblown Sediment Deposition Areas

Map 5-18 Spatial Distribution of Zinc Concentrations in the $<2-\mathrm{mm}$ Fraction of Soil Collected from the Windblown Sediment Deposition Areas

Map 5-19 Spatial Distribution of Arsenic Concentrations in the $<149-\mu$ m Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-20 Spatial Distribution of Lead Concentrations in the < 149- $\mu$ m Fraction of Soil Collected from the Aerial Deposition Areas

Map 5-21 Spatial Distribution of Arsenic Concentrations in the $<149-\mu$ m Fraction of Soil Collected from the Relict Floodplain Deposition Areas

Map 5-22 Spatial Distribution of Lead Concentrations in the < 149- $\mu$ m Fraction of Soil Collected from the Relict Floodplain Deposition Areas

## LIST OF TABLES

Table 2-1 Target Method Detection Limits and Method Reporting Limits
Table 2-2
Table 3-1
Table 3-2
Table 3-3a
Table 3-3b
Table 3-3c

Table 3-4 Soil Sample Analysis Summary
Table 3-5 Analytical Methods for Soil Samples
Table 4-1a Aerial Deposition Area Summary of Qualifiers for Bulk Soil Sample Conventional Parameter Results

Table 4-1b Aerial Deposition Area Summary of Qualifiers for < 2-mm Fraction Metals and Conventional Parameter Results

Table 4-1c Aerial Deposition Area Summary of Qualifiers for $<149-\mu \mathrm{m}$ Fraction Metals and Conventional Parameter Results

Table 4-2a

Table 4-2b

Table 4-2c

Table 4-3a

Table 4-3b

Table 4-4
Table 5-1a
Relict Floodplain Deposition Area Summary of Qualifiers for Bulk Soil Sample Conventional Parameter Results

Relict Floodplain Deposition Area Summary of Qualifiers for < 2-mm Fraction Metals and Conventional Parameter Results

Relict Floodplain Deposition Area Summary of Qualifiers for $<149-\mu \mathrm{m}$ Fraction Metals and Conventional Parameter Results

Windblown Sediment Deposition Area Summary of Qualifiers for Bulk Soil Sample Conventional Parameter Results

Windblown Sediment Deposition Area Summary of Qualifiers for <2-mm Fraction Metals and Conventional Parameter Results

Summary of Qualifiers for IVBA Results
Aerial Deposition Area Summary Statistics for Bulk Soil Sample Conventional Parameter Results

Table 5-1b Aerial Deposition Area Summary Statistics for < 2-mm Fraction Metals and Conventional Parameter Results

Table 5-1c Aerial Deposition Area Summary Statistics for < 149- $\mu \mathrm{m}$ Fraction Metals and Conventional Parameter Results

Table 5-2a

Table 5-2b

Table 5-2c

Table 5-3a

Table 5-3b

Table 5-4
Table 5-5
Table 5-6
Table 5-7a

Table 5-7b

Table 5-7c

Table 5-8a

Table 5-8b

Table 5-8c

Table 5-9a

Relict Floodplain Deposition Area Summary Statistics for Bulk Soil Sample Conventional Parameter Results

Relict Floodplain Deposition Area Summary Statistics for < 2-mm Fraction Metals and Conventional Parameter Results

Relict Floodplain Deposition Area Summary Statistics for < $149-\mu \mathrm{m}$ Fraction Metals and Conventional Parameter Results

Windblown Sediment Deposition Area Summary Statistics for Bulk Soil Sample Conventional Parameter Results

Windblown Sediment Deposition Area Summary Statistics for $<2-\mathrm{mm}$ Fraction Metals and Conventional Parameter Results

Summary Statistics for IVBA Results
Relative Bioavailability (RBA) Data for Lead from the $<149-\mu$ m Fraction
Lead Data from < 149- $\mu$ m Fraction Adjusted for Bioavailability
Aerial Deposition Area Summary of Field Split and Triplicate Sample Results for Bulk Soil Samples

Aerial Deposition Area Summary of Field Split and Triplicate Sample Results for the < 2-mm Soil Fraction

Aerial Deposition Area Summary of Field Split and Triplicate Sample Results for the < 149- $\mu$ m Soil Fraction

Relict Floodplain Deposition Area Summary of Field Split and Triplicate Sample Results for Bulk Soil Samples

Relict Floodplain Deposition Area Summary of Field Split and Triplicate Sample Results for the < 2-mm Soil Fraction

Relict Floodplain Deposition Area Summary of Field Split and Triplicate Sample Results for the $<149-\mu \mathrm{m}$ Soil Fraction

Windblown Sediment Deposition Area Summary of Field Split and Triplicate Sample Results for Bulk Soil Samples

Table 5-9b Windblown Sediment Deposition Area Summary of Field Split and Triplicate Sample Results for the < 2-mm Soil Fraction

Table 5-10 Comparison of Actual Method Reporting Limits with Analytical Concentration Goals for Nondetected Samples

Table 5-11a Summary of Metals Data Compared with Available Eco-SSLs
Table 5-11b Comparison of Aerial Deposition Area Metals Data from the $<2-\mathrm{mm}$ Fraction with Available Eco-SSLs

Table 5-11c Comparison of Relict Floodplain Deposition Area Metals Data from the <2-mm Fraction with Available Eco-SSLs

Table 5-11d Comparison of Windblown Sediment Deposition Area Metals Data from the $<2-\mathrm{mm}$ Fraction with Available Eco-SSLs

Table 5-12a Summary of Metals Data from < $149-\mu \mathrm{m}$ Fraction Compared with Available Human Health Screening Levels

Table 5-12b Comparison of Aerial Deposition Area Metals Data from < $149-\mu \mathrm{m}$ Fraction with Available Human Health Screening Levels

Table 5-12c Comparison of Relict Floodplain Deposition Area Metals Data from $<149-\mu \mathrm{m}$ Fraction with Available Human Health Screening Levels

## ACRONYMS AND ABBREVIATIONS

| ACG | analytical concentration goal |
| :--- | :--- |
| AD | aerial deposition |
| ADA | aerial deposition area |
| ALS | ALS Environmental |
| ARCADIS | ARCADIS U.S., Inc. |
| BERA | baseline ecological risk assessment |
| CEC | cation exchange capacity |
| COC | chain-of-custody |
| CSM | conceptual site model |
| DL | detection limit |
| DQO | data quality objective |
| DU | ecological soil screening level |
| Eco-SSL | Environmental Standards, Inc. |
| ESI | U.S. Environmental Protection Agency |
| EPA | field sampling plan |
| FSP | quality assurance project plan |
| GPS | human health risk assessment |
| HHRA | inductively coupled plasma |
| ICP | incremental composite sampling |
| ICS | identification |
| ID | Interstate Technology \& Regulatory Council |
| ITRC | in vitro bioaccessibility assay |
| IVBA | method detection limit |
| LCS | meratory control sample reporting limit |
| MDL | MRL |


| RBA | relative bioavailability |
| :--- | :--- |
| RBC | risk-based concentration |
| RF | relict floodplain |
| RFDA | relict floodplain deposition area |
| RI/FS | remedial investigation and feasibility study |
| RL | reporting limit |
| RM | relative percent difference |
| RPD | relative standard deviation |
| RSD | Syracuse Research Corporation |
| SOP | target analyte list |
| SRC | Teck American Incorporated |
| TAL | total organic carbon |
| TAI | Upper Columbia River |
| TOC | U.S. Bureau of Reclamation |
| UCR | U.S. Geological Survey |
| USBR | windblown sediment |
| USGS | wSolown sediment deposition area |
| WSDA | Recedure |

## UNITS OF MEASURE

| ac | acre(s) |
| :--- | :--- |
| cm | centimeter(s) |
| g | gram(s) |
| gal. | gallon(s) |
| in. | inch(es) |
| m | meter(s) |
| $\mathrm{mg} / \mathrm{kg}$ | milligrams per kilogram |
| mm | millimeter(s) |
| $\mu \mathrm{m}$ | micrometer(s) |
| $\mathrm{mi}^{2}$ | square mile(s) |

## 1 INTRODUCTION

This report presents the results for the 2014 Soil Study (herein referred to as "the study") conducted for the Upper Columbia River (UCR) Site, herein referred to as the Site. ${ }^{1}$ Analyses were conducted under the U.S. Environmental Protection Agency (EPA)approved quality assurance project plan (QAPP) for the study (Exponent et al. 2014). The study was conducted as part of the remedial investigation and feasibility study (RI/FS) for the Site to evaluate if there is unacceptable risk to ecological receptors and people from exposure to metals in the upland soils. Data needs addressed by this study are intended to support the conduct of the baseline ecological risk assessment (BERA) (to be completed by Teck American Incorporated [TAI]) and the baseline human health risk assessment (HHRA) (to be completed by EPA).

### 1.1 BACKGROUND

A review of historical soil data from soil samples collected adjacent to the UCR identified gaps in the data necessary to evaluate ecological and human health risks (Exponent et al. 2014). These data gaps included the need for additional soil data to evaluate upland areas potentially affected by point sources (e.g., aerial deposition of smelter particulates), historical fluvial deposition of sediment onto relict floodplains, and re-deposition of windblown sediment. In addition, historical soil data sets did not include parameters needed to determine the bioavailability of metals to soil organisms (i.e., cation exchange capacity [CEC], total organic carbon [TOC], and pH ) or the bioaccessibility of lead in soil to which people may be exposed (i.e., in vitro bioaccessibility assay [IVBA]) (USEPA 2007a).

### 1.2 REPORT ORGANIZATION

This report is organized into the following sections:

- Section 1 - Introduction. This section provides background information for the study and outlines the report organization.
${ }^{1}$ The Site, as defined in the June 2, 2006, Settlement Agreement (USEPA 2006b), is "the areal extent of hazardous substances contamination within the United States in or adjacent to the Upper Columbia River, including the Franklin D. Roosevelt Lake ("Lake Roosevelt"), from the border between the United States and Canada downstream to the Grand Coulee Dam, and all suitable areas in proximity to such contamination necessary for implementation of the response actions...."
- Section 2 - Study Design. This section describes the purpose and objectives of the study and provides an overview of the study design.
- Section 3 - Methods. This section provides the methods used for the study, including target sampling locations, collection methods, field analyses, laboratory analyses, and the approach used to summarize the data for the data report. This section also discusses any changes or deviations from the QAPP and field sampling plan (FSP).
- Section 4 - Validation Assessment. This section provides an overview of the validation assessment conducted for the analytical results of the study samples.
- Section 5 - Results. This section presents a summary of the field and analytical results and provides a comparison of the results with soil screening levels.
- Section 6 - Summary and Recommendations. This section presents a summary of the results and provides recommendations.
- Section 7 - References. This section presents bibliographic information for the documents cited within this report.

Figures, maps, and data tables are provided following Section 7. Appendices and the raw data are provided in electronic format (see enclosed CD-ROM). Data may also be obtained directly from the project database, accessible at: http://teck-ucr.exponent.com.

## 22014 SOIL STUDY DESIGN

### 2.1 PURPOSE OF STUDY

The purpose of the study was to collect information on concentrations of analytes in upland soil. This information will be used to evaluate whether there is unacceptable risk to ecological and human receptors from exposure to metals in the upland soil adjacent to the UCR (Exponent et al. 2014).

### 2.2 DATA QUALITY OBJECTIVES

As described in the QAPP (Exponent et al. 2014), EPA's seven-step data quality objective (DQO) process (USEPA 2006a) was used to guide the study design for the collection of upland soil. The DQO process is used to determine the type, quantity, and quality of data needed to achieve study goals and establishes performance and acceptance criteria for the data. The seven steps of the DQO process are listed below and discussed in the subsections that follow.

1. State the problem
2. Identify the goals of the study
3. Identify information inputs
4. Define the boundaries of the study
5. Develop the analytical approach
6. Specify performance or acceptance criteria
7. Develop the plan for obtaining the data.

### 2.2.1 Step 1 - State the Problem

The preliminary conceptual site model (CSM) for the UCR RI/FS (Parametrix et al. 2010) identified soil as a potential exposure pathway for ecological receptors (i.e., terrestrial invertebrates, amphibians, reptiles, plants, birds, and mammals) and human receptors. As presented in the QAPP (Exponent et al. 2014), the spatial extent of historical soil data was limited and considered to be insufficient to evaluate potential risks to ecological and human receptors. Therefore, additional soil data were collected during the study to evaluate three types of areas, representing different soil transport mechanisms, where concentrations of analytes in soil might present an unacceptable risk to ecological and human receptors.

These three types of areas are:

- Aerial deposition areas (ADAs)-areas potentially influenced by smelter particulate deposition in the northern portion of the Site
- Relict floodplain deposition areas (RFDAs) - areas that may have been inundated under historical hydraulic conditions in the UCR
- Windblown sediment deposition areas (WSDAs) - areas where sediment from the UCR shoreline may have been transported by the wind during periods of water drawdown.

At the direction of EPA, the analytes evaluated in this study included EPA's target analyte list (TAL) metals ${ }^{2}$ and molybdenum (USEPA 2012a). In addition, data that were not available in the historical data sets but were needed to determine the bioavailability of metals to soil organisms (i.e., CEC, TOC, and pH ) and the bioaccessibility of lead in soil to which people may be exposed (i.e., IVBA) were also evaluated in the study. ${ }^{3}$

### 2.2.2 Step 2 - Identify the Goals of the Study

The primary goal of the study was to generate data to be used in characterizing the exposure of ecological and human receptors to upland soil. As presented in the QAPP (Exponent et al. 2014), the primary questions developed to meet this goal were:

- Where have analyte concentrations in soil been influenced by the deposition of particulates in air emissions, sediment onto relict flood plains adjacent to the UCR, or re-deposition of windblown sediment?
- What are the concentrations of analytes in soil that have potentially been influenced by the deposition of particulates in air emissions, sediment onto relict floodplains adjacent to the UCR, or re-deposition of windblown sediment?
- Do analyte concentrations in soil pose an unacceptable risk to ecological or human receptors?
${ }^{2}$ TAL metals include aluminum, antimony, arsenic, barium, beryllium, cadmium, calcium, chromium, cobalt, copper, iron, lead, magnesium, manganese, mercury, nickel, potassium, selenium, silver, sodium, thallium, vanadium, and zinc (USEPA 2015) (see http://www.epa.gov/superfund/programs/clp/ismtarget.htm).
${ }^{3}$ At the direction of EPA, IVBA was expanded to include TAL metals and molybdenum rather than just lead as specified in the QAPP (Exponent et al. 2014). Documentation regarding the rationale for this decision is provided in Appendix $B$.

The specific risk questions that will be addressed in the BERA and HHRA using the data collected in this study include:

- Is reproduction, growth, or survival of terrestrial invertebrates or plants adversely affected by chemicals of potential concern (COPCs ${ }^{4}$ ) in UCR soil?
- Are COPCs in UCR soil at concentrations that will adversely affect reproduction, growth, or survival of amphibians or reptiles (herpetofauna) during adult life stages?
- Are COPCs in UCR soil at concentrations that will adversely affect reproduction, growth, or survival of terrestrial birds or mammals?
- Is the health of people working, recreating, or living on the Site adversely affected by COPCs in UCR soil?


### 2.2.3 Step 3 - Identify Information Inputs

Careful consideration was given to identify the types and sources of information needed to determine whether exposure to soil at the Site poses unacceptable risks to ecological or human receptors. This information included the following:

- Analytical data for TAL metals and molybdenum in the $<2-\mathrm{mm}$ fraction of soil collected from the three area types (i.e., ADAs, RFDAs, and WSDAs) for use in the BERA
- Analytical data for TAL metals and molybdenum in the $<149-\mu \mathrm{m}$ fraction of soil ${ }^{5}$ collected from two area types (i.e., ADAs and RFDAs) for use in the HHRA ${ }^{6}$
- Grain size distribution in bulk soil samples

[^0]- Soil data for geochemical parameters that could affect the bioavailability of metals to ecological receptors (i.e., pH [bulk soil samples ${ }^{7}$ ] and TOC and CEC [< 2-mm soil fraction]) from the three area types (i.e., ADAs, RFDAs, and WSDAs)
- IVBA analyses on 20 percent of the soil samples from two area types (i.e., ADAs and RFDAs) that had lead concentrations $>100 \mathrm{mg} / \mathrm{kg}$
- Soil screening levels (i.e., ecological soil screening levels [Eco-SSLs] and human [i.e., residential] soil screening levels).


### 2.2.4 Step 4 - Define the Boundaries of the Study

The Site encompasses the UCR from the U.S.-Canada border (river mile [RM] 745) to the Grand Coulee Dam (approximately RM 596) and includes Franklin D. Roosevelt Lake. As discussed in Step 1, three types of areas within the Site were selected to represent three different means of soil deposition: aerial deposition from smelter stacks (ADAs), sediment deposition on relict floodplains (RFDAs), and re-deposition of windblown sediment during low water (drawdown) conditions (WSDAs). These areas are shown on Map 2-1 and described in more detail, below.

- Aerial Deposition Areas. ADAs are areas located within the northernmost $100 \mathrm{mi}^{2}$ of the Site and extend south from the U.S.-Canada border. The eastern border follows the ridge east of the UCR toward Northport, Washington. The western border is bisected by drainages and extends northwesterly along a straight line to the U.S.-Canada border.
- Relict Floodplain Deposition Areas. For the purpose of the study, relict floodplains are areas that may have been subjected to flooding under past flow conditions but are not expected to be inundated under current pool level management controls. The relict floodplain is the delineated area between high-pool seasonally inundated lands and the maximum pre-1973 strandline. The five largest UCR relict floodplains are located near Northport, Washington. All five of these floodplains were designated as RFDAs for the study.
- Windblown Sediment Deposition Areas. Areas on the west bank of northern Marcus Flats and just north of Seven Bays at Columbia Beach were identified as WSDAs because they represent locations where the windblown re-deposition of

[^1]sediment is most likely to occur and represent a possible worst-case scenario. Sampling locations were selected based on site-specific wind dynamics and further refined to avoid areas adjacent to nearby mining activities.

In addition to spatial factors, the timing for sampling was also considered. The only temporal consideration for the soil sampling event was that the soil needed to be accessible (i.e., not snow-covered or frozen). Thus, the targeted time frame for sampling was between June and October.

### 2.2.5 Step 5 - Develop the Analytical Approach

Except for the soil screening levels, all of the information listed in Step 3 was collected as part of the study. The Eco-SSLs presented in the QAPP (Exponent et al. 2014) are the lowest of the screening levels adopted by EPA for plants, soil invertebrates, birds, and mammals (USEPA 2010a). ${ }^{8}$ The human health soil screening levels presented in the QAPP were derived by Syracuse Research Corporation (SRC 2013) and represent residential riskbased screening levels for soil. With the exception of antimony, arsenic, and mercury, all of the SRC (2013) screening levels were calculated using EPA's regional screening level calculator and default values. ${ }^{9}$ SRC (2013) adjusted the screening levels for antimony and mercury to reflect changes to the default reference dose values for those metals. SRC (2013) also adjusted the human health screening level for arsenic for natural background. ${ }^{10}$ The target method detection limits (MDLs) and method reporting limits (MRLs) (Table 2-1) for analytical methods selected for the study were five-fold lower than the EcoSSLs and the human health soil screening levels.

The analytical approaches adopted for the study are intended to provide quality data for evaluating whether there are unacceptable risks to ecological or human receptors exposed to soil from the Site. Soil samples were collected using incremental composite sampling
${ }^{8}$ According to EPA (USEPA 2010a), the metals with Eco-SSLs are those that typically exist as cationic species.
${ }^{9}$ The comparison with screening values provided in this data report are for screening purposes only and are only intended to identify chemicals that should be evaluated in the risk assessments. They do not represent cleanup or action levels (USEPA 2002, 2003).
${ }^{10}$ The human health screening level for arsenic was based on the 2012 default arsenic residential soil screening level for a 1-in-1-million risk level (USEPA 2012b) plus an estimate of the concentration of arsenic in natural background ( $9 \mathrm{mg} / \mathrm{kg}$ ). Since SRC's development of this screening level, EPA's 2012 default arsenic screening level ( $0.39 \mathrm{mg} / \mathrm{kg}$ ) has been updated to include a default oral relative bioavailability assumption of 0.6 or 60 percent for arsenic in soil. The current default arsenic screening level is $0.67 \mathrm{mg} / \mathrm{kg}$. However, for the purpose of this report, the screening value identified in the QAPP was used.
(ICS) methods (ITRC 2012) within specifically selected decision units (DUs). ICS methods are designed to reduce variability in the data and provide more accurate estimates of the mean soil concentrations to which ecological or human receptors are exposed than those obtained from single, discrete soil samples.

Soil samples collected using the ICS method were sieved into two fractions (<2-mm and $<149-\mu \mathrm{m})$. Analytical data from the $<2-\mathrm{mm}$ and $<149-\mu \mathrm{m}$ fractions will be used for the evaluation of risk to ecological and human receptors, respectively.

The data were compared with soil screening levels to determine if additional data are needed. The Eco-SSLs and human health soil screening levels used in this screening process are provided in Section 5.6; they are conservative screening levels to be used only for initial screening purposes and are only intended to identify chemicals for further evaluation in the risk assessments. In the risk assessments, the site-specific bioavailability of metals to ecological receptors in the $<2-\mathrm{mm}$ fraction will be determined using the relationships among $\mathrm{pH}, \mathrm{CEC}$, and TOC, which affect the ability of organisms to take up metals from soils (e.g., Smolders et al. 2009; Checkai et al. 2014). Site-specific adjustments may be made for copper, nickel, zinc, cobalt, and molybdenum using pH, CEC, and TOC. ${ }^{11}$ In addition, IVBA results from the $<149-\mu \mathrm{m}$ fraction may be used to calculate relative bioavailability (RBA) values for metals in the < 2-mm fraction. ${ }^{12}$ The RBA adjustment for the $<2-\mathrm{mm}$ fraction will be conducted in the BERA.

IVBA results for lead in the $<149-\mu \mathrm{m}$ fraction have been used to calculate site-specific oral RBA values for lead in soil. In addition, arsenic concentrations in the $<149-\mu \mathrm{m}$ fraction have been adjusted for EPA's default RBA of 60 percent arsenic in soil (USEPA 2012b). Spatial representations of results for metals concentrations in the $<149-\mu \mathrm{m}$ fraction have been prepared to identify locations where concentrations exceed human health risk-based concentrations (RBCs). ${ }^{13}$ The spatial evaluation of data from the $<2-\mathrm{mm}$ fraction will be conducted as part of the BERA, and data may be adjusted for bioavailability. Locations

[^2]exceeding screening levels will be compared against regional background concentrations. ${ }^{14}$

### 2.2.6 Step 6 - Specify Performance or Acceptance Criteria

Performance or acceptance criteria are derived to minimize the possibility of either making erroneous conclusions or failing to keep uncertainty regarding the estimates to within acceptable levels.

The sampling goal was to collect 100 percent of the targeted samples. Reserve DUs were established to mitigate sample collection challenges in specific areas (e.g., impassable roads, flooding, rocky outcrops, cultural resource issues, lack of landowner permission, steep terrain, or erosion). For sampling challenges within DUs, procedures were established to shift increment locations if it was not possible to sample a targeted increment location listed in the QAPP.

Soil sampling and analysis were conducted using standard EPA-approved methods. DQOs followed EPA guidelines for precision, accuracy, representativeness, completeness, comparability, and analytical sensitivity. Composite samples were submitted to the analytical laboratory for ICS processing and subsampling according to the Interstate Technology \& Regulatory Council (ITRC) guidance (ITRC 2012) and EPA-approved laboratory standard operating procedures (SOPs). Composited samples were homogenized, sieved into fractions ( $<2-\mathrm{mm}$ and $<149-\mu \mathrm{m}$ ), and subsampled at the analytical laboratory. Metals analysis was conducted on a 2 g subsample, which is greater than the standard mass of 1 g (dry weight) for EPA Method 3050B but the minimum mass required to obtain a representative sample using ICS methods (Crumbling 2014).

Field quality control included the following: ${ }^{15}$

- Triplicate samples were collected from a total of 22 DUs. A detailed breakdown of triplicate samples by area is described below.
- Primary ADA (includes reserve DUs in the reserve sampling area) - 10 DUs were sampled in triplicate in the primary ADA (8 in the main area and 2 in

[^3]the reserve area). Of the 107 DUs sampled in the primary ADA, slightly less than 10 percent were sampled in triplicate (i.e., 10 of 107 or 9.4 percent), which met the QAPP requirement. This requirement specified that slightly under 10 percent of the DUs from the primary ADA be collected in triplicate because several additional sampling DUs were added at the request of EPA without the stipulation that additional triplicate locations be added to meet the 10 percent minimum criterion.

- High-density ADA - Six DUs were sampled in triplicate in the high-density ADA. Of the 35 DUs sampled in the high-density ADA, 17 percent were sampled in triplicate, which met the QAPP requirement that at least 10 percent of the DUs from the high-density ADA be collected in triplicate.
- RFDA-4 DUs were sampled in triplicate in the RFDA, one from each of the 4 RFDAs sampled (RFA, RFB, RFC, and RFD) as specified in the QAPP. Of the 16 DUs sampled in the RFDA, 25 percent were sampled in triplicate, which is above the minimum QAPP requirement that at least 10 percent of the DUs from the RFDA be collected in triplicate.
- WSDA - 2 DUs were sampled in triplicate in the WSDA: one DU in the Columbia Beach North deposition area and one DU in the Marcus Flats East deposition area. Of the 13 DUs sampled in the WSDA, 15 percent were sampled in triplicate, which is above the minimum QAPP requirement that at least 10 percent of the DUs from the WSDA be collected in triplicate.
- Two types of split samples were prepared. These samples were collected in the same manner as standard samples in accordance with the QAPP and as summarized below:
- Field split samples were pre-selected for certain DUs to assess the homogeneity of samples collected in the field. ALS Environmental (ALS) performed sample homogenization and took two aliquots of sample from the homogenized soil to generate the field split samples. Field split samples were prepared from 10 percent of the collected samples.
- EPA split samples were pre-selected by EPA representatives for chemical analysis as part of their quality assurance and quality control (QA/QC) program. ALS performed sample homogenization and took two aliquots of sample from the homogenized soil to generate the EPA split sample. EPA split samples were prepared from 15 percent of the collected samples.
- Equipment rinsate blanks were collected to identify possible contamination from the sampling environment or sampling equipment.

Laboratory quality control included the following:

- Matrix spike/matrix spike duplicate (MS/MSD) quality control was conducted for every 20 samples during analysis.
- Laboratory blanks were used to identify possible contamination from the preparation methods (i.e., sieving).


### 2.2.7 Step 7 - Develop the Plan for Obtaining the Data

A resource-effective design for collecting and processing the upland soil samples that would achieve the performance criteria for the study was described in the QAPP (Exponent et al. 2014). TAI and its technical team worked with potentially affected parties to assess the effects of the planned work and seek ways to avoid, minimize, or mitigate any adverse effects on properties with historical significance. A study-specific cultural resources coordination plan (Appendix B of the QAPP) provided relevant background information about Site-related cultural resources, defined measures for protecting resources, and defined procedures for consulting with the appropriate state, federal, and tribal parties with interests in the cultural resources of the Site.

### 2.3 STUDY DESIGN

This section summarizes the study design for the collection of soil samples and the rationale for the design, as presented in detail in the QAPP (Exponent et al. 2014). The sampling approach was developed based on the primary objective of the study, which was to collect information on analytes in upland soils adjacent to the UCR for use in the BERA and HHRA.

### 2.3.1 Overall Design

Soil samples were collected from predetermined DUs within three area types (i.e., ADAs, RFDAs, and WSDAs) following ICS methods (ITRC 2012). Increment samples (increments) were collected from the top 7.5 cm (3 in.) of soil at 30 increment locations within each DU (or 90 increment locations if the DU was sampled in triplicate). Increments were composited in the field to create one sample representing the entire DU. At the analytical laboratory, composite samples were homogenized and sieved into two soil fractions (i.e., $<2-\mathrm{mm}$ and $<149-\mu \mathrm{m}$ ). Data for analyte concentrations in the two fractions will be used in the BERA and HHRA, respectively and are therefore discussed separately in this report. Data for select conventional soil parameters (e.g., CEC, TOC, and pH will be used to assess the RBA of metals in soil to ecological receptors. IVBA analysis was
conducted on a subset (i.e., slightly more than 20 percent) of the $<149-\mu \mathrm{m}$ fraction samples with lead concentrations > $100 \mathrm{mg} / \mathrm{kg}$. IVBA data for lead provide an estimate of the sitespecific oral RBA for lead in soil. ${ }^{16}$

### 2.3.2 Selection of Sampling Areas

Soil samples were collected from the ADAs, RFDAs, and WSDAs as shown in Maps 2-2 through 2-4. The QAPP (Exponent et al. 2014) includes detailed information regarding how the sampling areas were defined and selected; brief summaries (by area type) are provided in the subsections below.

### 2.3.2.1 Aerial Deposition Areas

ADAs are lands adjacent to the UCR and within the river valley that most likely received aerial deposition from historical smelter stack emissions. Two ADAs were designated for sampling: the ADA high-density area, which comprises a $23-\mathrm{mi}^{2}$ corridor along the section of the UCR immediately downstream of the U.S.-Canada border and the ADA primary area, which comprises approximately $99 \mathrm{mi}^{2}$ (Map 2-2). The ADA high-density area was designated for a more extensive sampling due to the perceived likelihood of higher historical deposition rates in that area. In addition to these two subareas, a $16-\mathrm{mi}^{2}$ reserve area situated east of the ADA primary area was designated for sampling in the event that the target number of samples could not be collected from the high-density and primary areas (Map 2-2).

### 2.3.2.2 Relict Floodplain Deposition Areas

For the purpose of this evaluation, the RFDAs were defined as areas of the Site that were flooded under pre-1973 flow conditions but are not expected to flood under current flow and pool level management controls because changes in upstream flow regulations since 1973 have altered the magnitude of flood events. Thus, in the RFDAs, there is a potential for contamination from historical sediment deposition to exist beyond the present-day flood plain limits. The RFDAs are the areas between the maximum pre-1973 and post-1973 high-pool flood levels on the UCR. Five RFDAs that ranged in size from approximately 81 ac $\left(0.13 \mathrm{mi}^{2}\right)$ to 268 ac $\left(0.42 \mathrm{mi}^{2}\right)$ were designated for sampling (RFA through RFE; Map 2-3).

[^4]
### 2.3.2.3 Windblown Sediment Deposition Areas

WSDAs were determined by analyzing wind conditions (i.e., speed, direction, and frequency), concentrations of analytes in nearshore sediment, and percent fines in sediment along the UCR. Wind data were gathered from meteorological data collection systems operated by the U.S. Geological Survey (USGS) and U.S. Bureau of Reclamation (USBR) and used to generate plots of wind speed patterns. The areas with the maximum winds, highest percent fines, and highest concentrations of critical analytes were selected to represent soil in areas with the greatest probability of having the highest analyte concentrations from the re-deposition of windblown sediment. The combined analyses indicated that Marcus Flats and Seven Bays on the left bank (east side) of the UCR (Map 2-4) represented the reasonable worst-case scenario for the enrichment of soils by analytes in windblown sediment. Marcus Flats had the highest analyte concentrations in bank sediment, and Seven Bays had the highest percentage of particles of a size most likely to be transported by wind. Thus, two beach areas each at Marcus Flats (East and West) and near Seven Bays (Columbia Beach North and South) were designated as WSDAs (Map 2-4).

### 2.3.3 Identification of Target DUs and Sampling Locations

The QAPP (Exponent et al. 2014) included detailed information regarding the selection of DUs within the three sampling areas. The process is briefly summarized by area type in the subsections below. Maps 2-2 through 2-4 show the locations of DUs in each area.

### 2.3.3.1 Aerial Deposition Areas

A total of 142 DUs were targeted for sampling in the ADAs: 39 in the ADA high-density area and 103 in the ADA primary area. In addition, reserve DUs were pre-selected for sampling in the event that the target number of DUs in the ADA high-density and primary areas could not be sampled: 7 in the high-density area, 19 in the primary area, and 16 in the reserve area (Table 2-2). The selection of DUs within the ADAs considered factors such as accessibility for sample collection (e.g., areas with less than 30-degree slope due to safety concerns, areas within 550 m of roads to minimize travel time for field personnel to reach sampling locations). DUs were not located within 50 m of roads and railways or within no-sample buffer zones established for active and abandoned mine sites. ${ }^{17}$ DUs

[^5]were also excluded from areas within relict floodplain depositional areas or areas near the surface of the Columbia River at full pool elevation. Using these criteria, a total of 63.94 square miles $\left(\mathrm{mi}^{2}\right)$ were removed from the ADAs ( $10.62 \mathrm{mi}^{2}$ from the ADA high-density area, $44.48 \mathrm{mi}^{2}$ from the ADA primary area, and $8.84 \mathrm{mi}^{2}$ from the ADA reserve area. Specific details on the area removed from the ADAs due to exclusion features are provided in Table B1-4 of the QAPP (Exponent et al. 2014).

### 2.3.3.2 Relict Floodplain Deposition Areas

A total of 29 DUs were targeted for sampling in the RFDAs, with 3 to 9 DUs selected per RFDA (Table 2-2). Locations of DUs (Map 2-3) were determined based on the direction and magnitude of flood events and vegetation type.

### 2.3.3.3 Windblown Sediment Deposition Areas

Two beach areas each at Marcus Flats and Columbia Beach were designated as the WSDAs. The beach areas were identified as being relatively undisturbed because they had no roads, railways, mines, or other sources of dust-producing activity. DUs at each beach area were located within two elongated polygons situated perpendicular to the primary direction of onshore winds. Seven target DUs were selected for each polygon, for a total of 28 DUs (Table 2-2 and Map 2-4). The WSDAs were not evaluated for human health because WSDA sampling was focused on evaluating risks to ecological receptors (USEPA 2012a). Prior sampling showed that the beaches sampled nearest the WSDAs (i.e., Summer Island and Marcus Island for the Marcus Flats WSDAs and Seven Bays for the Columbia Beach WSDAs) had no lead or arsenic concentrations above human health soil screening levels (USEPA 2012c).

Map B1-2 of the QAPP (Exponent et al. 2014) shows the buffer zones established for the mine locations.

## 3 METHODS

### 3.1 FIELD METHODS

The sampling program for the study was outlined in the QAPP (Exponent et al. 2014), which included the FSP as Appendix A. The FSP detailed the procedures and methods for sample collection and processing, field quality control, sample documentation, packaging, and transport, field documentation, laboratory analyses, and data management and reporting.

Field sampling was conducted by ARCADIS U.S., Inc. (ARCADIS). Upland soil samples were collected from the three area types within the Site (i.e., ADAs, RFDAs, and WSDAs) between September 8 and October 23, 2014. Sampling activities were conducted under the direct oversight of EPA or their authorized representatives. Cultural resource monitors from the Confederated Tribes of the Colville Reservation, the National Park Service, and/or the Spokane Tribe of Indians and archaeologists ${ }^{18}$ from URS Corporation were also present during sampling activities to provide oversight for the protection of cultural artifacts in accordance with the protocols outline in the cultural resources coordination plan (Appendix B of the QAPP [Exponent et al. 2014]). The sampling locations, sample collection methods, and field documentation are documented in the field activity report prepared by ARCADIS, which is included as Appendix A of this report. Field changes and deviations from the QAPP are also detailed in the field activity report (Appendix A) and summarized in Section 3.1.3.

### 3.1.1 Sampling Locations

As specified in the QAPP (Exponent et al. 2014) and discussed in Section 2.3, upland soil composite samples were targeted for collection from 199 DUs ( 142 from ADAs, 29 from RFDAs, and 28 from WSDAs). Because of access constraints (e.g., steep terrain, permission not provided by land owner), not all targeted DUs were sampled. Samples were collected from a total of 171 DUs ( 142 from ADAs, 16 from RFDAs, and 13 from WSDAs) (Maps 22 through 2-4 and Table 3-1). Increments were collected from 30 predetermined locations within each DU (or 90 increment locations if the DU was sampled in triplicate). ${ }^{19}$ The

[^6]increment locations were identified using a hand-held global positioning system (GPS) unit. Increments were generally collected within 2 to 10 m of the predetermined location. In some instances (because of steep slopes, access restrictions, or limited accuracy of the GPS units), sample increments were collected more than 10 m from the predetermined location. Table 3 of Appendix A indicates which sample increments were collected, either more than 2 m or more than 10 m , from the predetermined location. The coordinates for the new locations are provided on the Increment Collection Forms in Appendix D of the field activity report (Appendix A).

### 3.1.2 Methods for Sample Collection

This section summarizes the collection and field processing methods of soil samples, which were carried out in accordance with the methods presented in the QAPP (Exponent et al. 2014) and FSP (Appendix A of the QAPP). Field QC samples included triplicate, field split, EPA split, and equipment rinsate blank samples.

### 3.1.2.1 Incremental Composite Sampling

Within each DU, increments were collected at predetermined locations using ICS methods in accordance with the QAPP (Exponent et al. 2014). Once the locations had been cleared of any surface debris, increments were collected from the top 0 to 7.5 cm ( 0 to 3 in .) of soil using a 5 -cm-diameter AMS core sampler (soil punch). Each increment was placed in a dedicated plastic zippered storage bag and examined for cultural materials by a cultural resource monitor and/or archaeologist. Once the soil increment passed the cultural inspection, sampling continued. Sample buckets were labeled at the time of sampling. Labels included the alphanumeric sample identification (ID), as detailed in Section 2.6 of Appendix A of the QAPP, sampler's initials, and sample date and time.

Upon the completion of sampling at a DU, all 30 increments for that DU were composited in a laboratory-decontaminated 2-gal. plastic bucket to form a single incremental composite sample representative of the entire DU. Field observations and sampling activities were recorded in the field notebook and on a tablet computer. Sample collection equipment was grossly decontaminated (i.e., brushed off) between increments at the same DU and fully decontaminated between DUs in accordance with the procedures detailed in Appendix A of the QAPP (Exponent et al. 2014). Samples were placed on ice and stored in a secured, refrigerated truck located in the field. At least once a week, samples were transported by ARCADIS personnel to the analytical laboratory, ALS, in Kelso, Washington. At ALS, the contents of each bucket were homogenized and processed as described in Section 3.2 of this report. The sample ID was recorded on the appropriate
field sampling form in the tablet and on the chain-of-custody (COC) form. Sample labeling details and completed COC forms are provided in the field activity report (Appendix A). A total of 215 composite samples were collected (using ICS methods) from 171 DUs and submitted to ALS for analysis (see Section 3.2). This number includes triplicate samples collected (using ICS methods) at 22 of the DUs (Section 3.1.2.2). ${ }^{20}$ Table 3-2 identifies the numbers of incremental composite samples collected in each area. Tables 3-3a through 3-3c provide information for each of the composite samples in the three deposition areas, respectively. Details regarding the ICS are provided in the field activity report (Appendix A).

### 3.1.2.2 Triplicate Samples

A select number of DUs at each of the three sampling areas were sampled in triplicate to assess the precision of the sampling process in accordance with the QAPP (Exponent et al. 2014). Triplicate DUs were assigned based on a select percentage of DUs. Triplicate samples were collected from 22 DUs using ICS methods (Table 3-1) for a total of 66 triplicate samples. Table 3-2 provides the number of triplicate samples collected within each sampling area. Additional information regarding the triplicate samples is provided in the field activity report (Appendix A).

### 3.1.2.3 Split Samples

As specified in the QAPP (Exponent et al. 2014), two types of split samples were prepared from designated increment composite samples: field split samples and EPA split samples. Split samples were collected from pre-selected DUs to assess the homogeneity of samples collected in the field. A total of 22 (or at least 10 percent) field split samples were prepared and analyzed by ALS, as described in Section 3.2.

EPA split samples were selected by EPA for chemical analysis as part of their QA/QC program. EPA split samples were prepared by ALS from 32 (at least 15 percent) of the collected incremental composite samples and sent to the EPA for analysis, as described in Section 3.2 of this report.

### 3.1.2.4 Equipment Rinsate Blank Samples

In accordance with Appendix A of the QAPP (Exponent et al. 2014), equipment rinsate blank samples were collected to evaluate equipment decontamination procedures.

[^7]Multiple soil punches were used to collect increments at each DU. Soil punches were decontaminated in a manner similar to that used to decontaminate equipment between DUs. To collect a single equipment rinsate blank representative of all soil punches used at a given DU, deionized water was poured over the decontaminated soil punches into a laboratory-supplied sample jar. A total of 22 equipment rinsate blanks were collected from 22 DUs during the field sampling effort. Table 3-1 provides the number of DUs within each sampling area that had equipment rinsate blanks collected. The analysis of equipment rinsate blanks is described in Section 3.2 of this report.

### 3.1.3 Field Changes and Deviations

Procedures presented in the QAPP (Exponent et al. 2014) were followed to the extent possible during implementation of the study. Modifications to the QAPP were categorized as either "changes" or "deviations." Changes and deviations are summarized in Sections 4.1 and 4.2 of the field activity report (Appendix A).

Changes that were identified prior to the initiation of field work and during implementation of the study were documented on change request forms. Eight change request forms were prepared, submitted, and approved by EPA. The EPA-approved forms are included in the field activity report (Appendix A). The following types of changes were documented in the change request forms:

- Increments in 15 DUs were relocated because they were in areas where access to properties was not granted by the land owner.
- The triplicate selection was adjusted to meet the QAPP requirement to have one triplicate DU per RFDA.
- Increments in DUs were moved because the pre-selected increment locations were in areas that were too steep to sample.
- The boundaries for 13 DUs were adjusted in areas where access to the DUs was not granted by the land owners.
- New DUs were selected to replace two DUs that were in areas that were too steep to sample.
- One DU in the ADAs was not sampled because it was located in an area that was too steep to sample, and no suitable alternative location was identified.

Deviations from the QAPP that were identified during implementation of the study and subsequent corrective actions, if required, were documented on deviation/corrective action report forms included in Appendix A. The following deviations were documented in the deviation/corrective action reports:

- Nine increments were collected just outside of the DU boundary but still on property where TAI had permission to sample, due to limited GPS accuracy, terrain restrictions, and/or potential typographical errors.
- Fifteen of the 30 increments in DU ADA-101 were not collected due to access concerns and steep terrain.
- Forty-three increments were collected more than 2 m from the predetermined increment locations due to physical or access restrictions.
- Actual sampling coordinates were not recorded for 16 increments due to equipment malfunction ( 12 increments) or typographical errors (4 increments).

In addition to the deviations listed above, increments were identified after the completion of sample collection that had been collected more than 2 m from the predetermined increment location due to discrepancies between the proposed coordinates and the coordinates recorded during sample collection. Information is provided in Table 3 of Appendix A.

The majority of the sampling changes and deviations did not affect the sampling procedure and consisted primarily of the relocation of increments due to GPS malfunction or inaccuracy, safety concerns, lack of access across private property, or sampling obstacles such as the presence of cobbles or a lack of soil.

### 3.2 LABORATORY METHODS

Following the procedures specified in the QAPP (Exponent et al. 2014), soil samples were processed and analyzed by ALS. Upon receipt at ALS, all incremental composite samples were stored at room temperature, and an aliquot was taken from each sample for the analysis of grain size distribution and pH . The remaining sample underwent ICS processing according to the QAPP and was apportioned for sieving into two fractions: < 2-mm for ecological risk assessments (for ADA, RFDA, and WSDA DUs) and < 149- $\mu \mathrm{m}$ for human health risk assessments (for ADA and RFDA DUs). No $<149-\mu \mathrm{m}$ fraction was prepared for samples collected from the WSDAs because WSDA sampling was focused on evaluating risks to ecological receptors as discussed in Section 2.3.3.3. Any laboratory deviations from the QAPP are discussed in Section 3.2.2.

Table 3-4 summarizes the analyses conducted on the two fractions of soil from the three sampling areas. The $<2-\mathrm{mm}$ fraction was analyzed for total solids, CEC, TOC, and TAL metals, plus molybdenum. The $<149-\mu \mathrm{m}$ fraction was analyzed for total solids and TAL metals, plus molybdenum.

Approximately 20 percent of the samples with lead concentrations $>100 \mathrm{mg} / \mathrm{kg}$ were selected for IVBA analysis in consultation with EPA (see Appendix B). A subsample of the $<149-\mu \mathrm{m}$ fraction was apportioned for IVBA analysis and archived until results from the TAL metals and molybdenum analysis were completed. Soil samples for IVBA analysis were originally planned for the analysis of only lead (Exponent et al. 2014). However, at EPA's request, the IVBA analysis was later expanded to include all TAL metals (Appendix B).

Field split samples were prepared by ALS after homogenization and were assigned their own sample IDs. EPA split samples were also prepared by ALS after homogenization using ICS methods for subsampling, and were provided to EPA for separate analysis (Appendix A).

The 22 equipment rinsate blank samples collected in the field, as well as 30 sieve blanks prepared by ALS, were also analyzed for TAL metals and molybdenum.

### 3.2.1 Methods for Chemical Analysis

ALS prepared and analyzed all soil samples in accordance with the protocols and procedures specified in the QAPP (Exponent et al. 2014), as presented in Table 3-5. Soil samples for metals analyses were prepared with acid digestion following EPA methods 3050B and 7471B. Samples were analyzed for metals according to EPA methods 6010C, 6020A, and 7471B (see Table 3-5). Samples for IVBA were prepared according to EPA 9200.2-86 and analyzed according to EPA 6010B. Analytical concentration goals (ACGs) and MRLs are detailed in the QAPP.

### 3.2.2 Laboratory Deviations

Laboratory methods included two changes related to procedures specified in the QAPP (Exponent et al. 2014). The QAPP states that pH would be measured in the $<2-\mathrm{mm}$ fraction of each soil sample and grain size would be measured in both the $<2-\mathrm{mm}$ fraction and the < $149-\mu \mathrm{m}$ fraction; however, prior to sample collection the decision was made to analyze pH and grain size in the bulk soil sample so that the measurement would not be influenced (i.e., altered) by the soil drying and sieving process. In addition, soil samples were stored at room temperature after receipt in the laboratory rather than at $4^{\circ} \mathrm{C}\left( \pm 2{ }^{\circ} \mathrm{C}\right)$ as stated in the QAPP, and then air dried and sieved prior to analysis. The storage temperature is not as critical for chemicals that are known to be stable (i.e., total metals) as for chemicals that can be volatilized (e.g., mercury, volatile and semivolatile organic compounds) or degraded (e.g., organic compounds). EPA's national functional guidelines for inorganic Superfund data review (USEPA 2010b) leaves data qualification resulting
from not adhering to sample storage requirements of $4{ }^{\circ} \mathrm{C}\left( \pm 2{ }^{\circ} \mathrm{C}\right)$ up to the discretion of the data reviewer. No laboratory method deviations were noted in the data validation reports (available on the "Downloads" page in the project database [http://teckucr.exponent.com]).

### 3.3 DATA EVALUATION APPROACH

The QAPP (Exponent et al. 2014) included procedures for the documentation of field, laboratory, and data validation. The data management plan detailed information related to the storage and handling of all project data.

### 3.3.1 Methods

Sampling efforts were documented in field notebooks and forms, COCs, and GPS files. Deviations from the sampling plan were noted in the field and detailed in the field activity report (Appendix A). All documents were scanned and converted to electronic pdf files. Laboratory data were stored at the analytical laboratory and uploaded to the project database. Data were validated by an independent reviewer, Environmental Standards, Inc. (ESI). Data validation reports and the field activity report were submitted to EPA.

Validated data are tabulated and summarized in this report. Section 5 includes summary statistics for soil data (i.e., number of detections, range, and mean) and a field quality control sample assessment. Field split sample relative percent differences (RPDs) and field triplicate relative standard deviations (RSDs) were evaluated based on control limits of 20 and 35 percent, respectively. Results for each of the DUs are provided in figures, maps, and tables in Section 5; field split and triplicate sample replicates are averaged for applicable DUs.

The site-specific bioavailability of metals to ecological receptors in the $<2-\mathrm{mm}$ fraction using select conventional parameters (e.g., pH, CEC, and TOC) has not been determined for this data report. This evaluation will be conducted as part of the BERA. ${ }^{21}$ For the HHRA, the IVBA results for lead in the $<149-\mu \mathrm{m}$ fraction have been used to calculate oral RBA values for lead in soil. ${ }^{22}$ The human health screening level for lead includes a default

[^8]RBA adjustment of 60 percent. To ensure appropriate comparison of upland soil lead concentrations to the lead screening level, soil concentrations are multiplied by the ratio of the site-specific soil lead RBA value to EPA's default RBA. In addition, arsenic concentrations in the $<149-\mu \mathrm{m}$ fraction have been adjusted for EPA's default RBA of 60 percent arsenic in soil (USEPA 2012b).

### 3.3.2 Deviations from Planned Data Evaluation Approach

There were no changes to the data evaluation approach addressed in the field activity report (Appendix A) or in the data validation reports (available on the "Downloads" page in the project database [http://teck-ucr.exponent.com]).

## 4 VALIDATION ASSESSMENT

Data validation were performed by ESI of Valley Forge, Pennsylvania in accordance with the QAPP (Exponent et al. 2014) based on EPA guidance from the following documents:

- Guidance for Labeling Externally Validated Laboratory Analytical Data for Superfund Use (EPA 540-R-08-005) (USEPA 2009)
- USEPA Contract Laboratory Program National Functional Guidelines for Inorganic Data Review (EPA 540-R-04-004) (USEPA 2004)

Stage 2B validation was conducted for the majority of the soil data. Approximately 15 percent of the data underwent Stage 4 validation, which was in accordance with the QAPP (Exponent et al. 2014). Data were qualified, as needed, based on an evaluation of laboratory and field QC criteria, including holding times, initial and continuing calibration results, blank concentrations, laboratory duplicate and field split RPDs, field triplicate RSDs, serial dilution percent differences (\%Ds), and the recoveries of laboratory control samples (LCSs), internal standards, and MS/MSDs. ESI data validation reports are available on the "Downloads" page in the project database (http://teckucr.exponent.com). The results of the data validation are summarized in the following subsections.

### 4.1 OVERALL DATA QUALITY

A summary of the qualifiers assigned to the metals and conventional parameter data (i.e., $\mathrm{pH}, \mathrm{TOC}, \mathrm{CEC}$, total solids, and grain size) are presented in Tables 4-1a through 4-1c for the ADAs, Tables 4-2a through 4-2c for the RFDAs, and Tables 4-3a and 4-3b for the WSDAs. ${ }^{23}$ All data are usable with the qualifiers presented. The IVBA data are usable as qualified, with the exception of four IVBA results for molybdenum that were qualified as unusable (see Section 4.5.3). The data qualifiers were applied by ESI and included the following:

- "J"-The concentration was considered estimated due to one or more of the following: exceedance of project-specific holding time; analytical interference; LCS, MS/MSD, or reporting limit (RL) standard recovery not within acceptable

[^9]range; high \%D, RPD, or RSD for field or laboratory quality control; or the concentration is between the MDL and the MRL.

- "R"-The data point was unusable (i.e., rejected).
- "U"-The analyte was not detected at or above the MDL.
- "UJ"-The analyte was not detected, but the detection limit is likely higher than reported due to low bias.
- "U*"-The analyte was considered "not detected" because a similar concentration was detected in an associated blank sample. ESI considered the sample weight, percent solids, and dilution factor when evaluating blank contamination. For results qualified " $\mathrm{U}^{*}$," the MDL was changed to the concentration of the method blank.

Note that the numbers of qualified samples presented in Tables 4-1 through 4-4 (obtained from the project database) do not include laboratory QC samples, whereas the numbers of qualified samples presented in the text (obtained from the data validator) include laboratory QC samples. Therefore, the numbers in the text and the tables are not always consistent.

### 4.2 SAMPLE TRANSPORT AND HOLDING TIMES

There were no sample transport issues or exceedances of transport holding times. The QAPP-specified (Exponent et al. 2014) laboratory holding time of 14 days for pH was exceeded for 265 samples. Affected samples were qualified as estimated ("J" flagged).

The CEC data for many samples were originally qualified as estimated ("J" flagged) due to exceedance of the QAPP-specified (Exponent et al. 2014) holding time of 14 days. The qualifiers for holding time were initially applied based on the length of time between sample collection and analysis. However, because the samples were dried within 14 days of collection and the CEC is fixed upon drying, the "J" qualifiers related to hold time were subsequently deemed unnecessary. ESI issued an addendum to the data validation report that is available on the "Downloads" page in the project database (http://teckucr.exponent.com). The unnecessary qualifiers were removed from the project database and are not included in this report.

### 4.3 METALS

The soil metals data are usable as qualified; there are no rejected data for the metals analyses. Samples with reported results between the detection limit (DL) and reporting limit (RL) were qualified as estimated ("J" flagged). Numbers and percentages of qualified samples are presented in Tables 4-1b and 4-1c for the ADAs, Tables 4-2b and 4-2c for the RFDAs, and Table 4-3b for the WSDAs.

In addition, there were no rejected metals data for the 22 equipment rinsate blanks or the 30 laboratory sieve blanks.

### 4.3.1 Calibration

The nondetected concentrations of selenium in four samples were qualified "UJ" due to a low RL standard recovery.

### 4.3.2 Blanks

Concentrations of sodium in 29 samples were qualified as nondetected ("U*" flagged) due to the presence of the analyte at similar concentrations in an associated laboratory blank. The nondetected concentrations of magnesium in three samples were qualified "UJ" due to significant negative instrument bias in the associated calibration blanks. Concentrations of magnesium and/or calcium in five samples were qualified as estimated ("J" flagged) due to significant negative instrument bias in the associated calibration blanks.

### 4.3.3 Matrix Spikes

Concentrations of antimony, barium, cadmium, calcium, chromium, lead, manganese, potassium, and/or zinc in 486 samples were qualified as estimated ("J" flagged) due to MS/MSD recoveries or RPDs that were not within control limits. MS/MSD recoveries and RPDs are provided in the laboratory reports available on the "Downloads" page in the project database (http://teck-ucr.exponent.com).

### 4.3.4 Laboratory Control Samples

Concentrations of aluminum, antimony, molybdenum, and/or thallium in 122 samples were qualified as estimated ("J" flagged) due to LCS recoveries that were not within control limits.

### 4.3.5 Laboratory Duplicates, Field Split Samples, and Triplicate Samples

Concentrations of aluminum, antimony, arsenic, barium, beryllium, cadmium, calcium, chromium, cobalt, iron, lead, magnesium, manganese, mercury, molybdenum, nickel, potassium, sodium, and/or vanadium in 134 samples were qualified as estimated ("J" flagged) due to laboratory duplicate or field split RPDs, or triplicate RSDs that were not within control limits.

### 4.3.6 Interference Check Samples

Concentrations of magnesium, potassium, and/or sodium in 88 samples were qualified as estimated ("J" flagged) due to inductively coupled plasma (ICP) interference.

### 4.3.7 Serial Dilutions

Concentrations of antimony, beryllium, cadmium, magnesium, molybdenum, sodium, silver and/or thallium in 238 samples were qualified as estimated ("J" flagged) due to high serial dilution percent difference.

### 4.3.8 Internal Standards

All metals internal standard results were within acceptable limits.

### 4.4 CONVENTIONAL PARAMETERS

The soil conventional parameters data (i.e., $\mathrm{pH}, \mathrm{TOC}, \mathrm{CEC}$, total solids, and grain size) are usable as qualified. There are no rejected data for conventional parameters analyses. Samples with reported results between the DL and RL were qualified as estimated ("J" flagged). Numbers and percentages of qualified samples are presented in Tables 4-1a through 4-1c, 4-2a through 4-2c, and 4-3a through 4-3b for the ADAs, RFDAs, and WSDAs, respectively.

### 4.4.1 Laboratory and Field Duplicates and Triplicates

Concentrations of CEC and/or TOC in 40 out of 274 samples (14.6\%) were qualified as estimated ("J" flagged) due to laboratory duplicate or field split RPDs, or triplicate RSDs that were not within control limits.

### 4.5 IVBA

The IVBA data are usable as qualified with the exception of four samples for molybdenum listed below in Section 4.5.3. Numbers of qualified samples are listed in Table 4-4. Samples with reported positive results between the DL and RL were qualified as estimated ("J" flagged).

IVBA data are reported as percent bioaccessible based on the concentrations of analyte detected in the soil sample and in a liquid extract. Bioaccessibility percentages calculated from qualified data are qualified as estimated ("J" flagged). Results qualified as not detected due to blank contamination ("U*" flagged) were not excluded from the bioaccessibility calculations. However, bioaccessibility percentages were not calculated for "U" or "R" flagged data, as indicated by "NC" in Table 4-4. The qualifiers detailed in Sections 4.5.1 through 4.5.8 apply to the IVBA soil concentrations (as opposed to the reported bioaccessibility percentages).

### 4.5.1 Calibration

All calibration standard recoveries for IVBA analyses were within control limits.

### 4.5.2 Blanks

Concentrations of molybdenum or sodium in 19 samples were qualified as nondetected ("U*" flagged) due to the presence of the analyte in an associated laboratory blank.

### 4.5.3 Matrix Spikes

Concentrations of molybdenum in 12 samples were qualified as unusable ("R" flagged for nondetected results) or estimated ("J" flagged for detected results) due to very low matrix spike recoveries. The IVBA analyses for molybdenum were rejected in the following samples:

- ADA-061-150um
- ADA-061-150umDUP (laboratory duplicate)
- ADA-057-150um
- ADA-057-150umDUP (laboratory duplicate).

The nondetected concentrations of molybdenum or selenium in five samples were qualified "UJ" due to low MS recoveries.

Concentrations of antimony, iron, manganese, and/or molybdenum in 33 samples were qualified as estimated ("J" flagged) due to MS/MSD recoveries that were not within control limits.

### 4.5.4 Laboratory Control Samples and Standard Reference Material

All LCS and standard reference material results for IVBA analyses were within acceptable limits.

### 4.5.5 Laboratory Duplicates and Field Split Samples

All laboratory and field split sample results for IVBA analyses were within acceptable limits.

### 4.5.6 Interference Check Samples

All interference check results for IVBA analyses were within acceptable limits.

### 4.5.7 Serial Dilutions

Concentrations of lead in 19 samples were qualified as estimated ("J" flagged) due to a high serial dilution percent difference.

### 4.5.8 Internal Standards

All internal standard results for IVBA analyses were within acceptable limits.

## 5 RESULTS

The following sections include summary statistics for all usable data, an evaluation of method reporting limits for nondetected samples, and a comparison of detected values with screening levels for ecological and human receptors. Summary statistics for each sampling area and analyte include the number of detected values and the minimum, maximum, and mean values. Summary statistics are presented in Tables 5-1a through $5-1 \mathrm{c}, 5-2 \mathrm{a}$ through 5-2c, and 5-3a and 5-3b for ADA, RFDA, and WSDA areas, respectively. Table 5-4 provides summary statistics for IVBA results (for ADA and RFDA samples). Figures 5-1a through 5-1c, 5-2a through 5-2c, and 5-3 show the results for conventional parameters in the bulk soil, < 2-mm fractions, and < 149- $\mu \mathrm{m}$ fractions, respectively. Figures 5-4a through 5-4x and 5-5a through 5-5x show results for metals in the $<2-\mathrm{mm}$ and $<149-\mu \mathrm{m}$ fractions, respectively. IVBA results for the TAL metals and molybdenum in the $<149-\mu \mathrm{m}$ fraction are shown in Figures 5-6a through 5-6x. Rejected IVBA data (i.e., four molybdenum results) are not used in the data summaries; however, all data are included in the project database.

In accordance with the draft data management plan (Exponent 2010), nondetected results are represented in calculations as one-half of the MDL. For field split samples and triplicate samples, the average of the replicate results is used to calculate the minimum, maximum, and mean values. Data for EPA split samples, equipment rinsate blanks, and laboratory QA/QC samples, such as MS/MSDs, are not included in the data summaries.

The QAPP (Exponent et al. 2014) identified the soil screening levels for ecological receptors and humans that would be used to determine ACGs for TAL metals and molybdenum. If no screening level was available, then the laboratory MRL was used as the ACG. Nondetected results for metals were compared with a value that was 10 times the ACG, as summarized in Section 5.4.

As discussed in Section 2.2.4, metals data for ecological receptors (i.e., concentrations in the $<2-\mathrm{mm}$ fraction) were compared with available Eco-SSLs. Metals data for human health (i.e., concentrations in the $<149-\mu \mathrm{m}$ fraction) were compared with human health screening levels. ${ }^{24}$ Screening levels are for comparison purposes only and are only intended to identify chemicals for further evaluation in the risk assessments. They do not represent cleanup or action levels (USEPA 2003, 2002).

[^10]
### 5.1 AERIAL DEPOSITION AREAS

For the ADAs, incremental composite samples were collected from 142 DUs as planned (a sampling completion rate of 100 percent [Appendix A]). Samples were collected from 35 DUs in the high-density area ( 6 of which were reserve locations), 91 DUs in the primary area (11 of which were reserve locations), and 16 DUs in the reserve area (see Table 3-2). Only 15 of the 30 planned increments were collected from one DU (ADA-101) due to access concerns and steep terrain (Deviation Report No. 2 [Appendix A]).

Triplicate samples (48) were collected from 16 of the DUs (6 from the high-density area and 10 from the primary area as shown in Table 3-1) for a total of 174 incremental composite samples collected in the ADA.

### 5.1.1 Metals and Conventional Parameters

Summary statistics for metals and conventional parameter data for the high-density and primary areas of the ADA are presented in Table 5-1a for bulk soil, Table 5-1b for the $<2-\mathrm{mm}$ fraction, and Table 5-1c for the $<149-\mu \mathrm{m}$ fraction.

### 5.1.2 IVBA

The IVBA analysis was conducted on samples (<149- $\mu \mathrm{m}$ fraction) from 11 DUs in both the high-density area and the primary area. The IVBA analysis also was conducted on one set of triplicate samples from the primary area (ADA-016). Molybdenum results for two samples in the primary area were not usable ("R" qualified). Summary statistics for the IVBA data are presented in Table 5-4. The data are reported as percent bioaccessible. For lead, EPA has a validated method for relating lead IVBA results to lead oral RBA in soil (USEPA 2007a). Lead RBA is an exposure input for assessing lead risks to humans. Table 5-5 summarizes lead RBA values extrapolated from each IVBA result. Based on the lead RBA results for ADA samples with IVBA results, the average lead RBA for the ADA is 71 percent. Table 5-6 provides the RBA-adjusted lead concentrations using the average RBA for the ADA for DUs without IVBA results.

### 5.2 RELICT FLOODPLAIN DEPOSITION AREAS

For the RFDAs, 24 incremental composite samples were collected from 16 DUs, including triplicate samples collected from four DUs (one DU from each RFDA was sampled in triplicate) (see Table 3-2). The sampling completion rate for the RFDA was 55 percent (Appendix A).

### 5.2.1 Metals and Conventional Parameters

Summary statistics for metals and conventional parameters in the four RFDAs are presented in Table 5-2a for bulk soil, Table 5-2b for the < 2-mm fraction, and Table 5-2c for the $<149-\mu \mathrm{m}$ soil fraction.

### 5.2.2 IVBA

The IVBA analysis was conducted on samples (<149- $\mu \mathrm{m}$ fraction) from three DUs in the RFDA. One set of triplicate samples from a DU in RFA (RFA-001) underwent IVBA analysis. Summary statistics for the IVBA data are presented in Table 5-4. The data are reported as percent bioaccessible. Table 5-5 summarizes lead RBA values extrapolated from each IVBA result. Table 5-6 provides RBA-adjusted lead concentrations. Empirical lead concentrations were adjusted using the ratio of site-specific RBA to EPA's default RBA (see Table 5-5). The ratio of the DU-specific RBA to EPA's default RBA was used when available. For ADA and RFDA DUs that did not have IVBA measured directly (i.e., those not listed in Table 5-5), the average RBA ratio for the ADA overall (including primary and high density) or the RFDA reported in Table 5-5 was applied.

### 5.3 WINDBLOWN SEDIMENT DEPOSITION AREAS

For the WSDAs, 17 incremental composite samples were collected from 13 DUs, including triplicate samples collected from two DUs (one DU each from Columbia Beach North and Marcus Flats East was sampled in triplicate) (see Table 3-2). The sampling completion rate for the WSDA was 46 percent (Appendix A).

### 5.3.1 Metals and Conventional Parameters

Summary statistics for metals and conventional parameters in the WSDAs are presented in Table 5-3a for bulk soil and Table 5-3b for the $<2-\mathrm{mm}$ fraction.

### 5.3.2 IVBA

There were no IVBA samples planned or analyzed for the WSDAs (Appendix B).

### 5.4 FIELD QC SUMMARY

Field split RPDs and triplicate RSDs are summarized in Tables 5-7a through 5-9b. A control limit of 20 percent was used to evaluate field split RPDs and a control limit of

35 percent was used to evaluate triplicate RSDs. The sections below discuss field split RPDs and triplicate RSDs by area ${ }^{25}$.

### 5.4.1 Aerial Deposition Areas

Field QC results for the ADA samples are summarized by splits and triplicates in Table 5-7a for bulk soil samples, Table 5-7b for the < 2-mm fraction, and Table 5-7c for the $<149-\mu \mathrm{m}$ fraction. Total field QC results greater than control limits are as follows (summarized by soil fraction and analyte group):

- Bulk fraction
- Grain size - 38out of 208 data points (18.3percent)
- Other conventional parameters-0 out of 68 data points
- <2-mm-fraction
- Metals - 37 out of 816 data points ( 4.5 percent)
- Conventional parameters - 21 out of 102 data points (20.6 percent)
- < 149- $\mu \mathrm{m}$ fraction
- Metals -4 out of 816 data points ( 0.5 percent)
- Conventional parameters-0 out of 34 data points


### 5.4.2 Relict Floodplain Deposition Areas

Field quality control results for the RFDA samples are summarized by splits and triplicates in Table 5-8a for bulk soil samples, Table 5-8b for the < 2-mm fraction, and Table $5-8 \mathrm{c}$ for the $<149-\mu \mathrm{m}$ fraction. Total field quality control results greater than control limits are as follows (summarized by sieve fraction and analyte group):

- Bulk fraction
- Grain size - 24 out of 52 data points (46.2 percent)
- Other conventional parameters-0 out of 12 data points
- $<2-\mathrm{mm}$ fraction
- Metals -7 out of 145 data points ( 4.8 percent)
- Conventional parameters-0 out of 17 data points

[^11]- $<149-\mu \mathrm{m}$ fraction
- Metals-6 out of 144 data points (4.2 percent)
- Conventional parameters -0 out of 6 data points


### 5.4.3 Windblown Sediment Deposition Areas

Field quality control results for the WSDA samples are summarized by splits and triplicates in Table 5-9a for the bulk soil samples and Table 5-9b for the < 2-mm fraction. Total field quality control results greater than control limits are as follows (summarized by sieve fraction and analyte group):

- Bulk fraction
- Grain size - 6 out of 26 data points (23.1 percent)
- Other conventional parameters-0 out of 8 data points
- <2-mm fraction
- Metals -5 out of 96 data points ( 5.2 percent)
- Conventional parameters -1 out of 12 data points ( 8.3 percent)


### 5.5 EVALUATION OF REPORTING LIMITS FOR NONDETECTED SAMPLES

Target MDLs, MRLs, and ACGs for metals were included in the QAPP (Exponent et al. 2014) (target MDLs and MRLs are presented in Table 2-1). Table 5-10 shows the minimum and maximum MRLs for nondetected metals results (applicable to only a portion of the sodium data in the ADAs and one selenium data point in the WSDAs). The MRLs for all nondetected data points are less than 10 times the ACG.

For conventional parameters, target MDLs and MRLs for total solids and TOC were detailed in the QAPP (Exponent et al. 2014). However, there are no nondetected results for conventional parameters. The QAPP did not include MDLs, MRLs, or ACGs for IVBA.

### 5.6 COMPARISON WITH SCREENING LEVELS

Data were compared with conservative screening levels protective of ecological receptors and human health. This section summarizes the comparisons with the ecological and human health screening levels by area. This comparison is for screening purposes only and is only intended to identify chemicals that should be evaluated in the risk assessments. They do not represent cleanup or action levels (USEPA 2003, 2002).

### 5.6.1 Ecological Screening Levels

Results from the $<2-\mathrm{mm}$ fraction were compared with Eco-SSLs using the values presented in the QAPP (Exponent et al. 2014), which were the lowest of the screening levels adopted by EPA for plants, soil invertebrates, birds, and mammals (USEPA 2010a). ${ }^{26}$ Eco-SSLs were available for 15 of the 24 metals analyzed in the study. ${ }^{27}$ Figures $5-4 a$ through $5-4 x$ provide a comparison of the metals data for DUs from each area with ecological screening levels, when available. Table 5-11a summarizes the number of DUs from the ADAs, RFDAs, and WSDAs that are greater than the available Eco-SSLs. Maps 51 through 5-18 provide a spatial representation of DUs with metals concentrations in the < 2-mm fraction that are greater than the Eco-SSLs. Maps are not provided when either no DUs or all DUs had concentrations greater than the screening level. Comparisons with ecological screening levels are discussed by area in the following subsections.

### 5.6.1.1 Aerial Deposition Areas

Comparisons of available Eco-SSLs with metals data from the ADAs are presented in Table 5-11b. Of the 142 DUs sampled from the ADAs, none had concentrations greater than the Eco-SSL for beryllium ( $21 \mathrm{mg} / \mathrm{kg}$ ) or silver ( $4.2 \mathrm{mg} / \mathrm{kg}$ ). All DUs had concentrations greater than the Eco-SSL for antimony ( $0.27 \mathrm{mg} / \mathrm{kg}$ ), cadmium $(0.36 \mathrm{mg} / \mathrm{kg})$, lead ( $11 \mathrm{mg} / \mathrm{kg}$ ), vanadium ( $7.8 \mathrm{mg} / \mathrm{kg}$ ), and zinc ( $46 \mathrm{mg} / \mathrm{kg}$ ). Some of the DUs had concentrations greater than the Eco-SSLs for the remaining eight metals as follows: 41 for arsenic ( $18 \mathrm{mg} / \mathrm{kg}$ ), 59 for barium ( $330 \mathrm{mg} / \mathrm{kg}$ ), 26 for chromium ( $26 \mathrm{mg} / \mathrm{kg}$ ), 5 for cobalt ( $13 \mathrm{mg} / \mathrm{kg}$ ), 21 for copper ( $28 \mathrm{mg} / \mathrm{kg}$ ), 141 for manganese ( $220 \mathrm{mg} / \mathrm{kg}$ ), 10 for nickel ( $38 \mathrm{mg} / \mathrm{kg}$ ), and 19 for selenium ( $0.52 \mathrm{mg} / \mathrm{kg}$ ).

### 5.6.1.2 Relict Floodplain Deposition Areas

Comparisons of available Eco-SSLs with metals data from the RFDAs are presented in Table 5-11c. Of the 16 DUs sampled from the RFDAs, none had concentrations greater than the Eco-SSL for beryllium ( $21 \mathrm{mg} / \mathrm{kg}$ ), nickel ( $38 \mathrm{mg} / \mathrm{kg}$ ), or silver ( $4.2 \mathrm{mg} / \mathrm{kg}$ ). All DUs had concentrations greater than the Eco-SSL for antimony ( $0.27 \mathrm{mg} / \mathrm{kg}$ ), cadmium $(0.36 \mathrm{mg} / \mathrm{kg})$, lead ( $11 \mathrm{mg} / \mathrm{kg}$ ), manganese ( $220 \mathrm{mg} / \mathrm{kg}$ ), vanadium ( $7.8 \mathrm{mg} / \mathrm{kg}$ ), and zinc $(46 \mathrm{mg} / \mathrm{kg})$. Thirteen DUs had concentrations greater than the Eco-SSL for copper ( 28 $\mathrm{mg} / \mathrm{kg}$ ), and nine DUs had concentrations greater than the Eco-SSL for chromium (26

[^12]$\mathrm{mg} / \mathrm{kg}$ ) and selenium ( $0.52 \mathrm{mg} / \mathrm{kg}$ ). Five DUs had concentrations greater than the EcoSSLs for the three remaining metals (i.e., arsenic [ $18 \mathrm{mg} / \mathrm{kg}$ ], barium [ $330 \mathrm{mg} / \mathrm{kg}$ ], and cobalt [ $13 \mathrm{mg} / \mathrm{kg}$ ]).

### 5.6.1.3 Windblown Sediment Deposition Areas

Comparisons of available Eco-SSLs with metals data from the WSDAs are presented in Table 5-11d. Of the 13 DUs sampled from the WSDAs, none had concentrations greater than the Eco-SSL for 9 of the 14 metals (i.e., arsenic [ $18 \mathrm{mg} / \mathrm{kg}$ ], barium [ $330 \mathrm{mg} / \mathrm{kg}$ ], beryllium [21 mg/kg], chromium [26 mg/kg], cobalt [13 mg/kg], copper [ $28 \mathrm{mg} / \mathrm{kg}$ ], nickel [ $38 \mathrm{mg} / \mathrm{kg}$ ], selenium [ $0.52 \mathrm{mg} / \mathrm{kg}$ ], and silver [ $4.2 \mathrm{mg} / \mathrm{kg}$ ]). All DUs had concentrations greater than the Eco-SSL for manganese ( $220 \mathrm{mg} / \mathrm{kg}$ ) and vanadium $(7.8 \mathrm{mg} / \mathrm{kg})$. For the remaining four metals, 11 DUs had concentrations greater than the Eco-SSL for zinc $(46 \mathrm{mg} / \mathrm{kg}), 8$ DUs had concentrations that were greater than the Eco-SSL for antimony $(0.27 \mathrm{mg} / \mathrm{kg})$ and lead ( $11 \mathrm{mg} / \mathrm{kg}$ ), and 7 DUs had concentrations greater than the Eco-SSL for cadmium $(0.36 \mathrm{mg} / \mathrm{kg})$.

### 5.6.2 Human Health Screening Levels

Results from the < 149- $\mu$ m fraction were compared with human health screening levels presented in the QAPP (Exponent et al. 2014). Human health screening levels were available for 19 of the 23 TAL metals analyzed. ${ }^{28}$ Four of the TAL metals (calcium, magnesium, potassium, and sodium) are essential nutrients and do not have human health screening levels. The QAPP did not include a screening level for molybdenum (which is not a TAL metal); however, for the purposes of evaluation in this report, a screening value of $390 \mathrm{mg} / \mathrm{kg}$ was used based on EPA's regional screening level table updated as of January 2015. Figures 5-5a through 5-5x present the metals data compared with human health screening levels, for those metals with human health RBCs. Maps 5-19 through 5-22 provide a spatial representation of concentrations in the $<149-\mu \mathrm{m}$ fraction for metals that are greater than the human health RBCs (i.e., arsenic and lead). Table 5-12a summarizes the number of DUs from the ADAs and RFDAs that are greater than the available screening levels. For lead, data were adjusted for soil lead RBA prior to comparison. ${ }^{29}$ In addition, arsenic concentrations were adjusted for 60 percent soil arsenic

[^13]oral RBA. Comparisons to human health screening levels are discussed by area in the following subsections.

### 5.6.2.1 Aerial Deposition Areas

Comparisons of human health soil screening levels with metals data for the 142 DUs in the ADAs are presented in Table 5-12b. Arsenic and lead were the only metals detected at levels that were greater than human health screening levels. For arsenic, the screening level ( $9.39 \mathrm{mg} / \mathrm{kg}$ ) was exceeded at 68 of 142 DUs after adjusting for 60 percent soil arsenic oral RBA (Map 5-19). For lead, 21 of 142 DUs had concentrations greater than the human health screening level of $400 \mathrm{mg} / \mathrm{kg}$ after adjusting for the ratio of site-specific RBA of lead to EPA's default RBA (Map 5-20).

### 5.6.2.2 Relict Floodplain Deposition Areas

Comparisons of human health soil screening levels with metals data for the RFDAs are presented in Table 5-12c. Of the 16 DUs in the RFDAs, arsenic and lead were the only metals detected at levels that were greater than the human health screening levels. After adjusting for 60 percent soil arsenic RBA, arsenic concentrations exceeded the screening level of $9.39 \mathrm{mg} / \mathrm{kg}$ at 8 of 16 DUs (Map 5-21). For lead, concentrations at 2 of 16 DUs exceeded the screening level of $400 \mathrm{mg} / \mathrm{kg}$ after adjusting for the ratio of site-specific RBA of lead to EPA's default RBA (Map 5-22).
concentrations are multiplied by the ratio of the site-specific soil lead RBA value to EPA's default RBA.

## 6 SUMMARY AND RECOMMENDATIONS

The purpose of this study was to collect additional soil data to evaluate upland areas adjacent to the UCR potentially affected by point sources (e.g., aerial deposition of smelter particulates), historical fluvial deposition of sediment onto relict floodplains, and redeposition of windblown sediment. These data will be used to assess risk to ecological and human receptors from exposure to metals in the upland soil adjacent to the UCR. The study measured analyte concentrations in soil samples collected from three areas (i.e., ADAs, RFDAs, and WSDAs).

The sampling design, as described in the QAPP (Exponent et al. 2014), used a compositing approach whereby increments were collected from the top 7.5 cm ( 3 in .) of soil at 30 increment locations within each DU. A total of 215 composite samples were collected between September 8 and October 23, 2014, at 171 DUs ( 142 from ADAs, 16 from RFDAs, and 13 from WSDAs).

The number of DUs sampled in the ADA included 35 DUs in the high-density area and 107 DUs in the primary area, of which 16 were from the reserve area. Overall, the collected and analyzed samples for the ADA met targets in the QAPP (Exponent et al. 2014). The selected DU locations in the RFDAs were informed largely by the direction and magnitude of different modeled flood events, as discussed in the QAPP (Exponent et al. 2014). The RFDAs were intended to represent five different relict floodplains. Samples from 55 percent of the RFDA DUs could not be collected either because access to the DU was denied or no response was received from the landowner after multiple requests. All five targeted DU samples from RFA were collected and analyzed. A portion of the DUs in each of RFB through RFD could not be collected, and no samples in RFE could be collected (Map 2-3).

The QAPP (Exponent et al. 2014) stated that Marcus Flats and Columbia Beach were predicted to represent a reasonable worst-case scenario for the enrichment of soil by chemicals in windblown sediment and "soil sampling in these areas will provide confirmatory data about whether or not this possibility is realized." For the WDSA, samples were collected and analyzed from Marcus Flats and Columbia Beach (Map 2-4). Samples from 46 percent WSDA DUs could not be collected because access was denied or no response was received from the landowner after multiple requests. All seven targeted Marcus Flats East DU samples were collected and analyzed. Samples could not be obtained from the Marcus Flats West DUs. For both Columbia Beach North and South, some of the more upland DUs (3 of 7 at Columbia Beach North and 2 of 7 at Columbia Beach South) could not be sampled. Previous beach sampling efforts have identified no
human health risks from exposed sediment at the Marcus Flats area and the Seven Bays area (which includes Columbia Beach) (USEPA 2012a).

The total number of field-collected samples (i.e., 215) included 66 triplicate samples collected in 22 of the DUs. The increment locations were identified using a hand-held GPS unit. Increments were generally collected within 2 to 10 m of the predetermined location. Because of access constraints (e.g., steep terrain, permission not provided by landowner), the predetermined increment locations within DUs sometimes required adjustment, and, as just discussed, not all 199 targeted DUs could be sampled. Sampling activities were conducted under the direct oversight of EPA or their authorized representatives.

Soil samples were processed and analyzed in accordance with the QAPP (Exponent et al. 2014). Samples were sieved into two fractions: < 2-mm for ecological risk assessments (for ADA, RFDA, and WSDA DUs) and $<149-\mu \mathrm{m}$ for human health risk assessments (for ADA and RFDA DUs). ${ }^{30}$ Prior to sieving, an aliquot was taken from each sample for the analysis of grain size, total solids, and pH to inform contaminant bioavailability. The $<2-\mathrm{mm}$ fraction was analyzed for total solids, CEC, TOC, and EPA's TAL metals, plus molybdenum. The $<149-\mu \mathrm{m}$ fraction was analyzed for total solids and TAL metals, plus molybdenum. Approximately 20 percent of the samples with lead concentrations $>100 \mathrm{mg} / \mathrm{kg}$ were selected for IVBA analysis to assess the RBA of lead in soil to which people might be exposed. Laboratory methods included one change related to procedures specified in the QAPP; pH was analyzed in the unsieved bulk fraction of each sample (rather than analyzing pH after drying and sieving the samples). No other laboratory method deviations were noted in the data validation reports. All chemical analyses specified in the QAPP were performed. ${ }^{31}$

Quality assurance and validation of soil chemistry data were performed in accordance with the QAPP (Exponent et al. 2014). Qualifiers were assigned to the metals and conventional parameter data, as appropriate. The MRLs for all nondetected data points for metals were less than 10 times the ACG. All conventional parameters were detected in all samples. The IVBA data are usable as qualified, with the exception of four IVBA results for molybdenum that were flagged as rejected.

[^14]DU-specific sampling results representing an estimate of the mean analyte concentration for each sampling area were compared with conservative screening levels that can be used to identify analytes and areas of potential concern for further evaluation in the ecological and human health risk assessments. Results from the $<2-\mathrm{mm}$ fractions were compared with Eco-SSLs, and results from the $<149-\mu \mathrm{m}$ fraction were compared with residential risk-based screening levels for soils. For the ecological screening, at least one DU from the ADA had concentrations greater than the Eco-SSLs for antimony, arsenic, barium, cadmium, chromium, cobalt, copper, lead, manganese, nickel, selenium, vanadium, and zinc. At least one DU in the RFDA had concentrations greater than Eco-SSLs for all these metals except nickel. In the WSDA, at least one DU had concentrations greater than the Eco-SSLs for antimony, cadmium, lead, manganese, vanadium, and zinc. None of the DUs in any of the areas had concentrations for beryllium or silver greater than the screening levels.

For the human health screening in both the ADAs and RFDAs, only arsenic and lead were at concentrations greater than screening levels. Lead and arsenic concentrations were RBA adjusted prior to comparing to screening levels. Lead was adjusted by the ratio of sitespecific RBA to EPA's default RBA and arsenic was adjusted for 60 percent soil arsenic oral RBA. All other metal concentrations in ADA and RFDA DUs were less than the human health screening levels.

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Figures


Figure 5-1a. pH in Bulk Soil Samples by Deposition Area


Figure 5-1b. Percent Solids in Bulk Soil Samples by Deposition Area


Figure 5-1c. Total Fines in Bulk Soil Samples by Deposition Area


Figure 5-2a. Percent Solids in <2-mm Soil Fractions by Deposition Area


Figure 5-2b. Organic carbon in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area


Figure 5-2c. Cation Exchange Capacity in < 2-mm Soil Fractions by Deposition Area


Figure 5-3. Percent Solids in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-4a. Aluminum Concentrations in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area


Figure 5-4b. Antimony Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure 5-4c. Arsenic Concentrations in < $2-\mathrm{mm}$ Soil Fractions by Deposition Area


Figure 5-4d. Barium Concentrations in < $2-\mathrm{mm}$ Soil Fractions by Deposition Area


Figure 5-4e. Beryllium Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure 5-4f. Cadmium Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure 5-4g. Calcium Concentrations in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area


Figure 5-4h. Chromium Concentrations in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area


Figure 5-4i. Cobalt Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure 5-4j. Copper Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure $5-4 \mathrm{k}$. Iron Concentrations in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area


Figure 5-4I. Lead Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure 5-4m. Magnesium Concentrations in <2-mm Soil Fractions by Deposition Area


Figure 5-4n. Manganese Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure 5-40. Mercury Concentrations in <2-mm Soil Fractions by Deposition Area


Figure 5-4p. Molybdenum Concentrations in <2-mm Soil Fractions by Deposition Area


Figure 5-4q. Nickel Concentrations in <2-mm Soil Fractions by Deposition Area


Figure 5-4r. Potassium Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure 5-4s. Selenium Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure 5-4t. Silver Concentrations in < 2-mm Soil Fractions by Deposition Area


Figure 5-4u. Sodium Concentrations in <2-mm Soil Fractions by Deposition Area


Figure $5-4 \mathrm{v}$. Thallium Concentrations in $<2-\mathrm{mm}$ Soil Fractions by Deposition Area


Figure 5-4w. Vanadium Concentrations in <2-mm Soil Fractions by Deposition Area


Figure $5-4 x$. Zinc Concentrations in < $2-\mathrm{mm}$ Soil Fractions by Deposition Area


Figure 5-5a. Aluminum Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5b. Antimony Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5c. Arsenic Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5d. Barium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5e. Beryllium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5f. Cadmium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure $5-5 \mathrm{~g}$. Calcium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure $5-5 \mathrm{~h}$. Chromium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5i. Cobalt Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5j. Copper Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5k. Iron Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5I. Lead Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5m. Magnesium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5n. Manganese Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-50. Mercury Concentrations in <149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5p. Molybdenum Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5q. Nickel Concentrations in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5r. Potassium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5s. Selenium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5t. Silver Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5u. Sodium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure $5-5 \mathrm{v}$. Thallium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure $5-5 \mathrm{w}$. Vanadium Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-5x. Zinc Concentrations in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6a. Percent Bioaccessible Aluminum in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6b. Percent Bioaccessible Antimony in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6c. Percent Bioaccessible Arsenic in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6d. Percent Bioaccessible Barium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6e. Percent Bioaccessible Beryllium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6f. Percent Bioaccessible Cadmium in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6g. Percent Bioaccessible Calcium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6h. Percent Bioaccessible Chromium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6i. Percent Bioaccessible Cobalt in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6j. Percent Bioaccessible Copper in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6k. Percent Bioaccessible Iron in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6I. Percent Bioaccessible Lead in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6m. Percent Bioaccessible Magnesium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6n. Percent Bioaccessible Manganese in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-60. Percent Bioaccessible Mercury in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6p. Percent Bioaccessible Molybdenum in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6q. Percent Bioaccessible Nickel in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6r. Percent Bioaccessible Potassium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6s. Percent Bioaccessible Selenium in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6t. Percent Bioaccessible Silver in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6u. Percent Bioaccessible Sodium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6v. Percent Bioaccessible Thallium in < 149- $\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6w. Percent Bioaccessible Vanadium in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


Figure 5-6x. Percent Bioaccessible Zinc in $<149-\mu \mathrm{m}$ Soil Fractions by Deposition Area


























## Legend

Relict floodplain deposition area decision unit
$>9.39$ and $\leq 20(\mathrm{mg} / \mathrm{kg})^{\mathrm{a}, \mathrm{b}}$
$\leq 9.39(\mathrm{mg} / \mathrm{kg})^{\mathrm{a}}$

- Populated Places

Major Road
Local Road

- River Mile (USGS)
for 60 percent soil arsenic oral relative bioavailability (USEPA 2012).
Maximum concentration is $16.32 \mathrm{mg} / \mathrm{kg}$.
Note: RF - relict floodplain; A through E identify individual areas


Tables

Table 2-1. Target Method Detection Limits and Method Reporting Limits

|  | Analytical | Analytical <br> Laboratory |
| :--- | :---: | :---: |
| Conventional Parameters | MDL | MRL |

## Notes:

Method detection limits/method reporting limits (MDLs/MRLs) are on a dry weight basis.
The MRL is provided on a dry-weight basis and assumes $50 \%$ moisture in the samples. The MRL for project samples will vary with the actual moisture content of the sample.
NA - not applicable, no MDL or MRL available for this method
CEC - cation exchange capacity
TOC - total organic carbon

Table 2-2. Summary of Target Sampling Locations

| Sampling Area | Number of Target DUs |
| :---: | :---: |
| ADA |  |
| High-density | $39^{\text {a }}$ |
| Primary | $103{ }^{\text {b }}$ |
| Reserve | 16 |
| Total | $142^{\text {c }}$ |
| RFDA |  |
| RFA | 5 |
| RFB | 9 |
| RFC | 8 |
| RFD | 3 |
| RFE | 4 |
| Total | 29 |
| WSDA |  |
| Columbia Beach North | 7 |
| Columbia Beach South | 7 |
| Marcus Flats East | 7 |
| Marcus Flats West | 7 |
| Total | 28 |
| All Areas |  |
| Total | 199 |
| Notes: |  |
| ${ }^{\text {a }}$ In addition to the target DUs in the high-density ADA, an additional 7 DUs were designated as reserve DUs to be used in the event that a target DU was unavailable for sampling. |  |
| ${ }^{\mathrm{b}}$ In addition to the target DUs in the primary ADA, an additional 19 DUs were designated as reserve DUs to be used in the event that a target DU was unavailable for sampling. |  |
| ${ }^{\text {c }}$ The 16 DUs in the reserve area are not included in the total number of target DUs. | of target DUs. |
| DU - decision unit |  |
| RFDA - relict flood plain deposition area |  |
| RFA, RFB, RFC, RFD, RFE - relict flood plain deposition areas A, B, C, D, and E |  |
| WSDA - windblown sediment deposition area |  |

Table 3-1. Summary of Sampling Locations

| Sampling Area | Number of DUs Sampled | Number of DUs Sampled for QC by Type |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | EPA Split | Field Split | Triplicate | Field Equipment Rinsate Blanks |
| ADA |  |  |  |  |  |
| High-density | 35 | 10 | 6 | 6 | 6 |
| Primary | 91 | 17 | 10 | 8 | 10 |
| Reserve | 16 | 0 | 2 | 2 | 0 |
| RFDA |  |  |  |  |  |
| RFA | 5 | 2 | 1 | 1 | 1 |
| RFB | 3 | 0 | 0 | 1 | 1 |
| RFC | 6 | 0 | 1 | 1 | 1 |
| RFD | 2 | 0 | 0 | 1 | 1 |
| RFE | 0 | 0 | 0 | 0 | 0 |
| WSDA |  |  |  |  |  |
| Columbia Beach North | 4 | 2 | 1 | 1 | 1 |
| Columbia Beach South | 2 | 0 | 0 | 0 | 0 |
| Marcus Flats East | 7 | 1 | 1 | 1 | 1 |
| Marcut Flats West | 0 | 0 | 0 | 0 | 0 |
| All Areas |  |  |  |  |  |
| Total | 171 | 32 | 22 | 22 | 22 |

## Notes:

ADA - areal deposition area
DU - decision unit
EPA - U.S. Environmental Protection Agency
QC - quality control
RFDA - relict flood plain deposition area
RFA, RFB, RFC, RFD - relict flood plain deposition areas A, B, C, and D
WSDA - windblown sediment deposition area

Table 3-2. Summary of Total Samples Collected in the Field

| Sampling Area | Number of DUs Sampled | Number of Triplicate Samples Per Sampling Area | Total Number of Samples Per Sampling Area |
| :---: | :---: | :---: | :---: |
| ADA |  |  |  |
| High-density | 35 | 18 | 47 |
| Primary | 91 | 24 | 107 |
| Reserve | 16 | 6 | 20 |
| Total | 142 | 48 | 174 |
| RFDA |  |  |  |
| RFA | 5 | 3 | 7 |
| RFB | 3 | 3 | 5 |
| RFC | 6 | 3 | 8 |
| RFD | 2 | 3 | 4 |
| RFE | 0 | 0 | 0 |
| Total | 16 | 12 | 24 |
| WSDA |  |  |  |
| Columbia Beach North | 4 | 3 | 6 |
| Columbia Beach South | 2 | 0 | 2 |
| Marcus Flats East | 7 | 3 | 9 |
| Marcus Flats West | 0 | 0 | 0 |
| Total | 13 | 6 | 17 |
| All Areas |  |  |  |
| Total | 171 | 66 | 215 |
| Notes: |  |  |  |
| ADA - areal deposition area |  |  |  |
| DU - decision unit |  |  |  |
| RFDA - relict flood plain deposition area |  |  |  |
| RFA, RFB, RFC, RFD - relict WSDA - windblown sedimen | depositional areas $A$ area | C, and D |  |

Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 3-3a. Aerial Deposition Area Planned and Sampled Decision Units

| DU | Sampled? | Reserve DU? | Collection Date | Increments Sampled | QC Samples Collected | Comments | X <br> Coordinate ${ }^{\text {a }}$ | Y <br> Coordinate ${ }^{\mathrm{a}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| High-density |  |  |  |  |  |  |  |  |
| ADA-123 | No | No | ns | ns |  | No access | ns | ns |
| ADA-124 | Yes | No | 10/4-10/5/2014 | 90 | Soil triplicate, EPA split |  | 445930.85 | 5418412.10 |
| ADA-125 | Yes | No | 10/23/2014 | 30 | EPA split |  | 446414.68 | 5419058.76 |
| ADA-126 | Yes | No | 9/12/2014 | 30 | EPA split |  | 444790.43 | 5419389.35 |
| ADA-127 | Yes | No | 10/15/2014 | 30 |  |  | 447387.57 | 5419964.38 |
| ADA-128 | Yes | No | 10/3/2014 | 30 |  |  | 443076.44 | 5420042.99 |
| ADA-129 | No | No | ns | ns |  | No access | ns | ns |
| ADA-130 | No | Yes | ns | ns |  | No access | ns | ns |
| ADA-131 | Yes | Yes | 9/19/2014 | 90 | Soil triplicate, EPA split |  | 449312.85 | 5421241.76 |
| ADA-132 | Yes | No | 9/16/2014 | 30 | Field split |  | 444488.36 | 5421679.27 |
| ADA-133 | Yes | No | 9/24/2014 | 30 |  |  | 442921.06 | 5422266.30 |
| ADA-134 | No | No | ns | ns |  | No access | ns | ns |
| ADA-135 | Yes | Yes | 9/18/2014 | 90 | Soil triplicate |  | 445774.68 | 5421960.23 |
| ADA-136 | Yes | No | 9/11/2014 | 30 | EPA split |  | 443716.50 | 5422241.23 |
| ADA-137 | No | No | ns | ns |  | No access | ns | ns |
| ADA-138 | No | No | ns | ns |  | No access | ns | ns |
| ADA-139 | Yes | No | 10/14/2014 | 30 |  |  | 451060.27 | 5422224.62 |
| ADA-140 | Yes | No | 9/16/2014 | 30 |  |  | 449270.55 | 5422882.66 |
| ADA-141 | Yes | No | 9/23/2014 | 30 | Field split |  | 442596.58 | 5422858.29 |
| ADA-142 | Yes | No | 9/25/2014 | 30 |  |  | 444830.00 | 5422890.11 |
| ADA-143 | Yes | No | 9/15/2014 | 30 |  |  | 446410.86 | 5422879.56 |
| ADA-144 | Yes | No | 9/30/2014 | 30 | EPA split |  | 448020.85 | 5422887.43 |
| ADA-145 | Yes | No | 9/24/2014 | 30 |  |  | 444382.50 | 5423458.64 |
| ADA-146 | Yes | No | 10/3/2014 | 30 |  |  | 445773.07 | 5423271.65 |
| ADA-147 | Yes | No | 9/30/2014 | 30 |  |  | 447710.68 | 5423527.46 |
| ADA-148 | Yes | No | 10/6/2014 | 30 | EPA split |  | 448959.94 | 5423535.21 |
| ADA-149 | No | No | ns | ns |  | No access | ns | ns |
| ADA-150 | Yes | No | 10/7/2014 | 30 |  |  | 449934.27 | 5423856.82 |
| ADA-151 | Yes | Yes | 10/5/2014 | 30 |  |  | 450555.29 | 5423841.69 |
| ADA-152 | Yes | Yes | 10/9/2014 | 30 |  |  | 453062.12 | 5423855.41 |
| ADA-153 | Yes | No | 9/21/2014 | 30 |  |  | 445801.86 | 5424216.90 |
| ADA-154 | Yes | Yes | 9/20-9/21/2014 | 90 | Soil triplicate |  | 446452.64 | 5424466.44 |
| ADA-155 | Yes | No | 9/16/2014 | 30 |  |  | 447068.93 | 5424510.59 |
| ADA-156 | Yes | No | 10/8/2014 | 30 | Field split |  | 448330.16 | 5424459.87 |
| ADA-157 | No | No | ns | ns |  | No access | ns | ns |

Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 3-3a. Aerial Deposition Area Planned and Sampled Decision Units

| DU | Sampled? | Reserve DU? | Collection Date | Increments Sampled | QC Samples Collected | Comments | Coordinate ${ }^{\text {a }}$ | Y Coordinate ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| High-density (continued) |  |  |  |  |  |  |  |  |
| ADA-158 | Yes | Yes | 10/7-10/8/2014 | 90 | Soil triplicate, EPA split |  | 450891.77 | 5424829.20 |
| ADA-159 | Yes | No | 10/4-10/5/2014 | 90 | Soil triplicate, EPA split |  | 448001.19 | 5425158.28 |
| ADA-160 | Yes | No | 10/4/2014 | 30 |  |  | 449285.64 | 5425128.87 |
| ADA-161 | Yes | No | 10/3/2014 | 30 | Field split |  | 450672.59 | 5426008.88 |
| ADA-162 | Yes | No | 10/10/2014 | 30 | Field split |  | 452001.67 | 5425721.97 |
| ADA-163 | No | No | ns | ns |  | No access | ns | ns |
| ADA-164 | Yes | No | 9/30/2014 | 30 | Field split |  | 454291.44 | 5425621.71 |
| ADA-165 | Yes | No | 9/17/2014 | 30 |  |  | 451970.55 | 5426465.76 |
| ADA-166 | No | No | ns | ns |  | No access | ns | ns |
| ADA-167 | No | No | ns | ns |  | No access | ns | ns |
| ADA-168 | Yes | No | 10/15/2014 | 30 | EPA split |  | 454736.09 | 5427633.42 |
| Primary |  |  |  |  |  |  |  |  |
| ADA-001 | Yes | No | 9/13/2014 | 30 | EPA split |  | 430454.64 | 5408511.52 |
| ADA-002 | Yes | No | 10/10/2014 | 30 |  |  | 433403.18 | 5408902.29 |
| ADA-003 | No | Yes | ns | ns |  | No access | ns | ns |
| ADA-004 | Yes | No | 10/10/2014 | 30 |  |  | 434841.35 | 5409509.76 |
| ADA-005 | Yes | No | 10/9/2014 | 30 |  |  | 433970.68 | 5410114.95 |
| ADA-006 | Yes | No | 10/11/2014 | 30 | Field split |  | 431738.01 | 5410479.8 |
| ADA-007 | No | No | ns | ns |  | No access | ns | ns |
| ADA-008 | Yes | No | 10/12/2014 | 30 | EPA split |  | 431121.98 | 5411071.92 |
| ADA-009 | No | No | ns | ns |  | No access | ns | ns |
| ADA-010 | Yes | No | 10/2/2014 | 30 |  |  | 431882.23 | 5411419.95 |
| ADA-011 | No | Yes | ns | ns |  | No access | ns | ns |
| ADA-012 | No | No | ns | ns |  | No access | ns | ns |
| ADA-013 | No | Yes | ns | ns |  | No access | ns | ns |
| ADA-014 | No | No | ns | ns |  | No access | ns | ns |
| ADA-015 | Yes | No | 9/14/2014 | 30 | Field split |  | 431758.61 | 5412692.02 |
| ADA-016 | Yes | Yes | 9/24-9/25/2014 | 90 | Soil triplicate, EPA split |  | 433047.91 | 5412676.30 |
| ADA-017 | Yes | No | 10/1/2014 | 30 |  |  | 436223.22 | 5412670.32 |
| ADA-018 | Yes | No | 10/1/2014 | 30 |  |  | 436860.07 | 5412665.23 |
| ADA-019 | Yes | No | 10/12/2014 | 30 |  |  | 432246.75 | 5413794.69 |
| ADA-020 | Yes | Yes | 9/13-9/14/2014 | 90 | Soil triplicate |  | 434933.00 | 5413770.79 |

Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 3-3a. Aerial Deposition Area Planned and Sampled Decision Units

| DU | Sampled? | Reserve DU? | Collection Date | Increments Sampled | QC Samples Collected | Comments | $X$ Coordinate ${ }^{\text {a }}$ | Y <br> Coordinate ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Primary (continued) |  |  |  |  |  |  |  |  |
| ADA-021 | Yes | No | 10/1/2014 | 30 |  |  | 433865.15 | 5414541.12 |
| ADA-022 | No | No | ns | ns |  | No access | ns | ns |
| ADA-023 | Yes | No | 9/13/2014 | 90 | Soil triplicate, EPA split |  | 436536.75 | 5414561.05 |
| ADA-024 | Yes | No | 10/1/2014 | 30 |  |  | 437401.59 | 5413570.48 |
| ADA-025 | Yes | No | 9/18/2014 | 30 |  |  | 441302.78 | 5413328.80 |
| ADA-026 | Yes | No | 9/17/2014 | 30 |  |  | 443474.57 | 5413075.78 |
| ADA-027 | No | No | na | na |  | No access | na | na |
| ADA-028 | Yes | No | 10/3/2014 | 30 | EPA split |  | 440972.65 | 5414580.01 |
| ADA-029 | No | No | na | na |  | No access | na | na |
| ADA-030 | No | No | na | na |  | No access | na | na |
| ADA-031 | No | No | na | na |  | No access | na | na |
| ADA-032 | No | Yes | na | na |  | No access | na | na |
| ADA-033 | Yes | No | 9/24/2014 | 30 |  |  | 449613.36 | 5414585.04 |
| ADA-034 | Yes | No | 10/10/2014 | 30 | EPA split |  | 449958.59 | 5413313.93 |
| ADA-035 | Yes | Yes | 10/4/2014 | 30 |  |  | 443883.84 | 5415485.41 |
| ADA-036 | No | Yes | ns | ns |  | No access | ns | ns |
| ADA-037 | No | No | ns | ns |  | No access | ns | ns |
| ADA-038 | No | No | ns | ns |  | No access | ns | ns |
| ADA-039 | Yes | No | 10/1/2014 | 30 |  |  | 451394.89 | 5415894.16 |
| ADA-040 | No | No | ns | ns |  | No access | ns | ns |
| ADA-041 | No | Yes | ns | ns |  | No access | ns | ns |
| ADA-042 | Yes | No | 10/9/2014 | 30 |  |  | 446424.57 | 5416827.09 |
| ADA-043 | Yes | No | 10/4/2014 | 30 | Field split |  | 448286.60 | 5417171.67 |
| ADA-044 | Yes | No | 9/18/2014 | 30 | Field split |  | 437517.69 | 5417874.07 |
| ADA-045 | Yes | No | 10/9/2014 | 30 |  |  | 443775.55 | 5417445.40 |
| ADA-046 | Yes | No | 10/2/2014 | 30 |  |  | 444800.65 | 5417460.34 |
| ADA-047 | Yes | No | 10/1/2014 | 30 | EPA split |  | 439605.07 | 5417642.93 |
| ADA-048 | Yes | No | 10/23/2014 | 30 |  |  | 448627.38 | 5418004.96 |
| ADA-049 | Yes | No | 9/18/2014 | 30 | EPA split |  | 450330.43 | 5417827.86 |
| ADA-050 | Yes | No | 10/5/2014 | 30 |  |  | 440794.62 | 5418192.02 |
| ADA-051 | Yes | No | 10/23/2014 | 30 |  | 1 of the 30 coordinates not collected due to tablet issues | 448031.66 | 5418403.88 |
| ADA-052 | Yes | No | 10/3/2014 | 30 |  |  | 439725.88 | 5418827.27 |
| ADA-053 | Yes | Yes | 10/8/2014 | 30 |  |  | 438078.38 | 5419121.05 |
| ADA-054 | Yes | No | 10/1/2014 | 30 |  |  | 441048.01 | 5419092.79 |
| ADA-055 | Yes | No | 10/8-10/9/2014 | 90 | Soil triplicate, EPA split |  | 448657.67 | 5419068.48 |

Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 3-3a. Aerial Deposition Area Planned and Sampled Decision Units

| DU | Sampled? | Reserve DU? | Collection Date | Increments Sampled | QC Samples Collected | Comments | $X$ Coordinate ${ }^{\text {a }}$ | Coordinate ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Primary (continued) |  |  |  |  |  |  |  |  |
| ADA-056 | Yes | No | 9/16/2014 | 30 |  |  | 437163.92 | 5419361.96 |
| ADA-057 | Yes | No | 10/8/2014 | 30 |  |  | 439041.12 | 5419194.04 |
| ADA-058 | Yes | No | 9/20/2014 | 30 |  |  | 450578.68 | 5419381.02 |
| ADA-059 | Yes | No | 10/7/2014 | 30 |  |  | 438831.39 | 5420042.20 |
| ADA-060 | Yes | No | 10/6-10/7/2014 | 90 | Soil triplicate, EPA split | 1 of the 30 coordinates not collected due to tablet issues | 440728.23 | 5419720.32 |
| ADA-061 | Yes | No | 9/17/2014 | 30 |  |  | 437829.58 | 5420006.53 |
| ADA-062 | Yes | No | 10/7/2014 | 30 |  |  | 440695.13 | 5420497.95 |
| ADA-063 | Yes | Yes | 9/18/2014 | 30 | Field split |  | 450559.98 | 5420322.58 |
| ADA-064 | Yes | No | 9/16/2014 | 30 |  |  | 437775.04 | 5420643.94 |
| ADA-065 | Yes | No | 10/8/2014 | 30 |  |  | 441470.61 | 5420669.41 |
| ADA-066 | Yes | No | 10/7/2014 | 30 |  |  | 442276.59 | 5420593.92 |
| ADA-067 | Yes | No | 9/18/2014 | 30 |  |  | 449938.4 | 5420651.82 |
| ADA-068 | No | No | ns | ns |  | No access | ns | ns |
| ADA-069 | No | No | ns | ns |  | No access | ns | ns |
| ADA-070 | Yes | Yes | 10/2/2014 | 30 |  |  | 440339.16 | 5421308.01 |
| ADA-071 | Yes | No | 10/8/2014 | 30 | EPA split |  | 441300.44 | 5421267.67 |
| ADA-072 | No | No | ns | ns |  | No access | ns | ns |
| ADA-073 | Yes | No | 10/4/2014 | 30 |  |  | 440978.67 | 5421926.71 |
| ADA-074 | No | No | ns | ns |  | No access | ns | ns |
| ADA-075 | No | Yes | ns | ns |  | No access | ns | ns |
| ADA-076 | Yes | No | 10/14/2014 | 30 |  |  | 452331.28 | 5422468.07 |
| ADA-077 | No | No | ns | ns |  | No access | ns | ns |
| ADA-078 | Yes | Yes | 9/30/2014 | 30 |  |  | 439741.01 | 5422599.58 |
| ADA-079 | Yes | No | 10/14/2014 | 30 | EPA split |  | 454236.08 | 5421793.21 |
| ADA-080 | No | No | ns | ns |  | No access | ns | ns |
| ADA-081 | Yes | No | 10/8/2014 | 30 |  |  | 440149.13 | 5423520.90 |
| ADA-082 | Yes | No | 10/4/2014 | 30 | Field split | 1 of the 30 coordinates not collected due to tablet issues | 441235.76 | 5423892.46 |
| ADA-083 | No | No | ns | ns |  | DU not sampled; unsafe to access steep terrain | ns | ns |
| ADA-084 | Yes | No | 10/10/2014 | 30 | Field split |  | 454849.78 | 5423489.49 |
| ADA-085 | Yes | No | 9/18/2014 | 30 | Field split |  | 461012.08 | 5423701.15 |
| ADA-086 | No | No | ns | ns |  | No access | ns | ns |
| ADA-087 | No | No | ns | ns |  | No access | ns | ns |
| ADA-088 | Yes | No | 10/3/2014 | 30 |  |  | 442908.22 | 5424496.56 |
| ADA-089 | Yes | No | 10/7/2014 | 30 |  |  | 444826.92 | 5424835.38 |
| ADA-090 | Yes | No | 10/7/2014 | 30 | Field split | 1 of the 30 coordinates not collected due to tablet issues | 442011.75 | 5424972.58 |
| ADA-091 | Yes | No | 10/3/2014 | 30 | EPA split |  | 445538.96 | 5425089.27 |

Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 3-3a. Aerial Deposition Area Planned and Sampled Decision Units

| DU | Sampled? | Reserve DU? | Collection Date | Increments Sampled | QC Samples Collected | Comments | Coordinate ${ }^{\text {a }}$ | Coordinate ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Primary (continued) |  |  |  |  |  |  |  |  |
| ADA-092 | Yes | No | 10/6/2014 | 30 |  |  | 455156.03 | 5424892.13 |
| ADA-093 | Yes | Yes | 9/17/2014 | 30 |  |  | 459485.85 | 5425132.23 |
| ADA-094 | Yes | Yes | 10/17/2014 | 30 | EPA split | 1 of the 30 coordinates not collected due to tablet issues | 442935.87 | 5425460.01 |
| ADA-095 | Yes | No | 10/9/2014 | 30 |  |  | 443519.63 | 5425506.74 |
| ADA-096 | Yes | No | 9/27/2014 | 30 |  |  | 444567.93 | 5425426.22 |
| ADA-097 | Yes | No | 9/25/2014 | 30 |  | 2 of the 30 coordinates not collected due to tablet issues | 446756.20 | 5425504.42 |
| ADA-098 | No | No | ns | ns |  | No access | ns | ns |
| ADA-099 | Yes | Yes | 10/11/2014 | 30 |  |  | 445469.14 | 5426068.74 |
| ADA-100 | No | Yes | na | na |  | No access | na | na |
| ADA-101 | Yes | No | 10/12/2014 | 15 |  | 15 of 30 increments not collected due to safety concerns | 457234.81 | 5425998.13 |
| ADA-102 | Yes | No | 10/9/2014 | 30 |  |  | 459162.26 | 5426045.50 |
| ADA-103 | Yes | No | 9/26/2014 | 30 |  |  | 460746.37 | 5426053.11 |
| ADA-104 | Yes | No | 9/20/2014 | 30 |  |  | 463360.88 | 5426078.09 |
| ADA-105 | Yes | No | 10/11/2014 | 30 | Field split |  | 456630.28 | 5426560.54 |
| ADA-106 | Yes | No | 10/16/2014 | 90 | Soil triplicate |  | 462060.78 | 5426417.34 |
| ADA-107 | Yes | No | 10/2/2014 | 90 | Soil triplicate, EPA split |  | 441309.86 | 5426718.58 |
| ADA-108 | Yes | No | 10/10-10/11/2014 | 90 | Soil triplicate |  | 448933.42 | 5426708.53 |
| ADA-109 | Yes | No | 9/30/2014 | 30 | EPA split |  | 450244.43 | 5426703.66 |
| ADA-110 | Yes | No | 9/26/2014 | 30 |  | 1 of the 30 coordinates not collected due to tablet issues | 451220.13 | 5426731.61 |
| ADA-111 | Yes | No | 10/7/2014 | 24 |  |  | 440003.76 | 5427024.60 |
| ADA-112 | Yes | No | 10/17/2014 | 30 |  |  | 442278.24 | 5427342.19 |
| ADA-113 | Yes | No | 9/20/2014 | 30 |  |  | 442900.90 | 5426988.87 |
| ADA-114 | Yes | No | 10/6/2014 | 30 |  |  | 444059.46 | 5427652.87 |
| ADA-115 | Yes | No | 10/18/2014 | 30 |  |  | 444533.83 | 5427152.76 |
| ADA-116 | Yes | No | 10/9/2014 | 30 |  |  | 445443.39 | 5427369.29 |
| ADA-117 | Yes | No | 9/30/2014 | 30 | EPA split |  | 446897.73 | 5427439.69 |
| ADA-118 | Yes | Yes | 9/30/2014 | 30 |  |  | 447704.37 | 5427630.14 |
| ADA-119 | Yes | No | 10/3/2014 | 30 |  |  | 449078.13 | 5427385.04 |
| ADA-120 | No | No | ns | ns |  | No access | ns | ns |
| ADA-121 | Yes | No | 10/15/2014 | 30 |  |  | 457851.02 | 5427244.43 |
| ADA-122 | Yes | No | 10/11/2014 | 30 |  |  | 461532.93 | 5427265.77 |
| ADA-169 | Yes | Yes | 9/20/2014 | 90 | Soil triplicate |  | 463116.51 | 5421657.07 |
| ADA-170 | Yes | Yes | 9/23/2014 | 30 |  |  | 461139.20 | 5422256.53 |
| ADA-171 | Yes | Yes | 9/24/2014 | 30 |  |  | 461111.06 | 5422860.89 |
| ADA-172 | Yes | Yes | 9/30/2014 | 30 |  |  | 466221.97 | 5422885.95 |
| ADA-173 | Yes | Yes | 9/13/2014 | 90 | Soil triplicate |  | 462347.18 | 5423522.85 |

Upper Columbia River
Soil Study Data Summary and Data Gap Report

## Table 3-3a. Aerial Deposition Area Planned and Sampled Decision Units

| DU | Sampled? | Reserve DU? | Collection Date | Increments Sampled | QC Samples Collected | Comments | Coordinate ${ }^{a}$ | Y <br> Coordinate ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Primary (continued) |  |  |  |  |  |  |  |  |
| ADA-174 | Yes | Yes | 9/12/2014 | 30 | Field split |  | 463357.73 | 5423907.06 |
| ADA-175 | Yes | Yes | 9/16/2014 | 30 |  |  | 464902.99 | 5423841.07 |
| ADA-176 | Yes | Yes | 9/13/2014 | 30 |  |  | 465577.16 | 5423872.26 |
| ADA-177 | Yes | Yes | 9/17/2014 | 30 |  |  | 466824.04 | 5424006.60 |
| ADA-178 | Yes | Yes | 9/12/2014 | 30 |  |  | 466252.63 | 5424581.36 |
| ADA-179 | Yes | Yes | 9/19/2014 | 30 |  |  | 463941.11 | 5424772.17 |
| ADA-180 | Yes | Yes | 9/19/2014 | 30 | Field split |  | 463390.86 | 5425403.66 |
| ADA-181 | Yes | Yes | 9/13/2014 | 30 |  |  | 464924.51 | 5426046.15 |
| ADA-182 | Yes | Yes | 10/14/2014 | 30 |  |  | 464328.05 | 5426703.96 |
| ADA-183 | Yes | Yes | 9/12/2014 | 30 |  |  | 468128.82 | 5426440.83 |
| ADA-184 | Yes | Yes | 9/12/2014 | 30 |  |  | 467752.89 | 5427343.69 |

[^15]Upper Columbia River
Soil Study Data Summary and Data Gap Report

| DU | Sampled? | Reserve DU? | Collection Date | Increments Sampled | QC Samples Collected | Comments | X Coordinate ${ }^{\text {a }}$ | Y <br> Coordinate ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RFA |  |  |  |  |  |  |  |  |
| RFA-001 | Yes | No | 9/11/2014 | 90 | Soil triplicate, EPA split |  | 446849.77 | 5421389.01 |
| RFA-002 | Yes | No | 9/9/2014 | 30 | Field split |  | 446822.53 | 5421184.20 |
| RFA-003 | Yes | No | 9/11/2014 | 30 |  |  | 447209.18 | 5421476.67 |
| RFA-004 | Yes | No | 9/10/2014 | 30 | EPA split |  | 446882.55 | 5421050.59 |
| RFA-005 | Yes | No | 9/9/2014 | 30 |  |  | 447290.32 | 5421427.34 |
| RFB |  |  |  |  |  |  |  |  |
| RFB-001 | No | No | ns | ns |  | No access | ns | ns |
| RFB-002 | Yes | No | 9/10/2014 | 30 |  |  | 443651.34 | 5421384.63 |
| RFB-003 | Yes | No | 9/8, 9/25/2014 | 90 | Soil triplicate |  | 443847.43 | 5421706.14 |
| RFB-004 | No | No | ns | ns |  | No access | ns | ns |
| RFB-005 | No | No | ns | ns |  | No access | ns | ns |
| RFB-006 | No | No | ns | ns |  | No access | ns | ns |
| RFB-007 | No | No | ns | ns |  | No access | ns | ns |
| RFB-008 | Yes | No | 9/8/2014 | 30 |  |  | 444317.65 | 5421245.24 |
| RFB-009 | No | No | ns | ns |  | No access | ns | ns |
| RFC |  |  |  |  |  |  |  |  |
| RFC-001 | No | No | ns | ns |  | No access | ns | ns |
| RFC-002 | No | No | ns | ns |  | No access | ns | ns |
| RFC-003 | Yes | No | 9/9/2014 | 30 | Field split |  | 441435.41 | 5417548.98 |
| RFC-004 | Yes | No | 9/9/2014 | 30 |  |  | 441582.87 | 5417581.77 |
| RFC-005 | Yes | No | 9/9, 9/25/2014 | 90 | Soil triplicate |  | 441406.34 | 5417267.81 |
| RFC-006 | Yes | No | 9/11/2014 | 30 |  |  | 441613.67 | 5417207.39 |
| RFC-007 | Yes | No | 9/10/2014 | 30 |  |  | 441350.94 | 5416878.35 |
| RFC-008 | Yes | No | 9/9/2014 | 30 |  |  | 441541.13 | 5416843.18 |
| RFD |  |  |  |  |  |  |  |  |
| RFD-001 | No | No | ns | ns |  | No access | ns | ns |
| RFD-002 | Yes | No | 9/9/2014 | 30 |  |  | 438218.77 | 5414719.77 |
| RFD-003 | Yes | No | 10/21/2014 | 90 | Soil triplicate |  | 437695.27 | 5414337.20 |
| RFE |  |  |  |  |  |  |  |  |
| RFE-001 | No | No | ns | ns |  | No access | ns | ns |
| RFE-002 | No | No | ns | ns |  | No access | ns | ns |
| RFE-003 | No | No | ns | ns |  | No access | ns | ns |
| RFE-004 | No | No | ns | ns |  | No access | ns | ns |

${ }^{\text {a }}$ Coordinates were calculated as the mean of the increment coordinates. Coordinates for decision units sampled in triplicate are from triplicate ' A '. DU - decision unit
ns - not sampled
RFA, RFB, RFC, RFD - relict flood plain depositional areas A, B, C, and D
Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 3-3c. Windblown Sediment Deposition Area Planned and Sampled Decision Units

| DU | Sampled? | Reserve DU? | Collection Date | Increments Sampled | QC Samples Collected | Comments | X Coordinate ${ }^{a}$ | Y Coordinate ${ }^{a}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Columbia Beach North |  |  |  |  |  |  |  |  |
| CBN-001 | Yes | No | 9/23/2014 | 30 | Field split |  | 399620.83 | 5307771.2 |
| CBN-002 | Yes | No | 9/24/2014 | 30 |  |  | 399711.62 | 5307823.61 |
| CBN-003 | Yes | No | 9/24/2014 | 90 | Soil triplicate, EPA split |  | 399803.78 | 5307878.29 |
| CBN-004 | Yes | No | 9/23/2014 | 30 | EPA split |  | 399883.97 | 5307920.73 |
| CBN-005 | No | No | ns | ns |  | No access | ns | ns |
| CBN-006 | No | No | ns | ns |  | No access | ns | ns |
| CBN-007 | No | No | ns | ns |  | No access | ns | ns |
| Columbia Beach South |  |  |  |  |  |  |  |  |
| CBS-001 | Yes | No | 9/23/2014 | 30 |  |  | 399792.35 | 5307475.73 |
| CBS-002 | Yes | No | 9/24/2014 | 30 |  |  | 399876.47 | 5307506.9 |
| CBS-003 | No | No | ns | ns |  | No access | ns | ns |
| CBS-004 | No | No | ns | ns |  | No access | ns | ns |
| CBS-005 | No | No | ns | ns |  | No access | ns | ns |
| CBS-006 | No | No | ns | ns |  | No access | ns | ns |
| CBS-007 | No | No | ns | ns |  | No access | ns | ns |
| Marcus Flats East |  |  |  |  |  |  |  |  |
| MFE-001 | Yes | No | 9/23/2014 | 30 | Field split |  | 423621.35 | 5392249.49 |
| MFE-002 | Yes | No | 9/23/2014 | 30 |  |  | 423595.26 | 5392357.05 |
| MFE-003 | Yes | No | 9/22/2014 | 30 | EPA split |  | 423569.15 | 5392447.14 |
| MFE-004 | Yes | No | 9/22/2014 | 30 |  |  | 423537.54 | 5392544.98 |
| MFE-005 | Yes | No | 9/23/2014 | 30 |  |  | 423508.46 | 5392636.86 |
| MFE-006 | Yes | No | 9/22/2014 | 30 |  |  | 423478.16 | 5392745.63 |
| MFE-007 | Yes | No | 9/22/2014 | 90 | Soil triplicate |  | 423462.73 | 5392844.19 |
| Marcus Flats West |  |  |  |  |  |  |  |  |
| MFW-001 | No | No | ns | ns |  | No access | ns | ns |
| MFW-002 | No | No | ns | ns |  | No access | ns | ns |
| MFW-003 | No | No | ns | ns |  | No access | ns | ns |
| MFW-004 | No | No | ns | ns |  | No access | ns | ns |
| MFW-005 | No | No | ns | ns |  | No access | ns | ns |
| MFW-006 | No | No | ns | ns |  | No access | ns | ns |
| MFW-007 | No | No | ns | ns |  | No access | ns | ns |

Notes:
a
Coordinates were calculated as the mean of the increment coordinates. Coordinates for decision units sampled in triplicate are from triplicate 'A'. CBN - Columbia Beach North
CBS - Columbia Beach South
DU - decision unit
MFE - Marcus Flats East
MFW - Marcus Flats West
ns - not sampled
Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 3-4. Soil Sample Analysis Summary

| Subarea | ICS Composites |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Bulk Soil |  |  | <2 mm Fraction |  |  |  | <149 $\mu \mathrm{m}$ Fraction |  |  |
|  | Grain Size | pH | Solids | TAL Metals + Mo | TOC | Solids | CEC | TAL Metals + Mo | Solids | IVBA Subset: TAL Metals + Mo ${ }^{\text {a }}$ |
| ADA |  |  |  |  |  |  |  |  |  |  |
| High-density | X | X | X | X | X | X | X | X | X | X |
| Primary | X | X | x | x | X | X | X | x | x | x |
| RFDA |  |  |  |  |  |  |  |  |  |  |
| RFA | X | X | X | X | X | X | X | X | X | x |
| RFB | X | X | X | X | X | X | X | X | X | $n s^{\text {b }}$ |
| RFC | x | $x$ | x | X | X | X | x | X | x | $n s^{\text {b }}$ |
| RFD | X | X | X | X | X | X | X | X | X | X |
| WSDA |  |  |  |  |  |  |  |  |  |  |
| Columbia Beach North | X | X | X | X | X | X | x | $n s^{\text {b }}$ | $n s^{\text {b }}$ | $n s^{\text {b }}$ |
| Columbia Beach South | x | X | x | X | X | X | $x$ | $n s^{\text {b }}$ | $n s^{\text {b }}$ | $n s^{\text {b }}$ |
| Marcus Flats East | X | X | X | X | X | X | X | $n s^{\text {b }}$ | $n s^{\text {b }}$ | $n s^{\text {b }}$ |

[^16]Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 3-5. Analytical Methods for Soil Samples

| Analytes | Sample Preparation |  | Quantitative Analysis |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Protocol | Procedure | Protocol | Procedure |
| Conventional Parameters |  |  |  |  |
| Grain size | NA | NA | PSEP | Sieves and pipette |
| pH | NA | NA | EPA 9045D | Electrometric |
| CEC | EPA 9080 | Displacement with ammonium acetate | EPA 9080 | AAS |
| TOC | SOP: GEN-ASTM | NA | ASTM D4129-05 | Coulometric |
| Percent moisture | NA | NA | EPA 160.3 | Gravimetric |
| TAL Metals/Metalloids |  |  |  |  |
| Calcium (Ca), iron (Fe), magnesium (Mg), potassium (K), and sodium ( Na ) | EPA 3050B | Acid digestion | EPA 6010C | ICP-AES |
| Aluminum (Al), antimony (Sb), arsenic (As), barium ( Ba ), beryllium ( Be ), cadmium ( Cd ), chromium (Cr), cobalt (Co), copper (Cu), lead (Pb), manganese (Mn), nickel (Ni), selenium (Se), silver (Ag), thallium (TI), vanadium (V), and zinc ( Zn$)^{\text {a }}$ | EPA 3050B | Acid digestion | EPA 6020A | ICP-MS |
| Mercury (Hg) (total) | EPA 7471B | Acid digestion/oxidation | EPA 7471B | CVAA |
| Molybdenum (Mo) | EPA 3050B | Acid digestion | EPA 6020A | ICP-MS |
| Lead (Pb) bioaccessibility | EPA 9200.2-86 | Glycine extraction | EPA 6010B | ICP-AES |

All methods subject to change upon consultation with the selected analytical laboratory
${ }^{\text {a }}$ Metals may be reported by EPA Method 6010 rather than EPA Method 6020 if the analyte concentrations are sufficiently high AAS - atomic absorption spectrometry
AES - atomic emission spectrometry
ALS - ALS Environmental
ASTM - American Society for Testing and Materials
CEC - cation exchange capacity
CVAA - cold vapor atomic absorption spectrometry
ICP - inductively coupled plasma
MS - mass spectrometry
NA - not applicable
PSEP - Puget Sound Estuary Program
SOP: GEN-ASTM - ALS standard operating procedure
TAL - target analyte list
$\qquad$



| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $6 \varepsilon 1$ | （\％001）681 | （\％0） 0 | $6 \varepsilon 1$ | səlq90つ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $6 \varepsilon 1$ | （\％001）6と | （\％0） 0 | $6 \varepsilon 1$ |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $6 \varepsilon 1$ | （\％001）68L | （\％0） 0 | $6 \varepsilon 1$ | рлелб әлıеоう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）681 | （\％0） 0 | 681 |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）68L | （\％0） 0 | 681 | ｜әлอ入б әu！ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $6 \varepsilon 1$ | （\％001）68L | （\％0） 0 | $6 \varepsilon 1$ |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）68L | （\％0） 0 | 681 | pues əsıeoo Кıə＾ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）681 | （\％0） 0 | 681 | pues əsıeoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）68 | （\％0） 0 | 681 | pues un！pow |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）68L | （\％0） 0 | 681 | pues əu！」 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）68L | （\％0） 0 | 681 | pues əu！！Кıə＾ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）68L | （\％0） 0 | 681 | H！S |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）68L | $(\% 0) 0$ | 681 | KıIO |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | әz！S U！ex |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）681 | （\％0） 0 | 681 | sp！los |
| 0 | 0 | $\angle 6$ | 0 | 0 | 0 | 0 | 62 | 12 | 0 | 0 | S\＆1 | 0 | 0 | 0 | 0 | 011 | 62 | 0 | （\％001） 68 上 | （\％0） 0 | 681 | Hd |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | s．әұәшел | моппиәлиоว－Kıеш！ıd |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon ¢$ | （\％001） EG | （\％0） 0 | $\varepsilon ¢$ | səl｜q0う |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon \varsigma$ | （\％001） E S | （\％0） 0 | \＆S |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon 9$ | （\％001） ES | （\％0） 0 | \＆S | əөлејб әsıeoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆G | （\％001）EG | （\％0） 0 | $\varepsilon \varsigma$ | рлелб แก！pә\％ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon \varsigma$ | （\％001）EG | （\％0） 0 | $\varepsilon \varsigma$ | ｜әлелб әи！ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆¢ | （\％001）ES | （\％0） 0 | \＆¢ |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon 9$ | （\％001） ES | （\％0） 0 | $\varepsilon 9$ | pues əsıeoo 人ıə $\Lambda$ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon \varsigma$ | （\％001）EG | （\％0） 0 | $\varepsilon ¢$ | pues əsıeoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆G | （\％001）\＆G | （\％0） 0 | \＆я | pues un！pen |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon 9$ | （\％001）ES | （\％0） 0 | \＆¢ | pues əu！」 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆S | （\％001） ES | （\％0） 0 | \＆S |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆S | （\％001）\＆S | （\％0） 0 | \＆S | H！S |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆G | （\％001）\＆G | （\％0） 0 | \＆G | KıIO |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | əz！S U！ex |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | £9 | （\％001）\＆G | （\％0） 0 | £9 | sp！los |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 18 | 61 | 0 | 0 | \＆S | 0 | 0 | 0 | 0 | $\varepsilon t$ | 01 | 0 | （\％001）\＆S | （\％0） 0 | \＆S | Hd |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ＊$\cap$ | ก | $\Gamma$ | ก | ＊N | N | $\Gamma$ | H | ＊ | ＊ 0 | ก | $\Gamma$ | ก | ＊N | N | $\Gamma$ | H | ＊ | sbeig onपi！M słnseyło łunoう |  | Sł｜nsəy ¡əə！əy | $\begin{gathered} \hline \hline \text { səldues } \\ \text { fo ıəqumn } \end{gathered}$ |  |
|  <br>  |  |  |  |  |  |  |  |  |  рәңдәээ૪ łо łunoう |  |  |  |  |  |  |  |  |  |  |  |  |  |


0000000000000000000



| Aluminum |
| :--- |
| Antimony |
| Arsenic |
| Barium |
| Beryllium |
| Cadmium |
| Calcium |
| Chromium |
| Cobalt |
| Copper |
| Iron |
| Lead |
| Magnesium |
| Manganese |
| Mercury |
| Molybdenum |
| Nickel |
| Potassium |




 | Solids |
| :--- |
| High-density - |

| High-density - Conventional Parame |
| :--- |
| CEC |
| Organic carbon |




 H The sample was analyzed as soon as possible after collection to minimize holding time

## koreroqe7

 samples. Therefore, the numbers in the text and the tables are not always consistent.CEC - cation exchange capacity


| 9 | 0 | St | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 601 | （\％001）681 | （\％0） 0 | $6 \varepsilon 1$ | un！pos |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon 1$ | 0 | 0 | 0 | 0 | 0 | 0 | 921 | （\％001）6とト | （\％0） 0 | 681 | 1əヘ！！ |
| 0 | 0 | 61 | 0 | 0 | 0 | 61 | 0 | 0 | 0 | 0 | $\angle Z$ | 0 | 0 | 0 | $\angle Z$ | 0 | 0 | でト | （\％001）68L | （\％0） 0 | $6 \varepsilon 1$ | un！uәə |
| 0 | 0 | $\downarrow$ | 0 | 1 | 乙 | 0 | 0 | 0 | 0 | 0 | G | 0 | Z | $\varepsilon$ | 0 | 0 | 0 | $\downarrow$ ¢ | （\％001）68L | （\％0） 0 | 681 | unissetod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6と1 | （\％001）68L | （\％0） 0 | 681 | ｜ə૫ग！ |
| 0 | 0 | カ1 | 0 | 0 | 0 | 0 | 0 | カ1 | 0 | 0 | 61 | 0 | 0 | 0 | 0 | 0 | 61 | 021 | （\％001）6とL | （\％0） 0 | $6 \varepsilon 1$ | unuəpq介iow |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）6とト | （\％0） 0 | 681 | Kınכaə |
| 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 9 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 8 | 1 เレ | （\％001）6とا | （\％0） 0 | 681 | əsəue6uew |
| 0 | 0 | Z1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 91 | 0 | 0 | 0 | 0 | 0 | 乙 | \＆え। | （\％001）68L | （\％0） 0 | $6 \varepsilon 1$ | un！səu6ew |
| 0 | 0 | 9 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 8 | 0 | 0 | 0 | 1 1． | （\％001）68L | （\％0） 0 | 681 | реә7 |
| 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | G | $\downarrow$ เ | （\％001）681 | （\％0） 0 | $6 \varepsilon 1$ | u0， |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $6 \varepsilon 1$ | （\％001）6とL | （\％0） 0 | $6 \varepsilon 1$ | ıəddoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）6とا | （\％0） 0 | 681 | Heqoう |
| 0 | 0 | 9 | 0 | G | 1 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | $L$ | 1 | 0 | 0 | 0 | 1 1． | （\％001）68L | （\％0） 0 | $6 \varepsilon 1$ | un！woxч๐ |
| 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 9\＆1 | （\％001）68L | （\％0） 0 | 681 | un！oleo |
| 0 | 0 | 01 | 0 | 0 | 01 | 0 | 0 | 0 | 0 | 0 | ャレ | 0 | 0 | カ1 | 0 | 0 | 0 | GZ1 | （\％001）68L | （\％0） 0 | 681 | un！upej |
| 0 | 0 | LE | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 88 | （\％001）6とL | （\％0） 0 | $6 \varepsilon 1$ | un！｜｜Kıəg |
| 0 | 0 | 11 | 0 | 0 | 01 | 0 | 0 | 0 | 0 | 0 | St | 0 | 0 | カ1 | 0 | 0 | 0 | ちて1 | （\％001）681 | （\％0） 0 | 681 | un！ueg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）68L | （\％0） 0 | 681 |  |
| 0 | 0 | 001 | 0 | $\dagger$ | 96 | 0 | 0 | 0 | 0 | 0 | 681 | 0 | G | $\downarrow$ ャレ | 0 | 0 | 0 | 0 | （\％001）681 | （\％0） 0 | 681 | Kuom！！${ }^{\text {a }}$ |
| 0 | 0 | $\dagger$ | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 乙 | عє1 | （\％001）681 | （\％0） 0 | 681 | unu！univ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）681 | （\％0） 0 | 681 | sp！os |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 8 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | 0 | 0 | $\downarrow$ | 0 | 0 | 0 | 67 | （\％001）$¢$ ¢ | （\％0） 0 | \＆S | vu！Z |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon 9$ | （\％001）\＆ | （\％0） 0 | \＆G | mn！peue＾ |
| 0 | 0 | $\varepsilon 1$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | 0 | 0 | 0 | 0 | 0 | 0 | $9 \downarrow$ | （\％001）$¢$ ¢ | （\％0） 0 | \＆G | un！｜｜еч। |
| $\dagger$ | 0 | † | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 67 | （\％001）\＆¢ | （\％0） 0 | $\varepsilon \varsigma$ | un！pos |
| 0 | 0 | $\Sigma$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | ZG | （\％001）\＆ऽ | （\％0） 0 | $\varepsilon \varsigma$ | 1ə＾！！ |
| 0 | 0 | 6 | 0 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | G | 0 | 0 | 0 | S | 0 | 0 | $8 t$ | （\％001）\＆ऽ | （\％0） 0 | $\varepsilon \subseteq$ | un！uəps |
| 0 | 0 | 6 | 0 | 9 | $t$ | 0 | 0 | 0 | 0 | 0 | G | 0 | $\varepsilon$ | 2 | 0 | 0 | 0 | $8 \downarrow$ | （\％001）\＆ | （\％0） 0 | \＆G | un！sselod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆G | （\％001）$¢$ ¢ | （\％0） 0 | \＆ 9 | ｜－YJ！ |
| 0 | 0 | t | 0 | 0 | 0 | 0 | 0 | 乙 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 1 | 19 | （\％001）\＆¢ | （\％0） 0 | $\varepsilon \varsigma$ | unuәpq\} |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆G | （\％001）\＆ऽ | （\％0） 0 | $\varepsilon \varsigma$ | KınכגəW |
| 0 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 11 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 9 | $\angle t$ | （\％001）\＆¢ | （\％0） 0 | $\varepsilon 9$ | əsəue6uew |
| 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 9 | 0 | 0 | $\dagger$ | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | $6 \downarrow$ | （\％001）\＆ | （\％0） 0 | $\varepsilon \varsigma$ | un！səu6ew |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon 9$ | （\％001）\＆ | （\％0） 0 | $\varepsilon \varsigma$ | реәך |
| 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | G | $8 \downarrow$ | （\％001）\＆ऽ | （\％0） 0 | $\varepsilon \subseteq$ | u0，｜ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon 9$ | （\％001）\＆¢ | （\％0） 0 | $\varepsilon \varsigma$ | ıəddoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆G | （\％001）\＆ऽ | （\％0） 0 | $\varepsilon \varsigma$ | Heqoう |
| 0 | 0 | 8 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | 0 | 0 | $\dagger$ | 0 | 0 | 0 | 67 | （\％001）\＆ | （\％0） 0 | \＆G | un！worys |
| 0 | 0 | $\dagger$ | 0 | 0 | 0 | 0 | 0 | $\dagger$ | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | Z | 19 | （\％001）\＆ | （\％0） 0 | \＆G | un！oleo |
| 0 | 0 | G1 | 0 | 0 | St | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 8 | 0 | 0 | 0 | St | （\％001）$¢$ ¢ | （\％0） 0 | \＆ 9 | un！upej |
| 0 | 0 | 乙\＆ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\angle 1$ | 0 | 0 | 0 | 0 | 0 | 0 | $9 \varepsilon$ | （\％001）\＆ | （\％0） 0 | $\varepsilon \varsigma$ | un！｜｜Kıəg |
| 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | 0 | 0 | 0 | 0 | 0 | 0 | 67 | （\％001）\＆ऽ | （\％0） 0 | $\varepsilon \varsigma$ | un！ueg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \＆9 | （\％001）\＆ | （\％0） 0 | \＆G | ग！uəS．$V$ |
| 0 | 0 | 001 | 0 | 6 | 16 | 0 | 0 | 0 | 0 | 0 | EG | 0 | G | 87 | 0 | 0 | 0 | 0 | （\％001）\＆G | （\％0） 0 | $\varepsilon \varsigma$ | Kuom！！uv |
| 0 | 0 | ع1 | 0 | 0 | 0 | 0 | 0 | 9 | 0 | 0 | $L$ | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 9t | （\％001）\＆¢ | （\％0） 0 | \＆S | unu！lun｜v |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | £G | （\％001）¢G | （\％0） 0 | \＆S | sp！los |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ＊ | ก | $\Gamma$ | ก | ＊N | N | $\Gamma$ | H | ＊ | ＊ก | ก | $\Gamma$ | ก | ＊N | N | $\Gamma$ | H | ＊ | sbely onч！̣M sןnsəyfo łunoう | $\begin{gathered} \hline \hline \text { sł\|nsəy } \\ \text { pəןdəoo } \end{gathered}$ | si｜nsəy <br>  | səldmesto ıəquinn | әৃK｜еuも |
| $\begin{gathered} \mathrm{s} \nmid 1 \\ 10 \% \end{gathered}$ |  |  |  |  |  |  |  |  |  <br>  |  |  |  |  |  |  |  |  |  |  |  |  |  |





H The sample was analyzed as soon as possible after collection to minimize holding time．
The result is an estimated value that was detected outside the quantitation range．
 The numbers of qualified samples（obtained from the project database）do not include
samples．Therefore，the numbers in the text and the tables are not always consistent．

| 0 | 0 | $\varepsilon$ | 0 | 0 | $\varepsilon$ | 0 | 0 |  | 0 | 0 | $\dagger$ | 0 | 0 | t | 0 | 0 | 0 | ¢\＆1 | （\％001）681 | （\％0） 0 | 681 | गu！Z |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 681 | （\％001）68L | （\％0） 0 | 681 | un！peue＾ |
| 0 | 0 | tr | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 61 | 0 | 0 | 0 | 0 | 0 | 0 | 021 | （\％001）6\＆ | （\％0） 0 | 681 | un！｜⿺𠃊⿻上丨 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | pənu！？ | sp！o\｜｜rezaW／ |  |
| ＊$\cap$ | ก | r | ก | ＊ | N | $\Gamma$ | H |  | ＊$\cap$ | ก | r | ก | ＊ | N | r | H | ＊ |  |  | Stlnsoy 1．oəəy | sə｜dues „๐ дəqunn | ग十र人，euv |
|  <br>  |  |  |  |  |  |  |  |  |  pəədəээ૪ ł0 łunoう |  |  |  |  |  |  |  |  |  |  |  |  |  |


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| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | เәлецб әлıео |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | ןəлебб un！pә\％ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | ｜əлелб์ әu！」 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | †əィе入б әu！！イıəへ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | pues əsıeoo Кıə＾ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | pues əsıeoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | pues un！pew |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | pues əu！ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | H！S |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | Kelo |
| әz！s |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | （\％001） 6 | （\％0） 0 | 6 | sp！los |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | カt | 99 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | t | G | 0 | （\％001） 6 | （\％0） 0 | 6 | $\mathrm{H}^{\text {d }}$ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） S | （\％0） 0 | G | səlqqoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | ןллеб әлıеоう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | （\％001） s | （\％0） 0 | G | เəлехб un！pə\％ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | ｜әлеגб әu！ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | ןəлелб̆ әu！！イıəへ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | （\％001）¢ | （\％0） 0 | G | pues əsıeoo Кıə＾ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | （\％001） s | （\％0） 0 | G | pues əsıeoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） s | （\％0） 0 | G | pues un！pəw |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | pues əu！ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） s | （\％0） 0 | G | pues əu！！Кıə＾ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | H！S |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） S | （\％0） 0 | G | Kelo |
| әz！s uile |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） S | （\％0） 0 | G | sp！los |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 09 | $0 t$ | 0 | 0 | G | 0 | 0 | 0 | 0 | $\varepsilon$ | 2 | 0 | （\％001） S | （\％0） 0 | G | Hd |
| s．əәәшелед ןеио！！ ¢ $^{\text {a }}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | solq90 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | əлелб әsıeoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | рлелб шก！pә\％ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | ｜әлелб әи！ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | †əィе入б әu！！イıəへ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | pues əsıeoo Kıə $\Lambda$ |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | pues əu！！¢ıə＾ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | H！S |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | Kelo |
| əz！s uluen |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | sp！los |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | $\varepsilon 9$ | $8 \varepsilon$ | 0 | 0 | 8 | 0 | 0 | 0 | 0 | G | $\varepsilon$ | 0 | （\％001） 8 | （\％0） 0 | 8 | Hd |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ＊ | $\bigcirc$ | $\Gamma$ | ก | ＊N | N | $\Gamma$ | H | ＊ | ＊$\cap$ | ก | $\Gamma$ | ก | ＊N | N | $\Gamma$ | H | ＊ | sbela onपt！M sinnsəyjo zunoう |  | ş｜nsəy ๒əə！əу | səlduesfo ıəqunn |  |
|  |  |  |  |  |  |  |  |  | sбिएן рәłdəoว૪ ！0 łunoう |  |  | sб̂eן |  |  |  |  |  |  |  |  |  |  |



| 0 | 0 | $0 t$ | 0 | 0 | $0 t$ | 0 | 0 | 0 | 0 | 0 | 乙 | 0 | 0 | 乙 | 0 | 0 | 0 | $\varepsilon$ | （\％001） S | （\％0） 0 | G | un！ssetod |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | g | （\％001）¢ | （\％0） 0 | G | ןəyग！ |
| 0 | 0 | 09 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001）S | （\％0） 0 | G | unuәpqKiow |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） s | （\％0） 0 | G | KınodeN |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） s | （\％0） 0 | G | әsəueбuew |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | un！səuбеw |
| 0 | 0 | 09 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | $z$ | （\％001） s | （\％0） 0 | G | реә7 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） s | （\％0） 0 | G | UOA｜ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | ıəddo |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） s | （\％0） 0 | G | Heqoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）S | （\％0） 0 | G | un！woxч |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | un！olej |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） s | （\％0） 0 | G | un！ueg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） g | （\％0） 0 | G |  |
| 0 | 0 | 001 | 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | G | 0 | 0 | G | 0 | 0 | 0 | 0 | （\％001）¢ | （\％0） 0 | G | Kuom！！？ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | unulumity |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）S | （\％0） 0 | G | sp！los |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | ¢ | иоqıeग ग！иебио |
| 0 | 0 | $0 t$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | （\％001）¢ | （\％0） 0 | G | ОヨО |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | गu！Z |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！peue＾ |
| 0 | 0 | $\varepsilon 9$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | S | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | （\％001） 8 | （\％0） 0 | 8 | un！！｜⿺𠃊 |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | （\％001） 8 | （\％0） 0 | 8 | un！pos |
| 0 | 0 | $8 \varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） 8 | （\％0） 0 | 8 | 1ə＾！！ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！uəəəS |
| 0 | 0 | $8 \varepsilon$ | 0 | 0 | $8 \varepsilon$ | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | G | （\％001） 8 | （\％0） 0 | 8 | un！ssetod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | 1อ＞ग！ N |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | （\％001） 8 | （\％0） 0 | 8 | unuәpqKiow |
| 0 | 0 | $\varepsilon 9$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | （\％001） 8 | （\％0） 0 | 8 | KınכגəN |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | әsəuебuew |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！səu6ew |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | реә7 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | Uod｜ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | ıəddo |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | Heqos |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！woxyo |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！olej |
| 0 | 0 | $8 \varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） 8 | （\％0） 0 | 8 | un！upes |
| 0 | 0 | $\varepsilon 9$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | （\％001） 8 | （\％0） 0 | 8 | un！｜｜K人20g |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | unueg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | ग！uəs．${ }^{\text {a }}$ |
| 0 | 0 | 001 | 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | （\％001） 8 | （\％0） 0 | 8 | Kuom！！？ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | unulumig |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | sp！los |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | иоqıes ग！uebio |
| 0 | 0 | $8 \varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） 8 | （\％0） 0 | 8 | Оヨコ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ＊ | ก | 「 | ก | ＊N | N | $r$ | H | ＊ | ＊$\cap$ | ก | $r$ | ก | $* \mathrm{~N}$ | N | r | H | ＊ | sbej］onपı！M sınseyło łunoう |  | $\begin{gathered} \hline \hline \text { słlnsey } \\ \text { ןəə!əy } \end{gathered}$ | $\begin{aligned} & \hline \hline \text { səldues } \\ & \text { ןo dequmn } \end{aligned}$ | әł¢K｜еu |
| $\begin{array}{r} \mathrm{s} \mid \mathrm{II} \\ \mathrm{f} 0 \% \end{array}$ | $\begin{aligned} & y \\ & \text { y pe } \\ & \text { an } \end{aligned}$ |  |  |  |  |  |  |  |  <br>  |  |  |  |  |  |  |  |  |  |  |  |  |  |


 $\overline{\text { подер!!ел }}$
 J The result is an estimated value that was detected outside the quantitation range.
N The matrix spike sample recovery is not within control limits. See case narrative. H The sample was analyzed as soon as possible after collection to minimize holding time
The result is an estimated value that was detected outside the quantitation range.
 RFA, RFB, RFC, RFD - relict flood plain depositional areas A, B, C, and D
CEC - cation exchange capacity The numbers of qualified samples (obtained from the tables are not always consistent.

## :səłon

| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | (\%001) $\downarrow$ | (\%0) 0 |  | गu! Z |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | (\%001) $\downarrow$ | (\%0) 0 | $\dagger$ | un!prue^ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | un!\||еч⿺ |
| 0 | 0 | gz | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | un!pos |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | (\%001) $\downarrow$ | (\%0) 0 | $\dagger$ | 1əл!! |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | t | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | un!uajes |
| 0 | 0 | G2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | un!ssetod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | \|อYग! |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | (\%00L) $\downarrow$ | (\%0) 0 | $\downarrow$ | unuəpq<10W |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | (\%00L) $\downarrow$ | (\%0) 0 | $\downarrow$ | KınכxaN |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | (\%00L) $\downarrow$ | (\%0) 0 | † | esouebuew |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | † | (\%00L) $\dagger$ | (\%0) 0 |  | un!səuben |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | (\%00L) $\downarrow$ | (\%0) 0 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| * $\cap$ | ก | $\Gamma$ | ก | *N | N | $\Gamma$ | H | * | * $\cap$ | ก | $\Gamma$ | ก | *N | N | $\Gamma$ | H | * | $\begin{gathered} \hline \text { sbely on } \\ \text { y!p sinsoy } \\ \text { fo tunoo } \end{gathered}$ | $\begin{gathered} \hline \hline \text { splnsəy } \\ \text { pəpdəoov } \end{gathered}$ | $\begin{aligned} & \hline \hline \text { sylnsoug } \\ & \text { pooloy } \end{aligned}$ | $\begin{gathered} \hline \hline \text { soldues } \\ \text { !o дequmn } \end{gathered}$ | गुराएеи |
| sџ\|nsəy pətdəכ૭૪ <br>  |  |  |  |  |  |  |  |  |  <br>  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| 0 | 0 | 0t | 0 | 0 | $0 t$ | 0 | 0 | 0 | 0 | 0 | 乙 | 0 | 0 | Z | 0 | 0 | 0 | $\varepsilon$ | （\％001）¢ | （\％0） 0 | G | unisselod |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） S | （\％0） 0 | G | ｜əyग！ |
| 0 | 0 | 09 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | Z | （\％001）¢ | （\％0） 0 | G | unuәpq\ıow |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） S | （\％0） 0 | G | KınכגəN |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | əsəurбuew |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | un！səu¢ీew |
| 0 | 0 | 09 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001）¢ | （\％0） 0 | G | реә7 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | u0ㅣ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | ıəddoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | Heqoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | un！woxчว |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） S | （\％0） 0 | G | un！э｜eว |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 乙 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | （\％001）¢ | （\％0） 0 | G | un！upej |
| 0 | 0 | 02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | （\％001）¢ | （\％0） 0 | G | un！｜｜イıəg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | un！ueg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | ग！uəs．${ }^{\text {a }}$ |
| 0 | 0 | 001 | 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | S | 0 | 0 | G | 0 | 0 | 0 | 0 | （\％001）¢ | （\％0） 0 | 9 | Kuom！！uv |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） s | （\％0） 0 | 9 | unu！umiv |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001）¢ | （\％0） 0 | G | sp！！os |
| sләдәшелед ןеио！риәлиол－8лу |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | गu！Z |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | mn！peue＾ |
| 0 | 0 | $\varepsilon 9$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | S | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | （\％001） 8 | （\％0） 0 | 8 | un！｜｜еч⿺ |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | （\％001） 8 | （\％0） 0 | 8 | un！pos |
| 0 | 0 | $8 \varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） 8 | （\％0） 0 | 8 | 1ə＾！！ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！uəəəs |
| 0 | 0 | $8 \varepsilon$ | 0 | 0 | $8 \varepsilon$ | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | G | （\％001） 8 | （\％0） 0 | 8 | mn！sselod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | рәYग！ |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | （\％001） 8 | （\％0） 0 | 8 | unuәpq＾IOW |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | KınכגəN |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | əsəueб́uew |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！səu6ew |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | реәך |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | u0，${ }^{\text {a }}$ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | ıəddoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | \＃eqoう |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！woxчว |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！ole |
| 0 | 0 | $8 \varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | G | （\％001） 8 | （\％0） 0 | 8 | un！upej |
| 0 | 0 | $\varepsilon 9$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | G | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | （\％001） 8 | （\％0） 0 | 8 | mn！｜｜＜ıә |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | un！ueg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | ग！uəs．${ }^{\text {a }}$ |
| 0 | 0 | 001 | 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | （\％001） 8 | （\％0） 0 | 8 | Kuom！！uv |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | unu！umiv |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | sp！olfer |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | （\％001） 8 | （\％0） 0 | 8 | sp！！os |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | әursed leu | циәлиоつ－ $\boldsymbol{\text {－}}$ |
| ＊ก | ก | $\Gamma$ | ก | ＊N | N | $r$ | H | ＊ | ＊$\cap$ | ก | $\Gamma$ | ก | ＊N | N | r | H | ＊ |  | $\begin{gathered} \hline \hline \text { sł\|nsəy } \\ \text { pəןdəo৩V } \end{gathered}$ | $\begin{gathered} \hline \text { sł\|nsəy } \\ \text { ןכə!əəy } \end{gathered}$ | $\begin{gathered} \hline \hline \text { sədmes } \\ \text { fo } \begin{array}{c} \text { əqumnN } \end{array} \end{gathered}$ | әұ¢K｜еu |
|  <br>  |  |  |  |  |  |  |  |  |  <br>  |  |  |  |  |  |  |  |  |  |  |  |  |  |




 samples. Therefore, the numbers in the text and the tables are not always
RFA, RFB, RFC, RFD - relict flood plain depositional areas A, B, C, and D
107e.109е7

| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | † | (\%001) $\downarrow$ | (\%0) 0 | $\dagger$ | गu!Z |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | un!peue^ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | un!\||eपı |
| 0 | 0 | GZ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | un!pos |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | (\%001) $\downarrow$ | (\%0) 0 | $\dagger$ | גəハ!! |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\checkmark$ | (\%001) $\downarrow$ | (\%0) 0 | $\dagger$ | un!uәəə |
| 0 | 0 | 92 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | (\%001) $\downarrow$ | (\%0) 0 | $\square$ | un!sselod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | (\%001) $\downarrow$ | (\%0) 0 | † | \|əy ${ }^{\text {a }}$ N |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\downarrow$ | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | unuapq^iow |
| 0 | 0 | GL | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | 0 | 0 | 0 | 0 | 0 | 0 | 1 | (\%001) $\downarrow$ | (\%0) 0 | $\downarrow$ | KınJa\% |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | (\%001) $\downarrow$ | (\%0) 0 | $\dagger$ | әsәuебиен |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | (\%00ト) $\downarrow$ | (\%0) 0 | $\dagger$ | un!sou6en |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\dagger$ | (\%001) $\dagger$ | (\%0) 0 | $\dagger$ | реә7 |


Table 4-2c. Relict Floodplain Deposition Area Summary of Qualifiers for < 149- $\mu \mathrm{m}$ Fraction Metals and Conventional Parameter Results


| 0 | OG | 0 G | OG | 0 | 0 | 0 G | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | （\％001）$て$ | （\％0） 0 | 2 | un！uəəə |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 001 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 0 | z | 0 | 己 | 0 | 0 | 0 | 0 | 0 | （\％001）乙 | （\％0） 0 | Z | un！sselod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | Z | （\％001）乙 | （\％0） 0 | 乙 | Гәソग़ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001） 乙 | （\％0） 0 | 2 | unuapqKiow |
| 0 | 0 | 001 | 0 | 0 | 0 | 001 | 0 | 0 | 0 | 0 | z | 0 | 0 | 0 | Z | 0 | 0 | 0 | （\％001）乙 | （\％0） 0 | 2 | KınJa\％ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 乙 | （\％001）乙 | （\％0） 0 | $\tau$ | әsəue6uew |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 001 | 0 | 0 | z | 0 | 0 | 0 | 0 | 0 | 2 | 0 | （\％001） 乙 | （\％0） 0 | 2 | un！səu6ิ\％ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001）乙 | （\％0） 0 | 2 | реә7 |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 001 | 0 | 0 | z | 0 | 0 | 0 | 0 | 0 | 2 | 0 | （\％001） Z | （\％0） 0 | $\tau$ | uod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001） 乙 | （\％0） 0 | 2 | ı2ddo |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $z$ | （\％001）乙 | （\％0） 0 | 乙 | Heqoo |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 乙 | （\％001）乙 | （\％0） 0 | 乙 | un！woduว |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001）乙 | （\％0） 0 | 2 | un！olej |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001）乙 | （\％0） 0 | $\checkmark$ | un！upej |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | z | 0 | 0 | 0 | 0 | 0 | 0 | 0 | （\％001） 乙 | （\％0） 0 | 2 | un！｜｜ 1 ¢ıg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001）乙 | （\％0） 0 | 2 | unueg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001）$て$ | （\％0） 0 | 2 | ग！uəs．1V |
| 0 | 0 | 001 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 0 | z | 0 | 2 | 0 | 0 | 0 | 0 | 0 | （\％001） Z | （\％0） 0 | 2 | Kuom！！？ |
| 0 | 0 | 001 | 0 | 0 | 0 | 0 | 0 | 001 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | （\％001）乙 | （\％0） 0 | 2 | unulumiv |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001） Z | （\％0） 0 | $\tau$ | splos |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001）乙 | （\％0） 0 | 2 | иоqıes ग！ue6ı0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | （\％001）乙 | （\％0） 0 | 2 | ОヨО |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | L | （\％001）L | （\％0） 0 | $L$ | गu！Z |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | un！peue＾ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | แn！｜！еч⿺ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | un！pos |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | 1әл！！ |
| 0 | 0 | 001 | 0 | 0 | 0 | 001 | 0 | 0 | 0 | 0 | $L$ | 0 | 0 | 0 | $L$ | 0 | 0 | 0 | （\％001）L | （\％0） 0 | $L$ | un！uəəจ |
| 0 | 0 | $\angle 9$ | 0 | $\angle G$ | 0 | 0 | 0 | 0 | 0 | 0 | t | 0 | $\dagger$ | 0 | 0 | 0 | 0 | $\varepsilon$ | （\％001）L | （\％0） 0 | $L$ | un！sselod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | ｜อบग！ |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | unuәpq＾10． |
| 0 | 0 | 001 | 0 | 0 | 0 | 001 | 0 | 0 | 0 | 0 | $L$ | 0 | 0 | 0 | $L$ | 0 | 0 | 0 | （\％001）L | （\％0） 0 | $L$ | Kınכגә |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | әsəue6uew |
| 0 | 0 | LG | 0 | 0 | 0 | 0 | 0 | LS | 0 | 0 | † | 0 | 0 | 0 | 0 | 0 | t | $\varepsilon$ | （\％001）L | （\％0） 0 | $L$ | un！səuбеw |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | реәך |
| 0 | 0 | $\angle 9$ | 0 | 0 | 0 | 0 | 0 | $\angle 9$ | 0 | 0 | $t$ | 0 | 0 | 0 | 0 | 0 | $t$ | $\varepsilon$ | （\％001）L | （\％0） 0 | $L$ | uod |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | ıaddo |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | Heqos |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | un！wodч |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | un！olej |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | un！upes |
| 0 | 0 | $\angle 9$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | t | 0 | 0 | 0 | 0 | 0 | 0 | $\varepsilon$ | （\％001）L | （\％0） 0 | $L$ | un！｜｜ 1 ¢ıg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | unureg |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ |  |
| 0 | 0 | 001 | 0 | $\angle 9$ | $\varepsilon \downarrow$ | 0 | 0 | 0 | 0 | 0 | $L$ | 0 | $\dagger$ | $\varepsilon$ | 0 | 0 | 0 | 0 | （\％001）L | （\％0） 0 | $L$ | Kuom！！？ |
| 0 | 0 | $\angle 9$ | 0 | 0 | 0 | 0 | 0 | $\angle 9$ | 0 | 0 | $\dagger$ | 0 | 0 | 0 | 0 | 0 | $t$ | $\varepsilon$ | （\％001）L | （\％0） 0 | $L$ | unulumiv |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | sp！los |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | иоqıes ग！ue6ı0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $L$ | （\％001）L | （\％0） 0 | $L$ | ОヨО |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | s．әуәшелед ןеиопи |  |  |
| ＊ก | ก | $r$ | n | ＊N | N | $\Gamma$ | H | ＊ | ＊ 0 | n | $\Gamma$ | ก | ＊N | N | $\Gamma$ | H | ＊ |  |  | st｜nsəy છəəə๐્વ | $\begin{gathered} \hline \text { səldmes } \\ \text { fo дəqunn } \end{gathered}$ | गुर｜［位 |
| 10 \％ | 餉 | P！｜e＾ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

## :sapon

 | 10 | $0(0 \%$ |
| :--- | :--- |
| 10 | $0(0 \%$ | 10 (100\%) -



Table 4-4. Summary of Qualifiers for IVBA Results

| Analyte | Soil <br> Fraction | Number of Samples | Rejected Results | Accepted Results | Count of Results with No Flags | Count of Accepted Results Validator Flags |  | Validator Flags, \% of Accepted Results |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | J | NC | J | NC |
| ADA - High-density |  |  |  |  |  |  |  |  |  |
| Aluminum, \% | < 149- $\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 10 | 1 | 0 | 9 | 0 |
| Antimony, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 0 | 11 | 0 | 100 | 0 |
| Arsenic, \% | < 149- $\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 11 | 0 | 0 | 0 | 0 |
| Barium, \% | < 149 - $\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 11 | 0 | 0 | 0 | 0 |
| Beryllium, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 7 | 4 | 0 | 36 | 0 |
| Cadmium, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 10 | 1 | 0 | 9 | 0 |
| Calcium, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 10 | 1 | 0 | 9 | 0 |
| Chromium, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 11 | 0 | 0 | 0 | 0 |
| Cobalt, \% | < 149- $\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 11 | 0 | 0 | 0 | 0 |
| Copper, \% | < 149- $\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 11 | 0 | 0 | 0 | 0 |
| Iron, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 0 | 11 | 0 | 100 | 0 |
| Lead, \% | < 149- $\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 1 | 10 | 0 | 91 | 0 |
| Magnesium, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 10 | 1 | 0 | 9 | 0 |
| Manganese, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 0 | 11 | 0 | 100 | 0 |
| Mercury, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 0 | 9 | 2 | 82 | 18 |
| Molybdenum, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 0 | 10 | 1 | 91 | 9 |
| Nickel, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 11 | 0 | 0 | 0 | 0 |
| Potassium, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 9 | 2 | 0 | 18 | 0 |
| Selenium, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 0 | 1 | 10 | 9 | 91 |
| Silver, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 10 | 1 | 0 | 9 | 0 |
| Sodium, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 0 | 11 | 0 | 100 | 0 |
| Thallium, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 11 | 0 | 0 | 0 | 0 |
| Vanadium, \% | < 149 - $\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 11 | 0 | 0 | 0 | 0 |
| Zinc, \% | < $149-\mu \mathrm{m}$ | 11 | 0 (0\%) | 11 (100\%) | 11 | 0 | 0 | 0 | 0 |
| ADA - Primary |  |  |  |  |  |  |  |  |  |
| Aluminum, \% | < 149- $\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 12 | 1 | 0 | 8 | 0 |
| Antimony, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 0 | 13 | 0 | 100 | 0 |
| Arsenic, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Barium, \% | < 149- $\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Beryllium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 9 | 4 | 0 | 31 | 0 |
| Cadmium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Calcium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Chromium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Cobalt, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Copper, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Iron, \% | < 149- $\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 1 | 12 | 0 | 92 | 0 |
| Lead, \% | < 149- $\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 5 | 8 | 0 | 62 | 0 |
| Magnesium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Manganese, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 1 | 12 | 0 | 92 | 0 |
| Mercury, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 0 | 7 | 6 | 54 | 46 |
| Molybdenum, \% | < $149-\mu \mathrm{m}$ | 13 | 2 (15\%) | 11 (85\%) | 0 | 9 | 2 | 69 | 15 |
| Nickel, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Potassium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Selenium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 0 | 1 | 12 | 8 | 92 |
| Silver, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 6 | 7 | 0 | 54 | 0 |
| Sodium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 0 | 13 | 0 | 100 | 0 |
| Thallium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 11 | 2 | 0 | 15 | 0 |
| Vanadium, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 13 | 0 | 0 | 0 | 0 |
| Zinc, \% | < $149-\mu \mathrm{m}$ | 13 | 0 (0\%) | 13 (100\%) | 12 | 1 | 0 | 8 | 0 |
| RFDA-RFA |  |  |  |  |  |  |  |  |  |
| Aluminum, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Antimony, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 0 | 4 | 0 | 100 | 0 |
| Arsenic, \% | < $149-\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Barium, \% | < $149-\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Beryllium, \% | < $149-\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) |  | 3 | 0 | 75 | 0 |
| Cadmium, \% | < $149-\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 3 | 1 | 0 | 25 | 0 |
| Calcium, \% | < $149-\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |

Table 4-4. Summary of Qualifiers for IVBA Results

| Analyte | Soil Fraction | Number of Samples | Rejected Results | Accepted Results | Count of Results with No Flags | Count of Accepted Results Validator Flags |  | Validator Flags, \% of Accepted Results |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | J | NC | J | NC |
| RFDA - RFA (continued) |  |  |  |  |  |  |  |  |  |
| Chromium, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Cobalt, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Copper, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Iron, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 0 | 4 | 0 | 100 | 0 |
| Lead, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Magnesium, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Manganese, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 0 | 4 | 0 | 100 | 0 |
| Mercury, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 0 | 0 | 4 | 0 | 100 |
| Molybdenum, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 0 | 4 | 0 | 100 | 0 |
| Nickel, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Potassium, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 3 | 1 | 0 | 25 | 0 |
| Selenium, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 0 | 1 | 3 | 25 | 75 |
| Silver, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 3 | 1 | 0 | 25 | 0 |
| Sodium, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 0 | 4 | 0 | 100 | 0 |
| Thallium, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 1 | 3 | 0 | 75 | 0 |
| Vanadium, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| Zinc, \% | < 149- $\mu \mathrm{m}$ | 4 | 0 (0\%) | 4 (100\%) | 4 | 0 | 0 | 0 | 0 |
| RFDA - RFD |  |  |  |  |  |  |  |  |  |
| Aluminum, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 1 | 0 | 100 | 0 |
| Antimony, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 1 | 0 | 100 | 0 |
| Arsenic, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Barium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Beryllium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Cadmium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Calcium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Chromium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Cobalt, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Copper, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Iron, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 1 | 0 | 100 | 0 |
| Lead, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Magnesium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Manganese, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 1 | 0 | 100 | 0 |
| Mercury, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 1 | 0 | 100 | 0 |
| Molybdenum, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 1 | 0 | 100 | 0 |
| Nickel, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Potassium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 1 | 0 | 100 | 0 |
| Selenium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 0 | 1 | 0 | 100 |
| Silver, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 1 | 0 | 100 | 0 |
| Sodium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 0 | 1 | 0 | 100 | 0 |
| Thallium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Vanadium, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |
| Zinc, \% | < 149- $\mu \mathrm{m}$ | 1 | 0 (0\%) | 1 (100\%) | 1 | 0 | 0 | 0 | 0 |

## Notes:

No lab qualifiers were applied to the calculated bioavailability percentages
ADA - aerial deposition area
IVBA - in vitro bioaccessibility assay
$J$ - estimated value
NC - IVBA percentage could not be calculated because the concentration was less than the MRL
RFA, RFB, RFC, RFD - relict flood plain depositional areas A, B, C, and D
Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 5-1a. Aerial Deposition Area Summary Statistics for Bulk Soil Sample Conventional Parameter Results

| Analyte | Number of Samples | Number of Detected Values | Minimum Detected Values | Mean Detected Values | Maximum Detected Values | Minimum Nondetected Value ${ }^{\text {a }}$ | Mean Nondetected Value ${ }^{\text {a }}$ | Maximum Nondetected Value ${ }^{\text {a }}$ | Overall Minimum Value ${ }^{\text {a }}$ | Overall Mean Value ${ }^{\text {a }}$ | Overall Maximum Value ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| High-density - Conventional Parameters |  |  |  |  |  |  |  |  |  |  |  |
| pH (SU) | 35 | 35 | 4.8 | 5.76 | 6.44 | -- | -- | -- | 4.8 | 5.76 | 6.44 |
| Solids (\%) | 35 | 35 | 77.2 | 88.5 | 96.8 | -- | -- | -- | 77.2 | 88.5 | 96.8 |
| Grain Size (\%) |  |  |  |  |  |  |  |  |  |  |  |
| Clay | 35 | 35 | 0.775 | 2.21 | 8.74 | -- | -- | -- | 0.775 | 2.21 | 8.74 |
| Silt | 35 | 35 | 11.44 | 29.2 | 59.64 | -- | -- | -- | 11.44 | 29.2 | 59.64 |
| Very fine sand | 35 | 35 | 5.65 | 15.6 | 31.58 | -- | -- | -- | 5.65 | 15.6 | 31.58 |
| Fine sand | 35 | 35 | 6.36 | 21.2 | 44.43 | -- | -- | -- | 6.36 | 21.2 | 44.43 |
| Medium sand | 35 | 35 | 3.59 | 13.4 | 33.01 | -- | -- | -- | 3.59 | 13.4 | 33.01 |
| Coarse sand | 35 | 35 | 0.94 | 5.91 | 11.91 | -- | -- | -- | 0.94 | 5.91 | 11.91 |
| Very coarse sand | 35 | 35 | 0.58 | 3.8 | 8.58 | -- | -- | -- | 0.58 | 3.8 | 8.58 |
| Very fine gravel | 35 | 35 | 0.13 | 2.63 | 8.27 | -- | -- | -- | 0.13 | 2.63 | 8.27 |
| Fine gravel | 35 | 35 | 0 | 2.26 | 9.36 | -- | -- | -- | 0 | 2.26 | 9.36 |
| Medium gravel | 35 | 35 | 0 | 1.3 | 12.3 | -- | -- | -- | 0 | 1.3 | 12.3 |
| Coarse gravel | 35 | 35 | 0 | 0.046 | 1.61 | -- | -- | -- | 0 | 0.046 | 1.61 |
| Very coarse gravel | 35 | 35 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| Cobbles | 35 | 35 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| Primary - Conventional Parameters |  |  |  |  |  |  |  |  |  |  |  |
| $\mathrm{pH}(\mathrm{SU})$ | 107 | 107 | 4.84 | 6.01 | 8 | -- | -- | -- | 4.84 | 6.01 | 8 |
| Solids (\%) | 107 | 107 | 59.3 | 82.5 | 95.8 | -- | -- | -- | 59.3 | 82.5 | 95.8 |
| Grain Size (\%) |  |  |  |  |  |  |  |  |  |  |  |
| Clay | 107 | 107 | 0.39 | 3.71 | 14.61 | -- | -- | -- | 0.39 | 3.71 | 14.61 |
| Silt | 107 | 107 | 13.02 | 37 | 68 | -- | -- | -- | 13.02 | 37 | 68 |
| Very fine sand | 107 | 107 | 4.32 | 9.56 | 21.49 | -- | -- | -- | 4.32 | 9.56 | 21.49 |
| Fine sand | 107 | 107 | 2.77 | 8.92 | 36.5 | -- | -- | -- | 2.77 | 8.92 | 36.5 |
| Medium sand | 107 | 107 | 2.54 | 7.96 | 33.4 | -- | -- | -- | 2.54 | 7.96 | 33.4 |
| Coarse sand | 107 | 107 | 2.66 | 6.93 | 15.1 | -- | -- | -- | 2.66 | 6.93 | 15.1 |
| Very coarse sand | 107 | 107 | 0.707 | 7.08 | 14.93 | -- | -- | -- | 0.707 | 7.08 | 14.93 |
| Very fine gravel | 107 | 107 | 0.16 | 6.11 | 13.25 | -- | -- | -- | 0.16 | 6.11 | 13.25 |
| Fine gravel | 107 | 107 | 0 | 6.63 | 18.71 | -- | -- | -- | 0 | 6.63 | 18.71 |
| Medium gravel | 107 | 107 | 0 | 4.39 | 34.89 | -- | -- | -- | 0 | 4.39 | 34.89 |
| Coarse gravel | 107 | 107 | 0 | 0.203 | 10.42 | -- | -- | -- | 0 | 0.203 | 10.42 |
| Very coarse gravel | 107 | 107 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| Cobbles | 107 | 107 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |

Notes:
For d
For decision units (DUs) with field split and triplicate samples, summary
${ }^{\mathrm{a}}$ Calculated with nondetected results at one-half of the detection limit.
SU - standard unit
Upper Columbia River
Soil Study Data Summa

| Analyte | Number of Samples | Number of Detected Values | Minimum <br> Detected <br> Values | Mean <br> Detected <br> Values | Maximum <br> Detected Values | Minimum Nondetected Value ${ }^{\text {a }}$ | Mean Nondetected Value ${ }^{\text {a }}$ | Maximum Nondetected Value ${ }^{\text {a }}$ | Overall Minimum Value ${ }^{\text {a }}$ | Overall Mean Value ${ }^{\text {a }}$ | Overall Maximum Value ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Primary - Metals/Metalloids (mg/kg) (continued) |  |  |  |  |  |  |  |  |  |  |  |
| Lead | 107 | 107 | 44.5 | 183 | 592 | -- | -- | -- | 44.5 | 183 | 592 |
| Magnesium | 107 | 107 | 1700 | 4570 | 9510 | -- | -- | -- | 1700 | 4570 | 9510 |
| Manganese | 107 | 107 | 373 | 1090 | 2350 | -- | -- | -- | 373 | 1090 | 2350 |
| Mercury | 107 | 107 | 0.0167 | 0.0719 | 0.16 | -- | -- | -- | 0.0167 | 0.0719 | 0.16 |
| Molybdenum | 107 | 107 | 0.39 | 1.56 | 7.81 | -- | -- | -- | 0.39 | 1.56 | 7.81 |
| Nickel | 107 | 107 | 8.38 | 24.2 | 59.4 | -- | -- | -- | 8.38 | 24.2 | 59.4 |
| Potassium | 107 | 107 | 874 | 1770 | 4320 | -- | -- | -- | 874 | 1770 | 4320 |
| Selenium | 107 | 107 | 0.133 | 0.387 | 3.32 | -- | -- | -- | 0.133 | 0.387 | 3.32 |
| Silver | 107 | 107 | 0.0533 | 0.304 | 1.18 | -- | -- | -- | 0.0533 | 0.304 | 1.18 |
| Sodium | 107 | 99 | 55 | 118 | 258 | 25.7 | 45 | 62 | 25.7 | 113 | 258 |
| Thallium | 107 | 107 | 0.12 | 0.256 | 0.54 | -- | -- | -- | 0.12 | 0.256 | 0.54 |
| Vanadium | 107 | 107 | 13.7 | 31.9 | 63.2 | -- | -- | -- | 13.7 | 31.9 | 63.2 |
| Zinc | 107 | 107 | 72.4 | 293 | 1150 | -- | -- | -- | 72.4 | 293 | 1150 |

For decision units (DUs) with field split and triplicate samples, summary statistics are based on the average of results for the DU. Nondetected values (NDs) are included as half the reporting limits (RLs). ${ }^{a}$ Calculated with nondetected results at one-half of the detection limit. CEC - cation exchange capacity
$\mathrm{me} / 100 \mathrm{gm}$ - miliequivalents per 100 grams $\mathrm{mg} / \mathrm{kg}$ - milligram per kilogram
-- - no nondetected values
Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 5-1c. Aerial Deposition Area Summary Statistics for $<149-\mu \mathrm{m}$ Fraction Metals and Conventional Parameter Results

| Analyte | Number of Samples | Number of Detected Values | Minimum <br> Detected Values | Mean Detected Values | Maximum <br> Detected Values | Minimum Nondetected Value ${ }^{\text {a }}$ | Mean Nondetected Value ${ }^{\text {a }}$ | Maximum Nondetected Value ${ }^{\text {a }}$ | Overall <br> Minimum Value ${ }^{\text {a }}$ | Overall Mean Value ${ }^{\text {a }}$ | Overall Maximum Value ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| High-density - Conventional Parameters (\%) |  |  |  |  |  |  |  |  |  |  |  |
| Solids | 35 | 35 | 93.4 | 96.9 | 99 | -- | -- | -- | 93.4 | 96.9 | 99 |
| High-density - Metals/Metalloids (mg/kg) |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | 35 | 35 | 7180 | 12800 | 22600 | -- | -- | -- | 7180 | 12800 | 22600 |
| Antimony | 35 | 35 | 1.27 | 4.24 | 12.6 | -- | -- | -- | 1.27 | 4.24 | 12.6 |
| Arsenic | 35 | 35 | 9.03 | 22.2 | 55.9 | -- | -- | -- | 9.03 | 22.2 | 55.9 |
| Barium | 35 | 35 | 75.7 | 206 | 453 | -- | -- | -- | 75.7 | 206 | 453 |
| Beryllium | 35 | 35 | 0.265 | 0.417 | 0.83 | -- | -- | -- | 0.265 | 0.417 | 0.83 |
| Cadmium | 35 | 35 | 1.71 | 6.64 | 16.3 | -- | -- | -- | 1.71 | 6.64 | 16.3 |
| Calcium | 35 | 35 | 1760 | 3640 | 6130 | -- | -- | -- | 1760 | 3640 | 6130 |
| Chromium | 35 | 35 | 11.9 | 17.9 | 30.3 | -- | -- | -- | 11.9 | 17.9 | 30.3 |
| Cobalt | 35 | 35 | 3.69 | 6.03 | 11 | -- | -- | -- | 3.69 | 6.03 | 11 |
| Copper | 35 | 35 | 11.1 | 25.1 | 54.8 | -- | -- | -- | 11.1 | 25.1 | 54.8 |
| Iron | 35 | 35 | 10200 | 15300 | 20500 | -- | -- | -- | 10200 | 15300 | 20500 |
| Lead | 35 | 35 | 120 | 348 | 988 | -- | -- | -- | 120 | 348 | 988 |
| Magnesium | 35 | 35 | 1970 | 3440 | 5210 | -- | -- | -- | 1970 | 3440 | 5210 |
| Manganese | 35 | 35 | 237 | 594 | 942 | -- | -- | -- | 237 | 594 | 942 |
| Mercury | 35 | 35 | 0.0267 | 0.0948 | 0.5 | -- | -- | -- | 0.0267 | 0.0948 | 0.5 |
| Molybdenum | 35 | 35 | 0.39 | 0.692 | 0.99 | -- | -- | -- | 0.39 | 0.692 | 0.99 |
| Nickel | 35 | 35 | 9.37 | 15.1 | 32.8 | -- | -- | -- | 9.37 | 15.1 | 32.8 |
| Potassium | 35 | 35 | 828 | 1550 | 2300 | -- | -- | -- | 828 | 1550 | 2300 |
| Selenium | 35 | 35 | 0.155 | 0.327 | 0.68 | -- | -- | -- | 0.155 | 0.327 | 0.68 |
| Silver | 35 | 35 | 0.1 | 0.335 | 1.01 | -- | -- | -- | 0.1 | 0.335 | 1.01 |
| Sodium | 35 | 33 | 44.9 | 104 | 224 | 53 | 57.5 | 62 | 44.9 | 101 | 224 |
| Thallium | 35 | 35 | 0.2 | 0.37 | 0.77 | -- | -- | -- | 0.2 | 0.37 | 0.77 |
| Vanadium | 35 | 35 | 17.7 | 25.6 | 33.5 | -- | -- | -- | 17.7 | 25.6 | 33.5 |
| Zinc | 35 | 35 | 121 | 355 | 735 | -- | -- | -- | 121 | 355 | 735 |
|  |  |  |  |  |  |  |  |  |  |  |  |
| Solids | 107 | 107 | 86.4 | 95 | 98.1 | -- | -- | -- | 86.4 | 95 | 98.1 |
| Primary - Metals/Metalloids (mg/kg) |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | 107 | 107 | 10900 | 17300 | 25900 | -- | -- | -- | 10900 | 17300 | 25900 |
| Antimony | 107 | 107 | 0.45 | 1.9 | 5.91 | -- | -- | -- | 0.45 | 1.9 | 5.91 |
| Arsenic | 107 | 107 | 5.58 | 15.2 | 32.6 | -- | -- | -- | 5.58 | 15.2 | 32.6 |
| Barium | 107 | 107 | 160 | 360 | 1270 | -- | -- | -- | 160 | 360 | 1270 |
| Beryllium | 107 | 107 | 0.33 | 0.537 | 1.08 | -- | -- | -- | 0.33 | 0.537 | 1.08 |
| Cadmium | 107 | 107 | 0.42 | 4.24 | 11.8 | -- | -- | -- | 0.42 | 4.24 | 11.8 |
| Calcium | 107 | 107 | 1810 | 4790 | 14300 | -- | -- | -- | 1810 | 4790 | 14300 |
| Chromium | 107 | 107 | 8.69 | 21.4 | 65.5 | -- | -- | -- | 8.69 | 21.4 | 65.5 |
| Cobalt | 107 | 107 | 4.4 | 8 | 18 | -- | -- | -- | 4.4 | 8 | 18 |
| Copper | 107 | 107 | 7.28 | 21.7 | 65.3 | -- | -- | -- | 7.28 | 21.7 | 65.3 |
| Iron | 107 | 107 | 12100 | 18600 | 27200 | -- | -- | -- | 12100 | 18600 | 27200 |
| Lead | 107 | 107 | 34.8 | 166 | 456 | -- | -- | -- | 34.8 | 166 | 456 |
| Magnesium | 107 | 107 | 1470 | 4190 | 9880 | -- | -- | -- | 1470 | 4190 | 9880 |
| Manganese | 107 | 107 | 382 | 843 | 1650 | -- | -- | -- | 382 | 843 | 1650 |

Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 5-1c. Aerial Deposition Area Summary Statistics for $<149-\mu \mathrm{m}$ Fraction Metals and Conventional Parameter Results

| Analyte | Number of Samples | Number of Detected Values | Minimum Detected Values | Mean Detected Values | Maximum <br> Detected Values | Minimum Nondetected Value ${ }^{\text {a }}$ | Mean Nondetected Value ${ }^{\text {a }}$ | Maximum Nondetected Value ${ }^{\text {a }}$ | Overall Minimum Value ${ }^{\text {a }}$ | Overall Mean Value ${ }^{\text {a }}$ | Overall Maximum Value ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Primary - Metals/Metalloids (mg/kg) (continued) |  |  |  |  |  |  |  |  |  |  |  |
| Mercury | 107 | 107 | 0.02 | 0.0599 | 0.14 | -- | -- | -- | 0.02 | 0.0599 | 0.14 |
| Molybdenum | 107 | 107 | 0.39 | 1.25 | 5.72 | -- | -- | -- | 0.39 | 1.25 | 5.72 |
| Nickel | 107 | 107 | 11 | 24.2 | 66.4 | -- | -- | -- | 11 | 24.2 | 66.4 |
| Potassium | 107 | 107 | 778 | 1760 | 3860 | -- | -- | -- | 778 | 1760 | 3860 |
| Selenium | 107 | 107 | 0.09 | 0.322 | 2.17 | -- | -- | -- | 0.09 | 0.322 | 2.17 |
| Silver | 107 | 107 | 0.0333 | 0.261 | 0.94 | -- | -- | -- | 0.0333 | 0.261 | 0.94 |
| Sodium | 107 | 101 | 76.9 | 136 | 267 | 48.95 | 52.9 | 63.5 | 48.95 | 132 | 267 |
| Thallium | 107 | 107 | 0.12 | 0.26 | 0.52 | -- | -- | -- | 0.12 | 0.26 | 0.52 |
| Vanadium | 107 | 107 | 17.9 | 31.2 | 56.6 | -- | -- | -- | 17.9 | 31.2 | 56.6 |
| Zinc | 107 | 107 | 64.8 | 288 | 1160 | -- | -- | -- | 64.8 | 288 | 1160 |
| Notes: |  |  |  |  |  |  |  |  |  |  |  |
| For decision units ${ }^{\text {a }}$ Calculated with $\mathrm{mg} / \mathrm{kg}$ - milligram -- - no nondetected | s) with field s <br> detected resu <br> kilogram <br> ues | and triplicate at one-half of | les, summa detection lim | istics are b | n the avera | results for the D | Nondetected va | (NDs) are includ | d as half the reporti | ing limits (RLs). |  |


| Analyte | Number of Samples | Number of Detected Values | Minimum Detected Values | Mean Detected Values | Maximum Detected Values | Minimum Nondetected Value $^{\mathrm{a}}$ | Mean Nondetected Value ${ }^{\text {a }}$ | Maximum Nondetected Value ${ }^{\text {a }}$ | Overall Minimum Value ${ }^{\text {a }}$ | Overall Mean Value ${ }^{\text {a }}$ | Overall Maximum Value ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RFA - Conventional Parameters |  |  |  |  |  |  |  |  |  |  |  |
| pH (SU) | 5 | 5 | 7.25 | 7.5 | 7.8 | -- | -- | -- | 7.25 | 7.5 | 7.8 |
| Solids (\%) | 5 | 5 | 94.6 | 95.4 | 97.6 | -- | -- | -- | 94.6 | 95.4 | 97.6 |
| Grain Size (\%) |  |  |  |  |  |  |  |  |  |  |  |
| Clay | 5 | 5 | 0.45 | 0.694 | 1 | -- | -- | -- | 0.45 | 0.694 | 1 |
| Silt | 5 | 5 | 5.57 | 7.8 | 11.2 | -- | -- | -- | 5.57 | 7.8 | 11.2 |
| Very fine sand | 5 | 5 | 9.9 | 14.1 | 16 | -- | -- | -- | 9.9 | 14.1 | 16 |
| Fine sand | 5 | 5 | 29.8 | 33 | 37.91 | -- | -- | -- | 29.8 | 33 | 37.91 |
| Medium sand | 5 | 5 | 18.95 | 24.7 | 34.82 | -- | -- | -- | 18.95 | 24.7 | 34.82 |
| Coarse sand | 5 | 5 | 2.95 | 6.84 | 11.8 | -- | -- | -- | 2.95 | 6.84 | 11.8 |
| Very coarse sand | 5 | 5 | 0.62 | 0.927 | 1.38 | -- | -- | -- | 0.62 | 0.927 | 1.38 |
| Very fine gravel | 5 | 5 | 0.31 | 1.42 | 2.51 | -- | -- | -- | 0.31 | 1.42 | 2.51 |
| Fine gravel | 5 | 5 | 0.84 | 3.28 | 7.28 | -- | -- | -- | 0.84 | 3.28 | 7.28 |
| Medium gravel | 5 | 5 | 0 | 3.34 | 11.7 | -- | -- | -- | 0 | 3.34 | 11.7 |
| Coarse gravel | 5 | 5 | 0 | 2.16 | 10.76 | -- | -- | -- | 0 | 2.16 | 10.76 |
| Very coarse gravel | 5 | 5 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| Cobbles | 5 | 5 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| RFB - Conventional Parameters |  |  |  |  |  |  |  |  |  |  |  |
| pH (SU) | 3 | 3 | 5.69 | 5.93 | 6.08 | -- | -- | -- | 5.69 | 5.93 | 6.08 |
| Solids (\%) | 3 | 3 | 84.6 | 87.6 | 93 | -- | -- | -- | 84.6 | 87.6 | 93 |
| Grain Size (\%) |  |  |  |  |  |  |  |  |  |  |  |
| Clay | 3 | 3 | 2.94 | 6.87 | 9.21 | -- | -- | -- | 2.94 | 6.87 | 9.21 |
| Silt | 3 | 3 | 31.4 | 38.8 | 50.5 | -- | -- | -- | 31.4 | 38.8 | 50.5 |
| Very fine sand | 3 | 3 | 12.91 | 20.2 | 29.36 | -- | -- | -- | 12.91 | 20.2 | 29.36 |
| Fine sand | 3 | 3 | 7.97 | 12.9 | 20.24 | -- | -- | -- | 7.97 | 12.9 | 20.24 |
| Medium sand | 3 | 3 | 4.15 | 6.34 | 8.63 | -- | -- | -- | 4.15 | 6.34 | 8.63 |
| Coarse sand | 3 | 3 | 2.72 | 3.56 | 4.92 | -- | -- | -- | 2.72 | 3.56 | 4.92 |
| Very coarse sand | 3 | 3 | 1.36 | 1.8 | 2.03 | -- | -- | -- | 1.36 | 1.8 | 2.03 |
| Very fine gravel | 3 | 3 | 0.07 | 0.826 | 1.6 | -- | -- | -- | 0.07 | 0.826 | 1.6 |
| Fine gravel | 3 | 3 | 0.47 | 0.813 | 1.09 | -- | -- | -- | 0.47 | 0.813 | 1.09 |
| Medium gravel | 3 | 3 | 0.11 | 2.59 | 7.53 | -- | -- | -- | 0.11 | 2.59 | 7.53 |
| Coarse gravel | 3 | 3 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| Very coarse gravel | 3 | 3 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| Cobbles | 3 | 3 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| RFC - Conventional Parameters |  |  |  |  |  |  |  |  |  |  |  |
| pH (SU) | 6 | 6 | 6.26 | 6.35 | 6.59 | -- | -- | -- | 6.26 | 6.35 | 6.59 |
| Solids (\%) | 6 | 6 | 83.7 | 87.4 | 91.2 | -- | -- | -- | 83.7 | 87.4 | 91.2 |
| Grain Size (\%) |  |  |  |  |  |  |  |  |  |  |  |
| Clay | 6 | 6 | 2.99 | 10.1 | 16.75 | -- | -- | -- | 2.99 | 10.1 | 16.75 |
| Fine sand | 6 | 6 | 5.11 | 9.68 | 14.4 | -- | -- | -- | 5.11 | 9.68 | 14.4 |
| Medium sand | 6 | 6 | 1.53 | 4.97 | 10.52 | -- | -- | -- | 1.53 | 4.97 | 10.52 |
| Coarse sand | 6 | 6 | 0.47 | 3.87 | 7.34 | -- | -- | -- | 0.47 | 3.87 | 7.34 |
| Fine gravel | 6 | 6 | 0 | 0.341 | 1.42 | -- | -- | -- | 0 | 0.341 | 1.42 |
| Medium gravel | 6 | 6 | 0 | 0.0439 | 0.0933 | -- | -- | -- | 0 | 0.0439 | 0.0933 |
| Coarse gravel | 6 | 6 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| Cobbles | 6 | 6 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |

Upper Columbia River
Soil Study Data Summary and Data Gap Report

Table 5-2a. Relict Floodplain Deposition Area Summary Statistics for Bulk Soil Sample Conventional Parameter Results

| Analyte | Number of Samples | Number of Detected Values | Minimum Detected Values | Mean Detected Values | Maximum Detected Values | Minimum Nondetected Value $^{\mathrm{a}}$ | Mean Nondetected Value ${ }^{\text {a }}$ | Maximum Nondetected Value ${ }^{\text {a }}$ | Overall Minimum Value ${ }^{\text {a }}$ | Overall Mean Value ${ }^{\text {a }}$ | Overall Maximum Value ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RFC - Conventional Parameters (continued) |  |  |  |  |  |  |  |  |  |  |  |
| Silt | 6 | 6 | 27.46 | 43.8 | 63.65 | -- | -- | -- | 27.46 | 43.8 | 63.65 |
| Very fine sand | 6 | 6 | 14.1 | 21.3 | 30.4 | -- | -- | -- | 14.1 | 21.3 | 30.4 |
| Very coarse sand | 6 | 6 | 0.453 | 2.11 | 3.5 | -- | -- | -- | 0.453 | 2.11 | 3.5 |
| Very fine gravel | 6 | 6 | 0.09 | 0.719 | 3.12 | -- | -- | -- | 0.09 | 0.719 | 3.12 |
| Very coarse gravel | 6 | 6 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| RFD - Conventional Parameters |  |  |  |  |  |  |  |  |  |  |  |
| pH (SU) | 2 | 2 | 6.24 | 6.74 | 7.23 | -- | -- | -- | 6.24 | 6.74 | 7.23 |
| Solids (\%) | 2 | 2 | 90.7 | 93.5 | 96.2 | -- | -- | -- | 90.7 | 93.5 | 96.2 |
| Grain Size (\%) |  |  |  |  |  |  |  |  |  |  |  |
| Clay | 2 | 2 | 0.77 | 1.24 | 1.71 | -- | -- | -- | 0.77 | 1.24 | 1.71 |
| Silt | 2 | 2 | 14.48 | 15.5 | 16.6 | -- | -- | -- | 14.48 | 15.5 | 16.6 |
| Very fine sand | 2 | 2 | 21.7 | 26.5 | 31.2 | -- | -- | -- | 21.7 | 26.5 | 31.2 |
| Fine sand | 2 | 2 | 26.8 | 32.5 | 38.25 | -- | -- | -- | 26.8 | 32.5 | 38.25 |
| Medium sand | 2 | 2 | 8.26 | 11.4 | 14.5 | -- | -- | -- | 8.26 | 11.4 | 14.5 |
| Coarse sand | 2 | 2 | 1.68 | 3.88 | 6.07 | -- | -- | -- | 1.68 | 3.88 | 6.07 |
| Very coarse sand | 2 | 2 | 1.17 | 2.28 | 3.39 | -- | -- | -- | 1.17 | 2.28 | 3.39 |
| Very fine gravel | 2 | 2 | 0.53 | 1.57 | 2.6 | -- | -- | -- | 0.53 | 1.57 | 2.6 |
| Fine gravel | 2 | 2 | 0.67 | 0.875 | 1.08 | -- | -- | -- | 0.67 | 0.875 | 1.08 |
| Medium gravel | 2 | 2 | 0 | 1.62 | 3.23 | -- | -- | -- | 0 | 1.62 | 3.23 |
| Coarse gravel | 2 | 2 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| Very coarse gravel | 2 | 2 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |
| Cobbles | 2 | 2 | 0 | 0 | 0 | -- | -- | -- | 0 | 0 | 0 |

Fotes:
For decision units (DUs) with field split and triplicate samples, summary statistics are based on the average of results for the DU. Nondetected values (NDs) are included as half the reporting limits (RLs). ${ }^{a}$ Calculated with nondetected results at one-half of the detection limit. RFA, RFB, RFC, RFD - relict flood plain depositional areas A, B, C, and D
Upper Columbia River
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Upper Columbia River
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Upper Columbia River
Soil Study Data Summary and Data Gap Report
Upper Columbia River
Soil Study Data Summary and Data Gap Report
Upper Columbia River
Soil Study Data Summary and Data Gap Report
Table 5-3b. Windblown Sediment Deposition Area Summary Statistics for < 2-mm Fraction Metals and Conventional Parameter Results

| Analyte | Number of Samples | Number of Detected Values | Minimum <br> Detected Values | Mean Detected Values | Maximum <br> Detected Values | Minimum Nondetected Value ${ }^{\text {a }}$ | Mean <br> Nondetected Value ${ }^{\text {a }}$ | Maximum Nondetected Value ${ }^{\text {a }}$ | Overall Minimum Value ${ }^{\text {a }}$ |  | Overall Maximum Value ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Columbia Beach North - Conventional Parameters |  |  |  |  |  |  |  |  |  |  |  |
| CEC (me/100 gm) | 4 | 4 | 7.28 | 8.63 | 11.5 | -- | -- | -- | 7.28 | 8.63 | 11.5 |
| Organic carbon (\%) | 4 | 4 | 0.89 | 1.28 | 1.76 | -- | -- | -- | 0.89 | 1.28 | 1.76 |
| Solids (\%) | 4 | 4 | 99.2 | 99.4 | 99.6 | -- | -- | -- | 99.2 | 99.4 | 99.6 |
| Columbia Beach North - Metals/Metalloids (mg/kg) |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | 4 | 4 | 10800 | 11600 | 13200 | -- | -- | -- | 10800 | 11600 | 13200 |
| Antimony | 4 | 4 | 0.21 | 0.257 | 0.287 | -- | -- | -- | 0.21 | 0.257 | 0.287 |
| Arsenic | 4 | 4 | 6.62 | 7.38 | 8.26 | -- | -- | -- | 6.62 | 7.38 | 8.26 |
| Barium | 4 | 4 | 122 | 141 | 158 | -- | -- | -- | 122 | 141 | 158 |
| Beryllium | 4 | 4 | 0.39 | 0.433 | 0.47 | -- | -- | -- | 0.39 | 0.433 | 0.47 |
| Cadmium | 4 | 4 | 0.205 | 0.242 | 0.3 | -- | -- | -- | 0.205 | 0.242 | 0.3 |
| Calcium | 4 | 4 | 2480 | 2810 | 3220 | -- | -- | -- | 2480 | 2810 | 3220 |
| Chromium | 4 | 4 | 14.5 | 15.9 | 18.8 | -- | -- | -- | 14.5 | 15.9 | 18.8 |
| Cobalt | 4 | 4 | 6.16 | 6.57 | 7.16 | -- | -- | -- | 6.16 | 6.57 | 7.16 |
| Copper | 4 | 4 | 12.9 | 14.4 | 16 | -- | -- | -- | 12.9 | 14.4 | 16 |
| Iron | 4 | 4 | 15800 | 16500 | 17100 | -- | -- | -- | 15800 | 16500 | 17100 |
| Lead | 4 | 4 | 9.49 | 10.7 | 12.1 | -- | -- | -- | 9.49 | 10.7 | 12.1 |
| Magnesium | 4 | 4 | 3580 | 3890 | 4310 | -- | -- | -- | 3580 | 3890 | 4310 |
| Manganese | 4 | 4 | 349 | 374 | 387 | -- | -- | -- | 349 | 374 | 387 |
| Mercury | 4 | 4 | 0.0085 | 0.00913 | 0.01 | -- | -- | -- | 0.0085 | 0.00913 | 0.01 |
| Molybdenum | 4 | 4 | 0.46 | 0.521 | 0.58 | -- | -- | -- | 0.46 | 0.521 | 0.58 |
| Nickel | 4 | 4 | 12.7 | 13.5 | 15.6 | -- | -- | -- | 12.7 | 13.5 | 15.6 |
| Potassium | 4 | 4 | 1890 | 2280 | 2640 | -- | -- | -- | 1890 | 2280 | 2640 |
| Selenium | 4 | 4 | 0.08 | 0.0846 | 0.09 | -- | -- | -- | 0.08 | 0.0846 | 0.09 |
| Silver | 4 | 4 | 0.035 | 0.0429 | 0.05 | -- | -- | -- | 0.035 | 0.0429 | 0.05 |
| Sodium | 4 | 4 | 88.3 | 104 | 120 | -- | -- | -- | 88.3 | 104 | 120 |
| Thallium | 4 | 4 | 0.11 | 0.118 | 0.12 | -- | -- | -- | 0.11 | 0.118 | 0.12 |
| Vanadium | 4 | 4 | 24.1 | 25.5 | 27 | -- | -- | -- | 24.1 | 25.5 | 27 |
| Zinc | 4 | 4 | 48.9 | 52.9 | 59 | -- | -- | -- | 48.9 | 52.9 | 59 |
| Columbia Beach South - Conventional Parameters |  |  |  |  |  |  |  |  |  |  |  |
| CEC (me/100 gm) | 2 | 2 | 7.08 | 7.41 | 7.73 | -- | -- | -- | 7.08 | 7.41 | 7.73 |
| Organic carbon (\%) | 2 | 2 | 1.02 | 1.19 | 1.35 | -- | -- | -- | 1.02 | 1.19 | 1.35 |
| Solids (\%) | 2 | 2 | 99.4 | 99.5 | 99.5 | -- | -- | -- | 99.4 | 99.5 | 99.5 |
| Columbia Beach South - Metals/Metalloids (mg/kg) |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | 2 | 2 | 9110 | 9860 | 10600 | -- | -- | -- | 9110 | 9860 | 10600 |
| Antimony | 2 | 2 | 0.17 | 0.18 | 0.19 | -- | -- | -- | 0.17 | 0.18 | 0.19 |
| Arsenic | 2 | 2 | 5.28 | 5.43 | 5.58 | -- | -- | -- | 5.28 | 5.43 | 5.58 |
| Barium | 2 | 2 | 95.1 | 105 | 115 | -- | -- | -- | 95.1 | 105 | 115 |
| Beryllium | 2 | 2 | 0.34 | 0.355 | 0.37 | -- | -- | -- | 0.34 | 0.355 | 0.37 |
| Cadmium | 2 | 2 | 0.18 | 0.18 | 0.18 | -- | -- | -- | 0.18 | 0.18 | 0.18 |
| Calcium | 2 | 2 | 2140 | 2360 | 2580 | -- | -- | -- | 2140 | 2360 | 2580 |
| Chromium | 2 | 2 | 11.9 | 13.5 | 15 | -- | -- | -- | 11.9 | 13.5 | 15 |
| Cobalt | 2 | 2 | 4.88 | 5.33 | 5.78 | -- | -- | -- | 4.88 | 5.33 | 5.78 |
| Copper | 2 | 2 | 11.1 | 11.9 | 12.7 | -- | -- | -- | 11.1 | 11.9 | 12.7 |
| Iron | 2 | 2 | 13500 | 14000 | 14500 | -- | -- | -- | 13500 | 14000 | 14500 |
| Lead | 2 | 2 | 7.91 | 8.04 | 8.16 | -- | -- | -- | 7.91 | 8.04 | 8.16 |
| Magnesium | 2 | 2 | 3400 | 3490 | 3580 | -- | -- | -- | 3400 | 3490 | 3580 |
| Manganese | 2 | 2 | 306 | 307 | 307 | -- | -- | -- | 306 | 307 | 307 |
| Mercury | 2 | 2 | 0.007 | 0.0075 | 0.008 | -- | -- | -- | 0.007 | 0.0075 | 0.008 |
| Molybdenum | 2 | 2 | 0.37 | 0.38 | 0.39 | -- | -- | -- | 0.37 | 0.38 | 0.39 |
| Nickel | 2 | 2 | 10.3 | 11.5 | 12.6 | -- | -- | -- | 10.3 | 11.5 | 12.6 |
| Potassium | 2 | 2 | 1790 | 1880 | 1960 | -- | -- | -- | 1790 | 1880 | 1960 |

Upper Columbia River
Soil Study Data Summary and Data Gap Report
Upper Columbia River
Soil Study Data Summary and Data Gap Report

| Analyte | Soil Fraction | Number of Samples | Number of Detected Values | Minimum Detected Values | Mean Detected Values | Maximum Detected Values | $\qquad$ | Mean Nondetected Value $^{\mathrm{a}}$ | Maximum Nondetected Value ${ }^{\text {a }}$ | Overall Minimum Value ${ }^{\text {a }}$ | Overall Mean Value ${ }^{\text {a }}$ | Overall Maximum Value ${ }^{\text {a }}$ | Units |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ADA - High-density |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | < $149-\mu \mathrm{m}$ | 11 | 11 | 14.8 | 20.9 | 28.4 | -- | -- | -- | 14.8 | 20.9 | 28.4 | \% |
| Antimony | < $149-\mu \mathrm{m}$ | 11 | 11 | 19.7 | 37 | 53.7 | -- | -- | -- | 19.7 | 37 | 53.7 | \% |
| Arsenic | < $149-\mu \mathrm{m}$ | 11 | 11 | 26.3 | 36.4 | 49.2 | -- | -- | -- | 26.3 | 36.4 | 49.2 | \% |
| Barium | < $149-\mu \mathrm{m}$ | 11 | 11 | 45.8 | 61.8 | 69.7 | -- | -- | -- | 45.8 | 61.8 | 69.7 | \% |
| Beryllium | < $149-\mu \mathrm{m}$ | 11 | 11 | 27.8 | 30.6 | 32.7 | -- | -- | -- | 27.8 | 30.6 | 32.7 | \% |
| Cadmium | < $149-\mu \mathrm{m}$ | 11 | 11 | 68 | 81 | 91.9 | -- | -- | -- | 68 | 81 | 91.9 | \% |
| Calcium | < $149-\mu \mathrm{m}$ | 11 | 11 | 60.5 | 73.6 | 85.2 | -- | -- | -- | 60.5 | 73.6 | 85.2 | \% |
| Chromium | < $149-\mu \mathrm{m}$ | 11 | 11 | 2.1 | 3.56 | 4.9 | -- | -- | -- | 2.1 | 3.56 | 4.9 | \% |
| Cobalt | < $149-\mu \mathrm{m}$ | 11 | 11 | 13.6 | 16.6 | 21.3 | -- | -- | -- | 13.6 | 16.6 | 21.3 | \% |
| Copper | < $149-\mu \mathrm{m}$ | 11 | 11 | 20.6 | 26.9 | 40 | -- | -- | -- | 20.6 | 26.9 | 40 | \% |
| Iron | < $149-\mu \mathrm{m}$ | 11 | 11 | 4.3 | 6.51 | 8.8 | -- | -- | -- | 4.3 | 6.51 | 8.8 | \% |
| Lead | < $149-\mu \mathrm{m}$ | 11 | 11 | 67.4 | 85.2 | 95.9 | -- | -- | -- | 67.4 | 85.2 | 95.9 | \% |
| Magnesium | < $149-\mu \mathrm{m}$ | 11 | 11 | 1.9 | 4.35 | 6.6 | -- | -- | -- | 1.9 | 4.35 | 6.6 | \% |
| Manganese | < $149-\mu \mathrm{m}$ | 11 | 11 | 37.6 | 50.4 | 58.2 | -- | -- | -- | 37.6 | 50.4 | 58.2 | \% |
| Mercury | < $149-\mu \mathrm{m}$ | 11 | 9 | 2.8 | 6.4 | 9.5 | -- | -- | -- | 2.8 | 6.4 | 9.5 | \% |
| Molybdenum | < $149-\mu \mathrm{m}$ | 11 | 10 | 2.9 | 5.36 | 11.1 | -- | -- | -- | 2.9 | 5.36 | 11.1 | \% |
| Nickel | < $149-\mu \mathrm{m}$ | 11 | 11 | 3.4 | 5.75 | 9.4 | -- | -- | -- | 3.4 | 5.75 | 9.4 | \% |
| Potassium | < $149-\mu \mathrm{m}$ | 11 | 11 | 13.1 | 20 | 27.5 | -- | -- | -- | 13.1 | 20 | 27.5 | \% |
| Selenium | < $149-\mu \mathrm{m}$ | 11 | 1 | 11.8 | 11.8 | 11.8 | -- | -- | -- | 11.8 | 11.8 | 11.8 | \% |
| Silver | < $149-\mu \mathrm{m}$ | 11 | 11 | 53.1 | 60.4 | 67.5 | -- | -- | -- | 53.1 | 60.4 | 67.5 | \% |
| Sodium | < $149-\mu \mathrm{m}$ | 11 | 11 | 4 | 9.65 | 14.2 | -- | -- | -- | 4 | 9.65 | 14.2 | \% |
| Thallium | < $149-\mu \mathrm{m}$ | 11 | 11 | 13.2 | 19 | 26.8 | -- | -- | -- | 13.2 | 19 | 26.8 | \% |
| Vanadium | < $149-\mu \mathrm{m}$ | 11 | 11 | 3 | 4.76 | 5.8 | -- | -- | -- | 3 | 4.76 | 5.8 | \% |
| Zinc | < $149-\mu \mathrm{m}$ | 11 | 11 | 28.5 | 45.9 | 52.2 | -- | -- | -- | 28.5 | 45.9 | 52.2 | \% |
| ADA - Primary |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | < $149-\mu \mathrm{m}$ | 11 | 11 | 9.4 | 18.6 | 30.6 | -- | -- | -- | 9.4 | 18.6 | 30.6 | \% |
| Antimony | < $149-\mu \mathrm{m}$ | 11 | 11 | 21.6 | 30.2 | 37.6 | -- | -- | -- | 21.6 | 30.2 | 37.6 | \% |
| Arsenic | < $149-\mu \mathrm{m}$ | 11 | 11 | 14.9 | 27.8 | 42.9 | -- | -- | -- | 14.9 | 27.8 | 42.9 | \% |
| Barium | < $149-\mu \mathrm{m}$ | 11 | 11 | 54.3 | 62.7 | 76.6 | -- | -- | -- | 54.3 | 62.7 | 76.6 | \% |
| Beryllium | < $149-\mu \mathrm{m}$ | 11 | 11 | 22.1 | 31.6 | 38.5 | -- | -- | -- | 22.1 | 31.6 | 38.5 | \% |
| Cadmium | < $149-\mu \mathrm{m}$ | 11 | 11 | 70.5 | 82.4 | 87.1 | -- | -- | -- | 70.5 | 82.4 | 87.1 | \% |
| Calcium | < $149-\mu \mathrm{m}$ | 11 | 11 | 58.9 | 76.1 | 96 | -- | -- | -- | 58.9 | 76.1 | 96 | \% |
| Chromium | < $149-\mu \mathrm{m}$ | 11 | 11 | 1.6 | 3.48 | 6.9 | -- | -- | -- | 1.6 | 3.48 | 6.9 | \% |
| Cobalt | < $149-\mu \mathrm{m}$ | 11 | 11 | 8.1 | 18.8 | 30.7 | -- | -- | -- | 8.1 | 18.8 | 30.7 | \% |
| Copper | < $149-\mu \mathrm{m}$ | 11 | 11 | 13.2 | 21.5 | 26.7 | -- | -- | -- | 13.2 | 21.5 | 26.7 | \% |
| Iron | < $149-\mu \mathrm{m}$ | 11 | 11 | 3.1 | 6.17 | 10.8 | -- | -- | -- | 3.1 | 6.17 | 10.8 | \% |
| Lead | < $149-\mu \mathrm{m}$ | 11 | 11 | 65 | 81.9 | 94 | -- | -- | -- | 65 | 81.9 | 94 | \% |
| Magnesium | < $149-\mu \mathrm{m}$ | 11 | 11 | 1.9 | 8.09 | 24.7 | -- | -- | -- | 1.9 | 8.09 | 24.7 | \% |
| Manganese | < $149-\mu \mathrm{m}$ | 11 | 11 | 38.6 | 52.8 | 74.5 | -- | -- | -- | 38.6 | 52.8 | 74.5 | \% |
| Mercury | < $149-\mu \mathrm{m}$ | 11 | 6 | 6.55 | 7.58 | 9 | -- | -- | -- | 6.55 | 7.58 | 9 | \% |
| Molybdenum | < $149-\mu \mathrm{m}$ | 9 | 7 | 3.8 | 5.45 | 7.2 | -- | -- | -- | 3.8 | 5.45 | 7.2 | \% |
| Nickel | < $149-\mu \mathrm{m}$ | 11 | 11 | 3.1 | 6.16 | 8.9 | -- | -- | -- | 3.1 | 6.16 | 8.9 | \% |
| Potassium | < $149-\mu \mathrm{m}$ | 11 | 11 | 15.2 | 21.6 | 30.3 | -- | -- | -- | 15.2 | 21.6 | 30.3 | \% |
| Selenium | < $149-\mu \mathrm{m}$ | 11 | 1 | 21.7 | 21.7 | 21.7 | -- | -- | -- | 21.7 | 21.7 | 21.7 | \% |
| Silver | < $149-\mu \mathrm{m}$ | 11 | 11 | 50.7 | 61.2 | 67 | -- | -- | -- | 50.7 | 61.2 | 67 | \% |
| Sodium | < $149-\mu \mathrm{m}$ | 11 | 11 | 5.4 | 8.64 | 16.4 | -- | -- | -- | 5.4 | 8.64 | 16.4 | \% |
| Thallium | < $149-\mu \mathrm{m}$ | 11 | 11 | 8.7 | 17.1 | 41.1 | -- | -- | -- | 8.7 | 17.1 | 41.1 | \% |
| Vanadium | < $149-\mu \mathrm{m}$ | 11 | 11 | 3.5 | 5.28 | 8.6 | -- | -- | -- | 3.5 | 5.28 | 8.6 | \% |
| Zinc | < $149-\mu \mathrm{m}$ | 11 | 11 | 19.4 | 33.6 | 48.2 | -- | -- | -- | 19.4 | 33.6 | 48.2 | \% |

Upper Columbia River
Soil Study Data Summary and Data Gap Report

| Analyte | Soil Fraction | Number of Samples | Number of Detected Values | Minimum <br> Detected Values | Mean Detected Values | Maximum <br> Detected Values | Minimum Nondetected Value ${ }^{\text {a }}$ | Mean <br> Nondetected Value ${ }^{\text {a }}$ | Maximum Nondetected Value ${ }^{\text {a }}$ | Overall Minimum Value ${ }^{\text {a }}$ | Overall Mean Value ${ }^{\text {a }}$ | Overall Maximum Value ${ }^{\text {a }}$ | Units |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RFDA - RFA |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | < $149-\mu \mathrm{m}$ | 2 | 2 | 14.7 | 16.1 | 17.5 | -- | -- | -- | 14.7 | 16.1 | 17.5 | \% |
| Antimony | < $149-\mu \mathrm{m}$ | 2 | 2 | 18 | 23.2 | 28.3 | -- | -- | -- | 18 | 23.2 | 28.3 | \% |
| Arsenic | < $149-\mu \mathrm{m}$ | 2 | 2 | 15.4 | 18.1 | 20.7 | -- | -- | -- | 15.4 | 18.1 | 20.7 | \% |
| Barium | < $149-\mu \mathrm{m}$ | 2 | 2 | 29.4 | 35.2 | 40.9 | -- | -- | -- | 29.4 | 35.2 | 40.9 | \% |
| Beryllium | < $149-\mu \mathrm{m}$ | 2 | 2 | 34.5 | 34.5 | 34.5 | -- | -- | -- | 34.5 | 34.5 | 34.5 | \% |
| Cadmium | < $149-\mu \mathrm{m}$ | 2 | 2 | 36.8 | 37.3 | 37.7 | -- | -- | -- | 36.8 | 37.3 | 37.7 | \% |
| Calcium | < $149-\mu \mathrm{m}$ | 2 | 2 | 48.4 | 49.2 | 50 | -- | -- | -- | 48.4 | 49.2 | 50 | \% |
| Chromium | < $149-\mu \mathrm{m}$ | 2 | 2 | 12.3 | 12.6 | 12.9 | -- | -- | -- | 12.3 | 12.6 | 12.9 | \% |
| Cobalt | < $149-\mu \mathrm{m}$ | 2 | 2 | 28.3 | 29.1 | 29.9 | -- | -- | -- | 28.3 | 29.1 | 29.9 | \% |
| Copper | < $149-\mu \mathrm{m}$ | 2 | 2 | 51.7 | 52.4 | 53 | -- | -- | -- | 51.7 | 52.4 | 53 | \% |
| Iron | < $149-\mu \mathrm{m}$ | 2 | 2 | 13.5 | 14.8 | 16.1 | -- | -- | -- | 13.5 | 14.8 | 16.1 | \% |
| Lead | < $149-\mu \mathrm{m}$ | 2 | 2 | 61.3 | 62.5 | 63.7 | -- | -- | -- | 61.3 | 62.5 | 63.7 | \% |
| Magnesium | < $149-\mu \mathrm{m}$ | 2 | 2 | 35.4 | 36.1 | 36.7 | -- | -- | -- | 35.4 | 36.1 | 36.7 | \% |
| Manganese | < $149-\mu \mathrm{m}$ | 2 | 2 | 53.4 | 55.1 | 56.7 | -- | -- | -- | 53.4 | 55.1 | 56.7 | \% |
| Mercury | < $149-\mu \mathrm{m}$ | 2 | 0 | na | na | na | -- | -- | -- | na | na | na | \% |
| Molybdenum | < $149-\mu \mathrm{m}$ | 2 | 2 | 2.6 | 3.25 | 3.9 | -- | -- | -- | 2.6 | 3.25 | 3.9 | \% |
| Nickel | < $149-\mu \mathrm{m}$ | 2 | 2 | 11.2 | 11.8 | 12.3 | -- | -- | -- | 11.2 | 11.8 | 12.3 | \% |
| Potassium | < $149-\mu \mathrm{m}$ | 2 | 2 | 9.1 | 10.9 | 12.6 | -- | -- | -- | 9.1 | 10.9 | 12.6 | \% |
| Selenium | < $149-\mu \mathrm{m}$ | 2 | 1 | 13 | 13 | 13 | -- | -- | -- | 13 | 13 | 13 | \% |
| Silver | < $149-\mu \mathrm{m}$ | 2 | 2 | 18 | 21.1 | 24.2 | -- | -- | -- | 18 | 21.1 | 24.2 | \% |
| Sodium | < $149-\mu \mathrm{m}$ | 2 | 2 | 28.1 | 29.8 | 31.5 | -- | -- | -- | 28.1 | 29.8 | 31.5 | \% |
| Thallium | < $149-\mu \mathrm{m}$ | 2 | 2 | 10.9 | 12.1 | 13.2 | -- | -- | -- | 10.9 | 12.1 | 13.2 | \% |
| Vanadium | < $149-\mu \mathrm{m}$ | 2 | 2 | 14.7 | 15.1 | 15.4 | -- | -- | -- | 14.7 | 15.1 | 15.4 | \% |
| Zinc | < $149-\mu \mathrm{m}$ | 2 | 2 | 42.7 | 43.2 | 43.7 | -- | -- | -- | 42.7 | 43.2 | 43.7 | \% |
| RFDA - RFD |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aluminum | < $149-\mu \mathrm{m}$ | 1 | 1 | 17.6 | 17.6 | 17.6 | -- | -- | -- | 17.6 | 17.6 | 17.6 | \% |
| Antimony | < $149-\mu \mathrm{m}$ | 1 | 1 | 27.4 | 27.4 | 27.4 | -- | -- | -- | 27.4 | 27.4 | 27.4 | \% |
| Arsenic | < $149-\mu \mathrm{m}$ | 1 | 1 | 27.3 | 27.3 | 27.3 | -- | -- | -- | 27.3 | 27.3 | 27.3 | \% |
| Barium | < $149-\mu \mathrm{m}$ | 1 | 1 | 26.4 | 26.4 | 26.4 | -- | -- | -- | 26.4 | 26.4 | 26.4 | \% |
| Beryllium | < $149-\mu \mathrm{m}$ | 1 | 1 | 31.8 | 31.8 | 31.8 | -- | -- | -- | 31.8 | 31.8 | 31.8 | \% |
| Cadmium | < $149-\mu \mathrm{m}$ | 1 | 1 | 43.4 | 43.4 | 43.4 | -- | -- | -- | 43.4 | 43.4 | 43.4 | \% |
| Calcium | < $149-\mu \mathrm{m}$ | 1 | 1 | 48 | 48 | 48 | -- | -- | -- | 48 | 48 | 48 | \% |
| Chromium | < $149-\mu \mathrm{m}$ | 1 | 1 | 11.4 | 11.4 | 11.4 | -- | -- | -- | 11.4 | 11.4 | 11.4 | \% |
| Cobalt | < $149-\mu \mathrm{m}$ | 1 | 1 | 25.4 | 25.4 | 25.4 | -- | -- | -- | 25.4 | 25.4 | 25.4 | \% |
| Copper | < $149-\mu \mathrm{m}$ | 1 | 1 | 38.6 | 38.6 | 38.6 | -- | -- | -- | 38.6 | 38.6 | 38.6 | \% |
| Iron | < $149-\mu \mathrm{m}$ | 1 | 1 | 21.1 | 21.1 | 21.1 | -- | -- | -- | 21.1 | 21.1 | 21.1 | \% |
| Lead | < $149-\mu \mathrm{m}$ | 1 | 1 | 55.2 | 55.2 | 55.2 | -- | -- | -- | 55.2 | 55.2 | 55.2 | \% |
| Magnesium | < $149-\mu \mathrm{m}$ | 1 | 1 | 37.8 | 37.8 | 37.8 | -- | -- | -- | 37.8 | 37.8 | 37.8 | \% |
| Manganese | < $149-\mu \mathrm{m}$ | 1 | 1 | 56.2 | 56.2 | 56.2 | -- | -- | -- | 56.2 | 56.2 | 56.2 | \% |
| Mercury | < $149-\mu \mathrm{m}$ | 1 | 1 | 0.9 | 0.9 | 0.9 | -- | -- | -- | 0.9 | 0.9 | 0.9 | \% |
| Molybdenum | < $149-\mu \mathrm{m}$ | 1 | 1 | 3.4 | 3.4 | 3.4 | -- | -- | -- | 3.4 | 3.4 | 3.4 | \% |
| Nickel | < $149-\mu \mathrm{m}$ | 1 | 1 | 13.1 | 13.1 | 13.1 | -- | -- | -- | 13.1 | 13.1 | 13.1 | \% |
| Potassium | < $149-\mu \mathrm{m}$ | 1 | 1 | 14.9 | 14.9 | 14.9 | -- | -- | -- | 14.9 | 14.9 | 14.9 | \% |
| Selenium | < $149-\mu \mathrm{m}$ | 1 | 0 | na | na | na | -- | -- | -- | na | na | na | \% |
| Silver | < $149-\mu \mathrm{m}$ | 1 | 1 | 15.5 | 15.5 | 15.5 | -- | -- | -- | 15.5 | 15.5 | 15.5 | \% |
| Sodium | < $149-\mu \mathrm{m}$ | 1 | 1 | 35.5 | 35.5 | 35.5 | -- | -- | -- | 35.5 | 35.5 | 35.5 | \% |
| Thallium | < $149-\mu \mathrm{m}$ | 1 | 1 | 10.4 | 10.4 | 10.4 | -- | -- | -- | 10.4 | 10.4 | 10.4 | \% |
| Vanadium | < $149-\mu \mathrm{m}$ | 1 | 1 | 19.2 | 19.2 | 19.2 | -- | -- | -- | 19.2 | 19.2 | 19.2 | \% |
| Zinc | < $149-\mu \mathrm{m}$ | 1 | 1 | 53 | 53 | 53 | -- | -- | -- | 53 | 53 | 53 | \% |

For decision units (DUs) with field split and triplicate samples, summary statistics are based on the average of results for the DU. Nondetected values (NDs) are included as half the reporting limits (RLs). ADA- Aerial deposition area
BA - in vitro RFDA - relict flood plain deposition area

Table 5-5. Relative Bioavailability (RBA) Data for Lead from the $<149-\mu \mathrm{m}$ Fraction

| Decision Unit | Bioavailable Percentage Overall (IVBA) | Qualifier | Decision Unit RBA ${ }^{\text {a,b }}$ | Ratio of Site-specific RBA to EPA Default RBA ${ }^{\text {c }}$ |
| :---: | :---: | :---: | :---: | :---: |
| ADA - High-density |  |  |  |  |
| ADA-125 | 87.7 | J | 74.2 | 1.2 |
| ADA-126 | 89.1 | J | 75.4 | 1.3 |
| ADA-141 ${ }^{\text {d }}$ | 82.9 | J | 70.0 | 1.2 |
| ADA-142 | 91.0 | J | 77.1 | 1.3 |
| ADA-144 | 79.6 | J | 67.1 | 1.1 |
| ADA-145 | 83.3 | J | 70.3 | 1.2 |
| ADA-150 | 79.8 | J | 67.3 | 1.1 |
| ADA-152 | 95.9 | J | 81.4 | 1.4 |
| ADA-160 | 87.6 | J | 74.1 | 1.2 |
| ADA-161 ${ }^{\text {d }}$ | 93.0 | J | 78.9 | 1.3 |
| ADA-162 | 67.4 |  | 56.4 | 0.94 |
| Average for ADA High-density |  |  | 72.0 | 1.2 |
| ADA - Primary |  |  |  |  |
| ADA-001 | 87.2 | J | 73.8 | 1.2 |
| ADA-016-A | 88.6 |  | 70.4 | 1.2 |
| ADA-016-B | 83.8 |  | 70.4 | 1.2 |
| ADA-016-C | 77.8 |  | 70.4 | 1.2 |
| ADA-016 ${ }^{\text {e }}$ | 77.8 |  | 70.4 | 1.2 |
| ADA-035 | 83.6 | J | 70.6 | 1.2 |
| ADA-047 | 85.9 | J | 72.6 | 1.2 |
| ADA-048 | 82.3 | J | 69.5 | 1.2 |
| ADA-057 | 70.7 |  | 59.3 | 1.0 |
| ADA-059 | 79.1 | J | 66.6 | 1.1 |
| ADA-061 | 65.0 |  | 54.3 | 0.90 |
| ADA-076 | 94.0 | J | 79.7 | 1.3 |
| ADA-081 | 80.1 | J | 67.5 | 1.1 |
| ADA-096 | 89.9 | J | 76.1 | 1.3 |
| Average for ADA Primary |  |  | 69.1 | 1.2 |
| Average for ADA Overall' |  |  | 70.6 | 1.2 |
| RFDA |  |  |  |  |
| RFA-001-A | 67.0 |  | 53.1 | 0.89 |
| RFA-001-B | 65.8 |  | 53.1 | 0.89 |
| RFA-001-C | 58.3 |  | 53.1 | 0.89 |
| RFA-001 ${ }^{\text {e }}$ | 58.3 |  | 53.1 | 0.89 |
| RFA-005 | 61.3 |  | 51.0 | 0.85 |
| RFD-002 | 55.2 |  | 45.7 | 0.76 |
| Average for RFDA ${ }^{\dagger}$ |  |  | 49.9 | 0.8 |

## Notes:

${ }^{\text {a }}$ RBA equation from EPA (2007-lead estimation guidance) RBA $=0.878^{*}$ IVBA-0.028
${ }^{b}$ For decision units (DUs) with triplicate samples (ADA-016 and RFA-001), triplicate sample results were RBA adjusted and then averaged.
${ }^{c}$ EPA default RBA $=60 \%$ (EPA 2007a), empirical lead soil concentrations are multiplied by this ratio before comparison to the human health soil screening value to account for differences in bioavailability relative to the the screening value.
${ }^{d}$ Analysis of bioaccessibility percentage was performed on the split sample rather than the primary sample from ADA-141 (i.e., DIRT-011 $149 \mu \mathrm{~m}$ ) and from ADA-161 (i.e., DIRT-015 $149 \mu \mathrm{~m}$ ).
${ }^{e}$ Values are averages of the preceeding triplicate samples for that DU.
${ }^{\mathrm{f}}$ The average value for the area is calculated using the average of the triplicate values and not the individual values.
ADA - aerial deposition area
IVBA - in vitro bioaccessibility
RFDA - relict floodplain deposition area

Table 5-6. Lead Data from < 149- $\mu \mathrm{m}$ Fraction Adjusted for Bioavailability

| Decision Unit | Bioavailability Adjusted Lead Concentration ${ }^{2}$ $(\mathrm{mg} / \mathrm{kg})$ | Qualifier ${ }^{\text {b }}$ |
| :---: | :---: | :---: |
| ADA - High Density |  |  |
| ADA-124 | 178 |  |
| ADA-125 | 151 |  |
| ADA-126 | 427 |  |
| ADA-127 | 214 |  |
| ADA-128 | 365 |  |
| ADA-131 | 387 |  |
| ADA-132 | 163 |  |
| ADA-133 | 340 |  |
| ADA-135 | 141 |  |
| ADA-136 | 387 |  |
| ADA-139 | 222 |  |
| ADA-140 | 560 |  |
| ADA-141 | 191 |  |
| ADA-142 | 599 |  |
| ADA-143 | 279 |  |
| ADA-144 | 747 |  |
| ADA-145 | 373 |  |
| ADA-146 | 394 |  |
| ADA-147 | 772 |  |
| ADA-148 | 373 |  |
| ADA-150 | 627 |  |
| ADA-151 | 440 |  |
| ADA-152 | 300 |  |
| ADA-153 | 450 |  |
| ADA-154 | 399 |  |
| ADA-155 | 321 |  |
| ADA-156 | 385 |  |
| ADA-158 | 391 |  |
| ADA-159 | 520 |  |
| ADA-160 | 348 |  |
| ADA-161 | 452 |  |
| ADA-162 | 928 |  |
| ADA-164 | 496 |  |
| ADA-165 | 567 |  |
| ADA-168 | 315 |  |
| ADA - Primary |  |  |
| ADA-001 | 354 |  |
| ADA-002 | 61 |  |
| ADA-004 | 162 |  |
| ADA-005 | 85 | J |
| ADA-006 | 272 |  |
| ADA-008 | 349 |  |
| ADA-010 | 529 |  |
| ADA-015 | 187 |  |
| ADA-016 | 258 |  |
| ADA-017 | 292 |  |
| ADA-018 | 529 |  |
| ADA-019 | 115 |  |
| ADA-020 | 105 |  |
| ADA-021 | 107 |  |
| ADA-023 | 178 |  |
| ADA-024 | 513 |  |

Table 5-6. Lead Data from < 149- $\mu \mathrm{m}$ Fraction Adjusted for Bioavailability

| Decision Unit | Bioavailability Adjusted Lead Concentration ${ }^{2}$ $(\mathrm{mg} / \mathrm{kg})$ | Qualifier ${ }^{\text {b }}$ |
| :---: | :---: | :---: |
| ADA - Primary (continued) |  |  |
| ADA-025 | 245 |  |
| ADA-026 | 85 |  |
| ADA-028 | 243 |  |
| ADA-033 | 89 |  |
| ADA-034 | 77 |  |
| ADA-035 | 215 |  |
| ADA-039 | 41 |  |
| ADA-042 | 96 |  |
| ADA-043 | 178 |  |
| ADA-044 | 165 |  |
| ADA-045 | 381 |  |
| ADA-046 | 212 |  |
| ADA-047 | 326 |  |
| ADA-048 | 144 |  |
| ADA-049 | 54 |  |
| ADA-050 | 432 |  |
| ADA-051 | 124 |  |
| ADA-052 | 415 |  |
| ADA-053 | 103 |  |
| ADA-054 | 505 |  |
| ADA-055 | 211 | J |
| ADA-056 | 85 |  |
| ADA-057 | 132 |  |
| ADA-058 | 94 |  |
| ADA-059 | 143 |  |
| ADA-060 | 478 |  |
| ADA-061 | 124 |  |
| ADA-062 | 348 |  |
| ADA-063 | 152 |  |
| ADA-064 | 167 |  |
| ADA-065 | 385 |  |
| ADA-066 | 168 |  |
| ADA-067 | 107 |  |
| ADA-070 | 245 |  |
| ADA-071 | 276 |  |
| ADA-073 | 312 |  |
| ADA-076 | 286 |  |
| ADA-078 | 256 |  |
| ADA-079 | 151 |  |
| ADA-081 | 187 | J |
| ADA-082 | 194 |  |
| ADA-084 | 175 |  |
| ADA-085 | 135 |  |
| ADA-088 | 327 |  |
| ADA-089 | 331 |  |
| ADA-090 | 248 |  |
| ADA-091 | 282 |  |
| ADA-092 | 225 |  |
| ADA-093 | 182 |  |
| ADA-094 | 100 |  |
| ADA-095 | 165 | J |
| ADA-096 | 364 |  |

Table 5-6. Lead Data from < 149- $\mu \mathrm{m}$ Fraction Adjusted for Bioavailability

| Decision Unit | Bioavailability Adjusted Lead Concentration ${ }^{\text {a }}$ $(\mathrm{mg} / \mathrm{kg})$ | Qualifier ${ }^{\text {b }}$ |
| :---: | :---: | :---: |
| ADA - Primary (continued) |  |  |
| ADA-097 | 536 |  |
| ADA-099 | 259 |  |
| ADA-101 | 226 |  |
| ADA-102 | 127 | J |
| ADA-103 | 174 |  |
| ADA-104 | 80 |  |
| ADA-105 | 235 |  |
| ADA-106 | 80 |  |
| ADA-107 | 148 |  |
| ADA-108 | 228 |  |
| ADA-109 | 381 |  |
| ADA-110 | 267 |  |
| ADA-111 | 120 |  |
| ADA-112 | 103 |  |
| ADA-113 | 163 |  |
| ADA-114 | 115 |  |
| ADA-115 | 143 |  |
| ADA-116 | 113 | J |
| ADA-117 | 194 |  |
| ADA-118 | 192 |  |
| ADA-119 | 207 |  |
| ADA-121 | 240 |  |
| ADA-122 | 57 |  |
| ADA-169 | 69 |  |
| ADA-170 | 109 |  |
| ADA-171 | 95 |  |
| ADA-172 | 70 |  |
| ADA-173 | 90 |  |
| ADA-174 | 77 |  |
| ADA-175 | 83 |  |
| ADA-176 | 79 |  |
| ADA-177 | 72 |  |
| ADA-178 | 56 |  |
| ADA-179 | 67 |  |
| ADA-180 | 133 |  |
| ADA-181 | 101 |  |
| ADA-182 | 41 |  |
| ADA-183 | 58 |  |
| ADA-184 | 85 |  |
| RFDA - RFA |  |  |
| RFA-001 | 414 |  |
| RFA-002 | 252 |  |
| RFA-003 | 388 |  |
| RFA-004 | 290 |  |
| RFA-005 | 308 |  |
| RFDA - RFB |  |  |
| RFB-002 | 159 |  |
| RFB-003 | 107 | J |
| RFB-008 | 60 |  |

Table 5-6. Lead Data from < 149- $\mu \mathrm{m}$ Fraction Adjusted for Bioavailability

| Decision Unit | Bioavailability Adjusted Lead Concentration <br>  <br> $(\mathrm{mg} / \mathrm{kg})$ | Qualifier $^{\mathrm{b}}$ |
| :--- | :---: | :---: | | RFDA - RFC | 386 |
| :--- | :--- |
| RFC-003 | 199 |
| RFC-004 | 470 |
| RFC-005 | 343 |
| RFC-006 | 362 |
| RFC-007 | 339 |
| RFC-008 |  |
| RFDA - RFD | 314 |
| RFD-002 | 307 |
| RFD-003 |  |

## Notes:

${ }^{a}$ Lead concentrations adjusted for the ratio of site-specific relative bioavailability (RBA) to EPA's default RBA, see Table 5-5. The ratio of the DU-specific RBA to EPA's default RBA was used when available. For ADA and RFDA DUs that did not have IVBA measured directly (i.e. those DUs not listed in Table 5-5), the average RBA ratio for the ADA overall (including primary and high density) or RFDA reported in Table 5-5 was applied.
${ }^{\mathrm{b}}$ Qualifiers are from lead data before the adjustment for bioavailability.
ADA - aerial deposition area
$J$ - estimated value
mg/kg - milligram per kilogram
RFDA - relict floodplain deposition area
RFA, RFB, RFC, and RFD- relict flood plain depositional areas $A, B, C$, and $D$

Table 5-7a. Aerial Deposition Area Summary of Field Split and Triplicate Sample Results for Bulk Soil Samples

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | No. RPDs >20\% | Max RPD (\%) | Number of Samples | No. RSDs > $35 \%$ | Max RSD (\%) |
| High-density - Conventional Parameters |  |  |  |  |  |  |
| pH | 6 | 0 | 18.8 | 6 | 0 | 5.07 |
| Solids | 6 | 0 | 6.95 | 6 | 0 | 4.93 |
| Grain Size |  |  |  |  |  |  |
| Clay | 6 | na | 64 | 6 | 2 | 51.3 |
| Silt | 6 | na | 16.6 | 6 | 0 | 18.7 |
| Very fine sand | 6 | na | 34.4 | 6 | 0 | 34.4 |
| Fine sand | 6 | na | 30.8 | 6 | 1 | 43.5 |
| Medium sand | 6 | na | 24.5 | 6 | 1 | 46.3 |
| Coarse sand | 6 | na | 14.9 | 6 | 1 | 35.7 |
| Very coarse sand | 6 | na | 34.5 | 6 | 1 | 38.3 |
| Very fine gravel | 6 | na | 56.3 | 6 | 2 | 77 |
| Fine gravel | 6 | na | 139 | 6 | 6 | 111 |
| Medium gravel | 6 | na | 200 | 6 | 4 | 173 |
| Coarse gravel | 6 | na | 200 | 6 | 0 | 0 |
| Very coarse gravel | 6 | na | 0 | 6 | 0 | 0 |
| Cobbles | 6 | na | 0 | 6 | 0 | 0 |
| Primary - Conventional Parameters |  |  |  |  |  |  |
| pH | 12 | 0 | 3.05 | 10 | 0 | 5.63 |
| Solids | 12 | 0 | 5.61 | 10 | 0 | 5.97 |
| Grain Size |  |  |  |  |  |  |
| Clay | 12 | na | 64.9 | 10 | 1 | 94.2 |
| Silt | 12 | na | 28.3 | 10 | 0 | 32.1 |
| Very fine sand | 12 | na | 106 | 10 | 0 | 24.9 |
| Fine sand | 12 | na | 45 | 10 | 0 | 27.7 |
| Medium sand | 12 | na | 49.4 | 10 | 0 | 20.1 |
| Coarse sand | 12 | na | 35.8 | 10 | 0 | 34.4 |
| Very coarse sand | 12 | na | 32.4 | 10 | 0 | 27.5 |
| Very fine gravel | 12 | na | 63.8 | 10 | 2 | 88.6 |
| Fine gravel | 12 | na | 156 | 10 | 8 | 122 |
| Medium gravel | 12 | na | 200 | 10 | 9 | 173 |
| Coarse gravel | 12 | na | 200 | 10 | 0 | 0 |
| Very coarse gravel | 12 | na | 0 | 10 | 0 | 0 |
| Cobbles | 12 | na | 0 | 10 | 0 | 0 |

## Notes:

Highlighted cells identify where relative percent differences (RPDs) and relative standard deviations (RSDs) are greater than the control limit.
Control limits specified in the quality assurance project plan (QAPP) (Exponent et al. 2014) are 20\% for analytical RPDs (i.e., metals, mercury, total organic carbon [TOC], cation exchange capacity [CEC], and pH ) and $35 \%$ for field triplicate RSDs. The QAPP did not specify a quality objective for grain size RPDs.
na - not applicable

Table 5-7b. Aerial Deposition Area Summary of Field Split and Triplicate Sample Results for the < 2-mm Soil Fraction

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | $\begin{aligned} & \text { No. RPDs } \\ & >20 \% \end{aligned}$ | Max RPD (\%) | Number of Samples | $\begin{gathered} \text { No. RSDs } \\ >35 \% \end{gathered}$ | Max RSD (\%) |
| High-density - Conventional Parameters |  |  |  |  |  |  |
| CEC | 6 | 3 | 32.5 | 6 | 0 | 28.2 |
| Organic carbon | 6 | 2 | 23.5 | 6 | 0 | 26.6 |
| Solids | 6 | 0 | 3.16 | 6 | 0 | 2.89 |
| High-density - Metals/Metalloids |  |  |  |  |  |  |
| Aluminum | 6 | 1 | 41.2 | 6 | 0 | 14.2 |
| Antimony | 6 | 1 | 27.4 | 6 | 1 | 35.2 |
| Arsenic | 6 | 1 | 35.1 | 6 | 0 | 26 |
| Barium | 6 | 1 | 29.8 | 6 | 0 | 10.8 |
| Beryllium | 6 | 1 | 36.6 | 6 | 0 | 12.4 |
| Cadmium | 6 | 1 | 30.4 | 6 | 0 | 24.9 |
| Calcium | 6 | 1 | 36.3 | 6 | 0 | 14.2 |
| Chromium | 6 | 1 | 37.9 | 6 | 0 | 23.5 |
| Cobalt | 6 | 1 | 35.6 | 6 | 0 | 20.9 |
| Copper | 6 | 1 | 32.7 | 6 | 0 | 19.5 |
| Iron | 6 | 1 | 34.6 | 6 | 0 | 13.9 |
| Lead | 6 | 1 | 40.5 | 6 | 0 | 34.9 |
| Magnesium | 6 | 1 | 35.6 | 6 | 0 | 17.9 |
| Manganese | 6 | 2 | 33.4 | 6 | 0 | 14.5 |
| Mercury | 6 | 1 | 33.3 | 6 | 0 | 28.6 |
| Molybdenum | 6 | 1 | 32.1 | 6 | 0 | 14.2 |
| Nickel | 6 | 1 | 34.8 | 6 | 0 | 18 |
| Potassium | 6 | 1 | 38.8 | 6 | 0 | 21.9 |
| Selenium | 6 | 1 | 31.1 | 6 | 0 | 24 |
| Silver | 6 | 1 | 27 | 6 | 0 | 26.9 |
| Sodium | 6 | 1 | 43.9 | 6 | 0 | 27.8 |
| Thallium | 6 | 1 | 27.3 | 6 | 0 | 23.5 |
| Vanadium | 6 | 1 | 39.1 | 6 | 0 | 14.2 |
| Zinc | 6 | 1 | 26.9 | 6 | 0 | 24.1 |
| Primary - Conventional Parameters |  |  |  |  |  |  |
| CEC | 12 | 8 | 62.8 | 10 | 2 | 43.1 |
| Organic carbon | 12 | 6 | 52.9 | 10 | 0 | 22.6 |
| Solids | 12 | 0 | 2.42 | 10 | 0 | 4.7 |
| Primary - Metals/Metalloids |  |  |  |  |  |  |
| Aluminum | 12 | 1 | 20.1 | 10 | 0 | 7.58 |
| Antimony | 12 | 1 | 22.2 | 10 | 0 | 12.9 |
| Arsenic | 12 | 0 | 12.5 | 10 | 0 | 15.3 |
| Barium | 12 | 2 | 26.4 | 10 | 0 | 22.3 |
| Beryllium | 12 | 0 | 17.9 | 10 | 0 | 9.56 |
| Cadmium | 12 | 0 | 19.5 | 10 | 0 | 14.4 |
| Calcium | 12 | 0 | 17.9 | 10 | 1 | 35.1 |
| Chromium | 12 | 1 | 28.8 | 10 | 0 | 26.6 |
| Cobalt | 12 | 0 | 16.1 | 10 | 0 | 12.3 |
| Copper | 12 | 0 | 8.6 | 10 | 0 | 14.4 |
| Iron | 12 | 0 | 9.69 | 10 | 0 | 8.94 |
| Lead | 12 | 1 | 22.7 | 10 | 0 | 14.8 |
| Magnesium | 12 | 1 | 25.6 | 10 | 0 | 16.8 |
| Manganese | 12 | 1 | 20.5 | 10 | 0 | 16.6 |
| Mercury | 12 | 0 | 13.3 | 10 | 0 | 34.6 |
| Molybdenum | 12 | 0 | 20 | 10 | 1 | 35.9 |
| Nickel | 12 | 0 | 18 | 10 | 0 | 23 |
| Potassium | 12 | 0 | 17.2 | 10 | 0 | 14.1 |
| Selenium | 12 | 0 | 14.3 | 10 | 0 | 21 |
| Silver | 12 | 0 | 15.9 | 10 | 0 | 15.1 |
| Sodium | 12 | 0 | 15.4 | 10 | 0 | 30.2 |
| Thallium | 12 | 1 | 23 | 10 | 0 | 10.2 |
| Vanadium | 12 | 0 | 10.5 | 10 | 0 | 11.7 |
| Zinc | 12 | 0 | 19.4 | 10 | 0 | 11.5 |

Notes:
Highlighted cells identify where relative percent differences (RPDs) and relative standard deviations (RSDs) are greater than the control limit.
Control limits specified in the quality assurance project plan (QAPP) (Exponent et al. 2014) are 20\% for analytical RPDs (i.e., metals, mercury, total organic carbon [TOC], cation exchange capacity [CEC], and pH ) and $35 \%$ for field triplicate RSDs. The QAPP did not specify a quality objective for grain size RPDs.

Table 5-7c. Aerial Deposition Area Summary of Field Split and Triplicate Sample Results for the $<149-\mu \mathrm{m}$ Soil Fraction

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | No. RPDs >20\% | Max RPD (\%) | Number of Samples | No. RSDs >35\% | Max RSD (\%) |
| High-density - Conventional Parameters |  |  |  |  |  |  |
| Solids | 6 | 0 | 1.82 | 6 | 0 | 1.64 |
| High-density - Metals/Metalloids |  |  |  |  |  |  |
| Aluminum | 6 | 0 | 2.69 | 6 | 0 | 11.3 |
| Antimony | 6 | 0 | 8.56 | 6 | 0 | 29.7 |
| Arsenic | 6 | 0 | 5.11 | 6 | 0 | 22.4 |
| Barium | 6 | 0 | 5.95 | 6 | 0 | 11.6 |
| Beryllium | 6 | 0 | 3.77 | 6 | 0 | 13.9 |
| Cadmium | 6 | 0 | 8.93 | 6 | 0 | 23.6 |
| Calcium | 6 | 0 | 9.92 | 6 | 0 | 15.3 |
| Chromium | 6 | 0 | 2.88 | 6 | 0 | 28.8 |
| Cobalt | 6 | 0 | 1.75 | 6 | 0 | 20.1 |
| Copper | 6 | 0 | 6.91 | 6 | 0 | 16.2 |
| Iron | 6 | 0 | 4.17 | 6 | 0 | 9.89 |
| Lead | 6 | 0 | 10.1 | 6 | 0 | 31.1 |
| Magnesium | 6 | 0 | 2.13 | 6 | 0 | 15.6 |
| Manganese | 6 | 0 | 6.76 | 6 | 0 | 12.9 |
| Mercury | 6 | 0 | 13.3 | 6 | 0 | 25 |
| Molybdenum | 6 | 0 | 7.55 | 6 | 0 | 8.2 |
| Nickel | 6 | 0 | 3.64 | 6 | 0 | 14.3 |
| Potassium | 6 | 0 | 3.69 | 6 | 0 | 21.2 |
| Selenium | 6 | 0 | 19.4 | 6 | 0 | 20 |
| Silver | 6 | 0 | 8.96 | 6 | 0 | 18.9 |
| Sodium | 6 | 0 | 8 | 6 | 0 | 33.7 |
| Thallium | 6 | 0 | 8 | 6 | 0 | 19.5 |
| Vanadium | 6 | 0 | 4.13 | 6 | 0 | 10.7 |
| Zinc | 6 | 0 | 7.28 | 6 | 0 | 17.6 |
| Primary - Conventional Parameters |  |  |  |  |  |  |
| Solids | 12 | 0 | 1.27 | 10 | 0 | 3.46 |
| Primary - Metals/Metalloids |  |  |  |  |  |  |
| Aluminum | 12 | 0 | 8.84 | 10 | 0 | 8.13 |
| Antimony | 12 | 0 | 15.3 | 10 | 0 | 12.8 |
| Arsenic | 12 | 0 | 16.8 | 10 | 0 | 19 |
| Barium | 12 | 1 | 24.7 | 10 | 0 | 26.1 |
| Beryllium | 12 | 0 | 4.17 | 10 | 0 | 11.3 |
| Cadmium | 12 | 0 | 16.7 | 10 | 0 | 16.1 |
| Calcium | 12 | 0 | 8.55 | 10 | 0 | 31.8 |
| Chromium | 12 | 0 | 9.4 | 10 | 0 | 21.9 |
| Cobalt | 12 | 0 | 6.08 | 10 | 0 | 10.5 |
| Copper | 12 | 0 | 12.9 | 10 | 0 | 13.2 |
| Iron | 12 | 0 | 6.06 | 10 | 0 | 6.57 |
| Lead | 12 | 0 | 18.3 | 10 | 0 | 13.9 |
| Magnesium | 12 | 0 | 7.2 | 10 | 0 | 15.5 |
| Manganese | 12 | 0 | 11.1 | 10 | 0 | 14 |
| Mercury | 12 | 2 | 22.2 | 10 | 0 | 33.3 |
| Molybdenum | 12 | 0 | 9.66 | 10 | 0 | 27 |
| Nickel | 12 | 0 | 6.15 | 10 | 0 | 20.2 |
| Potassium | 12 | 0 | 8.22 | 10 | 0 | 15.9 |
| Selenium | 12 | 0 | 16 | 10 | 0 | 27.2 |
| Silver | 12 | 1 | 22.2 | 10 | 0 | 21.7 |
| Sodium | 12 | 0 | 12.1 | 10 | 0 | 22.3 |
| Thallium | 12 | 0 | 13.3 | 10 | 0 | 10.2 |
| Vanadium | 12 | 0 | 10.7 | 10 | 0 | 9.36 |
| Zinc | 12 | 0 | 13.9 | 10 | 0 | 12.2 |

## Notes:

Highlighted cells identify where relative percent differences (RPDs) and relative standard deviations (RSDs) are greater than the control limit.
Control limits specified in the quality assurance project plan (QAPP) (Exponent et al. 2014) are 20\% for analytical RPDs (i.e., metals, mercury, total organic carbon [TOC], cation exchange capacity [CEC], and pH ) and $35 \%$ for field triplicate RSDs. The QAPP did not specify a quality objective for grain size RPDs.

Table 5-8a. Relict Floodplain Deposition Area Summary of Field Split and Triplicate Sample Results for Bulk Soil Samples

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | No. RPDs $>20 \%$ | Max RPD (\%) | Number of Samples | No. RSDs $>35 \%$ | Max RSD (\%) |
| RFA - Conventional Parameters |  |  |  |  |  |  |
| pH | 1 | 0 | 1.47 | 1 | 0 | 2.12 |
| Solids | 1 | 0 | 0.841 | 1 | 0 | 2.43 |
| Grain Size |  |  |  |  |  |  |
| Clay | 1 | na | 19.7 | 1 | 1 | 53.4 |
| Silt | 1 | na | 4.51 | 1 | 0 | 10 |
| Very fine sand | 1 | na | 16.2 | 1 | 0 | 9.32 |
| Fine sand | 1 | na | 5.71 | 1 | 0 | 14.6 |
| Medium sand | 1 | na | 9.56 | 1 | 0 | 7.17 |
| Coarse sand | 1 | na | 5.67 | 1 | 1 | 36.9 |
| Very coarse sand | 1 | na | 1.57 | 1 | 0 | 29.2 |
| Very fine gravel | 1 | na | 21.9 | 1 | 0 | 19.8 |
| Fine gravel | 1 | na | 198 | 1 | 1 | 57.7 |
| Medium gravel | 1 | na | 200 | 1 | 1 | 98.4 |
| Coarse gravel | 1 | na | 0 | 1 | 1 | 173 |
| Very coarse gravel | 1 | na | 0 | 1 | 0 | 0 |
| Cobbles | 1 | na | 0 | 1 | 0 | 0 |
| RFB - Conventional Parameters |  |  |  |  |  |  |
| pH | ns | ns | ns | 1 | 0 | 0.285 |
| Solids | ns | ns | ns | 1 | 0 | 4.49 |
| Grain Size |  |  |  |  |  |  |
| Clay | ns | ns | ns | 1 | 0 | 24.7 |
| Silt | ns | ns | ns | 1 | 0 | 18 |
| Very fine sand | ns | ns | ns | 1 | 0 | 5.48 |
| Fine sand | ns | ns | ns | 1 | 1 | 44 |
| Medium sand | ns | ns | ns | 1 | 1 | 71.9 |
| Coarse sand | ns | ns | ns | 1 | 1 | 91.5 |
| Very coarse sand | ns | ns | ns | 1 | 1 | 74.2 |
| Very fine gravel | ns | ns | ns | 1 | 0 | 22.9 |
| Fine gravel | ns | ns | ns | 1 | 1 | 63.7 |
| Medium gravel | ns | ns | ns | 1 | 1 | 125 |
| Coarse gravel | ns | ns | ns | 1 | 0 | 0 |
| Very coarse gravel | ns | ns | ns | 1 | 0 | 0 |
| Cobbles | ns | ns | ns | 1 | 0 | 0 |
| RFC - Conventional Parameters |  |  |  |  |  |  |
| pH | 1 | 0 | 0.638 | 1 | 0 | 1.44 |
| Solids | 1 | 0 | 0.339 | 1 | 0 | 4.85 |
| Grain Size |  |  |  |  |  |  |
| Clay | 1 | na | 10.1 | 1 | 0 | 15.8 |
| Silt | 1 | na | 5.71 | 1 | 0 | 13.4 |
| Very fine sand | 1 | na | 10.9 | 1 | 0 | 12.9 |
| Fine sand | 1 | na | 8.34 | 1 | 0 | 28.3 |
| Medium sand | 1 | na | 2.53 | 1 | 1 | 102 |
| Coarse sand | 1 | na | 1.1 | 1 | 1 | 93.5 |
| Very coarse sand | 1 | na | 10.7 | 1 | 1 | 83.5 |
| Very fine gravel | 1 | na | 85.7 | 1 | 1 | 75.2 |
| Fine gravel | 1 | na | 96.3 | 1 | 1 | 163 |
| Medium gravel | 1 | na | 0 | 1 | 1 | 173 |
| Coarse gravel | 1 | na | 0 | 1 | 0 | 0 |
| Very coarse gravel | 1 | na | 0 | 1 | 0 | 0 |
| Cobbles | 1 | na | 0 | 1 | 0 | 0 |
| RFD - Conventional Parameters |  |  |  |  |  |  |
| pH | ns | ns | ns | 1 | 0 | 8.29 |
| Solids | ns | ns | ns | 1 | 0 | 3.17 |
| Grain Size |  |  |  |  |  |  |
| Clay | ns | ns | ns | 1 | 0 | 22.8 |
| Silt | ns | ns | ns | 1 | 1 | 35.8 |
| Very fine sand | ns | ns | ns | 1 | 0 | 34.4 |
| Fine sand | ns | ns | ns | 1 | 0 | 9.82 |
| Medium sand | ns | ns | ns | 1 | 1 | 48.7 |
| Coarse sand | ns | ns | ns | 1 | 1 | 50 |
| Very coarse sand | ns | ns | ns | 1 | 1 | 42.3 |
| Very fine gravel | ns | ns | ns | 1 | 1 | 68.9 |
| Fine gravel | ns | ns | ns | 1 | 1 | 37 |
| Medium gravel | ns | ns | ns | 1 | 1 | 173 |
| Coarse gravel | ns | ns | ns | 1 | 0 | 0 |
| Very coarse gravel | ns | ns | ns | 1 | 0 | 0 |
| Cobbles | ns | ns | ns | 1 | 0 | 0 |

## Notes:

Highlighted cells identify where relative percent differences (RPDs) and relative standard deviations (RSDs) are greater than the control limit.
Control limits specified in the quality assurance project plan (QAPP) (Exponent et al. 2014) are $20 \%$ for analytical RPDs (i.e., metals, mercury, total organic carbon [TOC], cation exchange capacity [CEC], and pH ) and $35 \%$ for field triplicate RSDs. The QAPP did not specify a quality objective for grain size RPDs. na - not applicable
ns - not sampled (field duplicates were not prepared)
RFA, RFB, RFC, and RFD - relict flood plain depositional areas A, B, C, and D

Table 5-8b. Relict Floodplain Deposition Area Summary of Field Split and Triplicate Sample Results for the < 2-mm Soil Fraction

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | No. RPDs $>20 \%$ | Max RPD (\%) | Number of Samples | No. RSDs $>35 \%$ | Max RSD (\%) |
| RFA - Conventional Parameters |  |  |  |  |  |  |
| CEC | 1 | 0 | 13.9 | 1 | 0 | 14 |
| Organic carbon | 1 | 0 | 5.88 | 1 | 0 | 14.6 |
| Solids | 1 | 0 | 0 | 1 | 0 | 0.823 |
| RFA - Metals/Metalloids |  |  |  |  |  |  |
| Aluminum | 1 | 0 | 11.3 | 1 | 0 | 8.1 |
| Antimony | 1 | 0 | 0 | 1 | 0 | 5.21 |
| Arsenic | 1 | 0 | 2.75 | 1 | 0 | 6.57 |
| Barium | 1 | 0 | 4.39 | 1 | 0 | 1.37 |
| Beryllium | 1 | 0 | 13.7 | 1 | 0 | 7.56 |
| Cadmium | 1 | 0 | 7.55 | 1 | 0 | 7.73 |
| Calcium | 1 | 0 | 3.17 | 1 | 0 | 5.41 |
| Chromium | 1 | 0 | 12.4 | 1 | 0 | 4.78 |
| Cobalt | 1 | 0 | 8.7 | 1 | 0 | 10.3 |
| Copper | 1 | 0 | 16.4 | 1 | 0 | 12.8 |
| Iron | 1 | 0 | 14.8 | 1 | 0 | 15 |
| Lead | 1 | 0 | 0.704 | 1 | 0 | 8.47 |
| Magnesium | 1 | 0 | 8.39 | 1 | 0 | 8.49 |
| Manganese | 1 | 0 | 13.3 | 1 | 0 | 13.3 |
| Mercury | 1 | 1 | 52.6 | 1 | 1 | 46.9 |
| Molybdenum | 1 | 0 | 8.97 | 1 | 0 | 6.13 |
| Nickel | 1 | 0 | 1.74 | 1 | 0 | 5.85 |
| Potassium | 1 | 0 | 10.5 | 1 | 0 | 9.71 |
| Selenium | 1 | 0 | 15.8 | 1 | 0 | 4.61 |
| Silver | 1 | 0 | 16.2 | 1 | 0 | 11.6 |
| Sodium | 1 | 0 | 15.7 | 1 | 0 | 18.7 |
| Thallium | 1 | 0 | 19.4 | 1 | 0 | 10.6 |
| Vanadium | 1 | 0 | 6.9 | 1 | 0 | 3.16 |
| Zinc | 1 | 0 | 15.5 | 1 | 0 | 15.7 |
| RFB - Conventional Parameters |  |  |  |  |  |  |
| CEC | ns | ns | ns | 1 | 0 | 26.3 |
| Organic carbon | ns | ns | ns | 1 | 0 | 4.67 |
| Solids | ns | ns | ns | 1 | 0 | 1.57 |
| RFB - Metals/Metalloids |  |  |  |  |  |  |
| Aluminum | ns | ns | ns | 1 | 0 | 6.01 |
| Antimony | ns | ns | ns | 1 | 1 | 45.5 |
| Arsenic | ns | ns | ns | 1 | 0 | 19.2 |
| Barium | ns | ns | ns | 1 | 0 | 10.7 |
| Beryllium | ns | ns | ns | 1 | 0 | 15.3 |
| Cadmium | ns | ns | ns | 1 | 1 | 39.5 |
| Calcium | ns | ns | ns | 1 | 0 | 8.65 |
| Chromium | ns | ns | ns | 1 | 0 | 7.19 |
| Cobalt | ns | ns | ns | 1 | 0 | 5 |
| Copper | ns | ns | ns | 1 | 0 | 13.9 |
| Iron | ns | ns | ns | 1 | 0 | 5.84 |
| Lead | ns | ns | ns | 1 | 1 | 45.9 |
| Magnesium | ns | ns | ns | 1 | 0 | 3.72 |
| Manganese | ns | ns | ns | 1 | 0 | 22.5 |
| Mercury | ns | ns | ns | 1 | 0 | 24.7 |
| Molybdenum | ns | ns | ns | 1 | 0 | 8.97 |
| Nickel | ns | ns | ns | 1 | 0 | 7.85 |
| Potassium | ns | ns | ns | 1 | 0 | 4.21 |
| Selenium | ns | ns | ns | 1 | 0 | 24.6 |
| Silver | ns | ns | ns | 1 | 0 | 25.9 |
| Sodium | ns | ns | ns | 1 | 0 | 4.26 |
| Thallium | ns | ns | ns | 1 | 0 | 20.3 |
| Vanadium | ns | ns | ns | 1 | 0 | 4.93 |
| Zinc | ns | ns | ns | 1 | 0 | 21.2 |
| RFC - Conventional Parameters |  |  |  |  |  |  |
| CEC | 1 | 0 | 0.509 | 1 | 0 | 14.5 |
| Organic carbon | 1 | 0 | 2.76 | 1 | 0 | 2.14 |
| Solids | 1 | 0 | 0 | 1 | 0 | 0.588 |
| RFC - Metals/Metalloids |  |  |  |  |  |  |
| Aluminum | 1 | 0 | 0.502 | 1 | 0 | 7.51 |
| Antimony | 1 | 0 | 0.574 | 1 | 0 | 10.4 |
| Arsenic | 1 | 0 | 0 | 1 | 0 | 2.65 |
| Barium | 1 | 0 | 2.36 | 1 | 0 | 5.91 |
| Beryllium | 1 | 0 | 0 | 1 | 0 | 9.6 |
| Cadmium | 1 | 0 | 0.702 | 1 | 0 | 2.79 |

Table 5-8b. Relict Floodplain Deposition Area Summary of Field Split and Triplicate Sample Results for the < 2-mm Soil Fraction

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | No. RPDs >20\% | Max RPD (\%) | Number of Samples | No. RSDs > 35\% | Max RSD (\%) |
| RFC - Metals/Metalloids (continued) |  |  |  |  |  |  |
| Calcium | 1 | 0 | 5.82 | 1 | 0 | 1.5 |
| Chromium | 1 | 0 | 1.84 | 1 | 0 | 4.18 |
| Cobalt | 1 | 0 | 2.45 | 1 | 0 | 2.2 |
| Copper | 1 | 0 | 4.44 | 1 | 0 | 4.95 |
| Iron | 1 | 0 | 1.07 | 1 | 0 | 4.09 |
| Lead | 1 | 0 | 4.8 | 1 | 0 | 4.8 |
| Magnesium | 1 | 0 | 1.05 | 1 | 0 | 4.11 |
| Manganese | 1 | 0 | 0.448 | 1 | 0 | 5.46 |
| Mercury | 1 | 0 | 13.3 | 1 | 0 | 5.33 |
| Molybdenum | 1 | 0 | 1.71 | 1 | 0 | 0.813 |
| Nickel | 1 | 0 | 2.61 | 1 | 0 | 3.17 |
| Potassium | 1 | 0 | 1.86 | 1 | 0 | 7.35 |
| Selenium | 1 | 0 | 3.97 | 1 | 0 | 3.7 |
| Silver | 1 | 0 | 1.87 | 1 | 0 | 4.91 |
| Sodium | 1 | 0 | 16.7 | 1 | 0 | 31.1 |
| Thallium | 1 | 0 | 5.71 | 1 | 0 | 9.74 |
| Vanadium | 1 | 0 | 0.93 | 1 | 0 | 3.37 |
| Zinc | 1 | 0 | 0.487 | 1 | 0 | 1.29 |
| RFD - Conventional Parameters |  |  |  |  |  |  |
| CEC | ns | ns | ns | 1 | 0 | 2.88 |
| Organic carbon | ns | ns | ns | 1 | 0 | 21.5 |
| RFD - Metals/Metalloids |  |  |  |  |  |  |
| Solids | ns | ns | ns | 1 | 0 | 0.354 |
| Aluminum | ns | ns | ns | 1 | 0 | 6.05 |
| Antimony | ns | ns | ns | 1 | 0 | 7.11 |
| Arsenic | ns | ns | ns | 1 | 0 | 13.2 |
| Barium | ns | ns | ns | 1 | 1 | 37.6 |
| Beryllium | ns | ns | ns | 1 | 0 | 5.8 |
| Cadmium | ns | ns | ns | 1 | 0 | 6.69 |
| Calcium | ns | ns | ns | 1 | 1 | 37.6 |
| Chromium | ns | ns | ns | 1 | 0 | 3.05 |
| Cobalt | ns | ns | ns | 1 | 0 | 4.79 |
| Copper | ns | ns | ns | 1 | 0 | 8.62 |
| Iron | ns | ns | ns | 1 | 0 | 5.36 |
| Lead | ns | ns | ns | 1 | 0 | 1.79 |
| Magnesium | ns | ns | ns | 1 | 0 | 28.1 |
| Manganese | ns | ns | ns | 1 | 0 | 6.08 |
| Mercury | ns | ns | ns | 1 | 0 | 31.4 |
| Molybdenum | ns | ns | ns | 1 | 0 | 30.7 |
| Nickel | ns | ns | ns | 1 | 0 | 6 |
| Potassium | ns | ns | ns | 1 | 0 | 5.11 |
| Selenium | ns | ns | ns | 1 | 0 | 12.3 |
| Silver | ns | ns | ns | 1 | 0 | 20.4 |
| Sodium | ns | ns | ns | 1 | 0 | 3.82 |
| Thallium | ns | ns | ns | 1 | 0 | 5.15 |
| Vanadium | ns | ns | ns | 1 | 0 | 4.01 |
| Zinc | ns | ns | ns | 1 | 0 | 19.8 |

## Notes:

Highlighted cells identify where relative percent differences (RPDs) and relative standard deviations (RSDs) are greater than the control limit.
Control limits specified in the quality assurance project plan (QAPP) (Exponent et al. 2014) are $20 \%$ for analytical RPDs (i.e., metals, mercury, total organic carbon [TOC], cation exchange capacity [CEC], and pH ) and $35 \%$ for field triplicate RSDs. The QAPP did not specify a quality objective for grain size RPDs.
ns - not sampled (field duplicates were not prepared)
RFA, RFB, RFC, and RFD - relict flood plain depositional areas A, B, C, and D

Table 5-8c. Relict Floodplain Deposition Area Summary of Field Split and Triplicate Sample Results for the < $149-\mu \mathrm{m}$ Soil Fraction

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | No. RPDs > $20 \%$ | Max RPD (\%) | Number of Samples | No. RSDs $>35 \%$ | Max RSD (\%) |
| RFA - Conventional Parameters |  |  |  |  |  |  |
| Solids | 1 | 0 | 0.502 | 1 | 0 | 0.637 |
| RFA - Metals/Metalliods |  |  |  |  |  |  |
| Aluminum | 1 | 0 | 5.27 | 1 | 0 | 6.63 |
| Antimony | 1 | 0 | 3.9 | 1 | 0 | 16.1 |
| Arsenic | 1 | 0 | 2.71 | 1 | 0 | 3.45 |
| Barium | 1 | 0 | 11 | 1 | 0 | 6.93 |
| Beryllium | 1 | 0 | 3.92 | 1 | 0 | 7.14 |
| Cadmium | 1 | 0 | 4.76 | 1 | 0 | 2.29 |
| Calcium | 1 | 0 | 0.603 | 1 | 0 | 6.2 |
| Chromium | 1 | 0 | 7.59 | 1 | 0 | 5.43 |
| Cobalt | 1 | 0 | 2.43 | 1 | 0 | 3.73 |
| Copper | 1 | 0 | 4.23 | 1 | 0 | 7.56 |
| Iron | 1 | 0 | 3.31 | 1 | 0 | 1.21 |
| Lead | 1 | 0 | 4.96 | 1 | 0 | 4.41 |
| Magnesium | 1 | 0 | 0.76 | 1 | 0 | 4.7 |
| Manganese | 1 | 0 | 3.34 | 1 | 0 | 1.96 |
| Mercury | 1 | 0 | 20 | 1 | 0 | 31.5 |
| Molybdenum | 1 | 0 | 8.35 | 1 | 0 | 12.7 |
| Nickel | 1 | 0 | 5.3 | 1 | 0 | 1.85 |
| Potassium | 1 | 0 | 4.2 | 1 | 0 | 6.3 |
| Selenium | 1 | 0 | 1.98 | 1 | 0 | 7.86 |
| Silver | 1 | 1 | 21.4 | 1 | 0 | 11.6 |
| Sodium | 1 | 0 | 2.74 | 1 | 0 | 4.09 |
| Thallium | 1 | 0 | 10.9 | 1 | 0 | 8.11 |
| Vanadium | 1 | 0 | 7.16 | 1 | 0 | 1.84 |
| Zinc | 1 | 0 | 6.37 | 1 | 0 | 1.18 |
| RFB - Conventional Parameters |  |  |  |  |  |  |
| Solids | ns | ns | ns | 1 | 0 | 0.446 |
| RFB - Metals/Metalliods |  |  |  |  |  |  |
| Aluminum | ns | ns | ns | 1 | 0 | 3.65 |
| Antimony | ns | ns | ns | 1 | 1 | 41.1 |
| Arsenic | ns | ns | ns | 1 | 0 | 17.7 |
| Barium | ns | ns | ns | 1 | 0 | 4.61 |
| Beryllium | ns | ns | ns | 1 | 0 | 4.44 |
| Cadmium | ns | ns | ns | 1 | 0 | 33.1 |
| Calcium | ns | ns | ns | 1 | 0 | 2.6 |
| Chromium | ns | ns | ns | 1 | 0 | 3.11 |
| Cobalt | ns | ns | ns | 1 | 0 | 4.03 |
| Copper | ns | ns | ns | 1 | 0 | 8.44 |
| Iron | ns | ns | ns | 1 | 0 | 2.5 |
| Lead | ns | ns | ns | 1 | 1 | 40.6 |
| Magnesium | ns | ns | ns | 1 | 0 | 5.43 |
| Manganese | ns | ns | ns | 1 | 0 | 12.8 |
| Mercury | ns | ns | ns | 1 | 0 | 31.5 |
| Molybdenum | ns | ns | ns | 1 | 0 | 6.72 |
| Nickel | ns | ns | ns | 1 | 0 | 2.66 |
| Potassium | ns | ns | ns | 1 | 0 | 7.24 |
| Selenium | ns | ns | ns | 1 | 0 | 7.26 |
| Silver | ns | ns | ns | 1 | 0 | 20.5 |
| Sodium | ns | ns | ns | 1 | 0 | 11.5 |
| Thallium | ns | ns | ns | 1 | 0 | 15.1 |
| Vanadium | ns | ns | ns | 1 | 0 | 4 |
| Zinc | ns | ns | ns | 1 | 0 | 14.2 |
| RFC - Conventional Parameters |  |  |  |  |  |  |
| Solids | 1 | 0 | 0.305 | 1 | 0 | 0.389 |

Table 5-8c. Relict Floodplain Deposition Area Summary of Field Split and Triplicate Sample Results for the < 149- $\mu \mathrm{m}$ Soil Fraction

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | No. RPDs $>20 \%$ | Max RPD (\%) | Number of Samples | No. RSDs $>35 \%$ | Max RSD (\%) |
| RFC-Metals/Metalliods |  |  |  |  |  |  |
| Aluminum | 1 | 0 | 0.456 | 1 | 0 | 3.3 |
| Antimony | 1 | 0 | 1.85 | 1 | 0 | 12.6 |
| Arsenic | 1 | 0 | 1.67 | 1 | 0 | 5.84 |
| Barium | 1 | 0 | 2.09 | 1 | 0 | 5.24 |
| Beryllium | 1 | 0 | 2.67 | 1 | 0 | 12.1 |
| Cadmium | 1 | 0 | 3.18 | 1 | 0 | 4.58 |
| Calcium | 1 | 0 | 1.68 | 1 | 0 | 1.4 |
| Chromium | 1 | 0 | 1.93 | 1 | 0 | 1.86 |
| Cobalt | 1 | 0 | 3.1 | 1 | 0 | 1.19 |
| Copper | 1 | 0 | 1.28 | 1 | 0 | 9.5 |
| Iron | 1 | 0 | 1.78 | 1 | 0 | 2.1 |
| Lead | 1 | 0 | 4.31 | 1 | 0 | 6.44 |
| Magnesium | 1 | 0 | 2.37 | 1 | 0 | 2.33 |
| Manganese | 1 | 0 | 2.15 | 1 | 0 | 3.87 |
| Mercury | 1 | 0 | 5.71 | 1 | 0 | 7.39 |
| Molybdenum | 1 | 0 | 4.82 | 1 | 0 | 5.04 |
| Nickel | 1 | 0 | 1.81 | 1 | 0 | 1.59 |
| Potassium | 1 | 0 | 3.19 | 1 | 0 | 10.2 |
| Selenium | 1 | 0 | 8.89 | 1 | 0 | 4.69 |
| Silver | 1 | 0 | 4.72 | 1 | 0 | 10.1 |
| Sodium | 1 | 0 | 5.67 | 1 | 1 | 37.2 |
| Thallium | 1 | 0 | 3.17 | 1 | 0 | 12.4 |
| Vanadium | 1 | 0 | 3.33 | 1 | 0 | 2.11 |
| Zinc | 1 | 0 | 0.469 | 1 | 0 | 0.422 |
| RFD - Conventional Parameters |  |  |  |  |  |  |
| Solids | ns | ns | ns | 1 | 0 | 0.101 |
| RFD-Metals/Metaliods |  |  |  |  |  |  |
| Aluminum | ns | ns | ns | 1 | 0 | 9.87 |
| Antimony | ns | ns | ns | 1 | 0 | 17.4 |
| Arsenic | ns | ns | ns | 1 | 0 | 16.8 |
| Barium | ns | ns | ns | 1 | 0 | 31.7 |
| Beryllium | ns | ns | ns | 1 | 0 | 4.88 |
| Cadmium | ns | ns | ns | 1 | 0 | 3.98 |
| Calcium | ns | ns | ns | 1 | 1 | 40 |
| Chromium | ns | ns | ns | 1 | 0 | 4.59 |
| Cobalt | ns | ns | ns | 1 | 0 | 5.22 |
| Copper | ns | ns | ns | 1 | 0 | 7.62 |
| Iron | ns | ns | ns | 1 | 0 | 4.22 |
| Lead | ns | ns | ns | 1 | 0 | 9.44 |
| Magnesium | ns | ns | ns | 1 | 0 | 32.3 |
| Manganese | ns | ns | ns | 1 | 0 | 6.44 |
| Mercury | ns | ns | ns | 1 | 1 | 36.2 |
| Molybdenum | ns | ns | ns | 1 | 0 | 24.6 |
| Nickel | ns | ns | ns | 1 | 0 | 2.94 |
| Potassium | ns | ns | ns | 1 | 0 | 13 |
| Selenium | ns | ns | ns | 1 | 0 | 10.9 |
| Silver | ns | ns | ns | 1 | 0 | 13.2 |
| Sodium | ns | ns | ns | 1 | 0 | 6.17 |
| Thallium | ns | ns | ns | 1 | 0 | 4.17 |
| Vanadium | ns | ns | ns | 1 | 0 | 1.07 |
| Zinc | ns | ns | ns | 1 | 0 | 19.5 |

## Notes:

Highlighted cells identify where relative percent differences (RPDs) and relative standard deviations (RSDs) are greater than the control limit.
Control limits specified in the quality assurance project plan (QAPP) (Exponent et al. 2014) are 20\% for analytical RPDs (i.e., metals, mercury, total organic carbon [TOC], cation exchange capacity [CEC], and pH ) and $35 \%$ for field triplicate RSDs. The QAPP did not specify a quality objective for grain size RPDs.
ns - not sampled (field duplicates were not prepared)
RFA, RFB, RFC, and RFD - relict flood plain depositional areas A, B, C, and D

Table 5-9a. Windblown Sediment Deposition Area Summary of Field Split and Triplicate Sample Results for Bulk Soil Samples

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | No. RPDs > $20 \%$ | Max RPD (\%) | Number of Samples | No. RSDs >35\% | Max RSD (\%) |
| Columbia Beach North - Conventional Parameters |  |  |  |  |  |  |
| pH | 1 | 0 | 3.62 | 1 | 0 | 1.07 |
| Solids | 1 | 0 | 0.101 | 1 | 0 | 0.0587 |
| Grain Size |  |  |  |  |  |  |
| Clay | 1 | na | 0.425 | 1 | 0 | 3.87 |
| Silt | 1 | na | 11.5 | 1 | 0 | 12.3 |
| Very fine sand | 1 | na | 4.27 | 1 | 0 | 3.59 |
| Fine sand | 1 | na | 6.22 | 1 | 0 | 6.28 |
| Medium sand | 1 | na | 8.1 | 1 | 0 | 10.4 |
| Coarse sand | 1 | na | 2.73 | 1 | 0 | 8.19 |
| Very coarse sand | 1 | na | 9.64 | 1 | 0 | 14.8 |
| Very fine gravel | 1 | na | 9.32 | 1 | 0 | 24.6 |
| Fine gravel | 1 | na | 60.1 | 1 | 1 | 83 |
| Medium gravel | 1 | na | 0 | 1 | 1 | 173 |
| Coarse gravel | 1 | na | 0 | 1 | 0 | 0 |
| Very coarse gravel | 1 | na | 0 | 1 | 0 | 0 |
| Cobbles | 1 | na | 0 | 1 | 0 | 0 |
| Marcus Flats East - Conventional Parameters |  |  |  |  |  |  |
| pH | 1 | 0 | 2.44 | 1 | 0 | 2.66 |
| Solids | 1 | 0 | 1.41 | 1 | 0 | 1.15 |
| Grain Size |  |  |  |  |  |  |
| Clay | 1 | na | 8.22 | 1 | 1 | 40.3 |
| Silt | 1 | na | 6.35 | 1 | 0 | 8.48 |
| Very fine sand | 1 | na | 0.477 | 1 | 0 | 4.9 |
| Fine sand | 1 | na | 1.06 | 1 | 0 | 10.4 |
| Medium sand | 1 | na | 8.59 | 1 | 0 | 9.03 |
| Coarse sand | 1 | na | 23.1 | 1 | 0 | 34 |
| Very coarse sand | 1 | na | 25.6 | 1 | 0 | 11.6 |
| Very fine gravel | 1 | na | 102 | 1 | 1 | 54.1 |
| Fine gravel | 1 | na | 28.9 | 1 | 1 | 51.3 |
| Medium gravel | 1 | na | 200 | 1 | 1 | 173 |
| Coarse gravel | 1 | na | 0 | 1 | 0 | 0 |
| Very coarse gravel | 1 | na | 0 | 1 | 0 | 0 |
| Cobbles | 1 | na | 0 | 1 | 0 | 0 |

## Notes

Highlighted cells identify where relative percent differences (RPDs) and relative standard deviations (RSDs) are greater than the control limit.
Control limits specified in the quality assurance project plan (QAPP) (Exponent et al. 2014) are 20\% for analytical RPDs (i.e., metals, mercury, total organic carbon [TOC], cation exchange capacity [CEC], and pH ) and $35 \%$ for field triplicate RSDs. The QAPP did not specify a quality objective for grain size RPDs.
na - not applicable

Table 5-9b. Windblown Sediment Deposition Area Summary of Field Split and Triplicate Sample Results for the < 2-mm Soil Fraction

| Analyte | Field Split Sample RPDs |  |  | Triplicate Sample RSDs |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of Samples | No. RPDs > $20 \%$ | Max RPD (\%) | Number of Samples | No. RSDs > $35 \%$ | Max RSD (\%) |
| Columbia Beach North - Conventional Parameters |  |  |  |  |  |  |
| CEC | 1 | 0 | 18.8 | 1 | 0 | 14 |
| Organic carbon | 1 | 0 | 1.71 | 1 | 0 | 13.8 |
| Solids | 1 | 0 | 0.1 | 1 | 0 | 0.0582 |
| Columbia Beach North - Metals/Metalloids |  |  |  |  |  |  |
| Aluminum | 1 | 0 | 2.76 | 1 | 0 | 1.79 |
| Antimony | 1 | 0 | 9.52 | 1 | 0 | 5.33 |
| Arsenic | 1 | 1 | 24.2 | 1 | 0 | 8.02 |
| Barium | 1 | 0 | 1.64 | 1 | 0 | 1.84 |
| Beryllium | 1 | 0 | 0 | 1 | 0 | 4.26 |
| Cadmium | 1 | 0 | 4.88 | 1 | 0 | 2.47 |
| Calcium | 1 | 0 | 11.3 | 1 | 0 | 2.8 |
| Chromium | 1 | 0 | 1.34 | 1 | 0 | 4.5 |
| Cobalt | 1 | 0 | 4.87 | 1 | 0 | 1.83 |
| Copper | 1 | 0 | 0.778 | 1 | 0 | 2.34 |
| Iron | 1 | 0 | 9.58 | 1 | 0 | 2.19 |
| Lead | 1 | 0 | 0.211 | 1 | 0 | 2.32 |
| Magnesium | 1 | 0 | 5.01 | 1 | 0 | 2.8 |
| Manganese | 1 | 0 | 4.01 | 1 | 0 | 2.7 |
| Mercury | 1 | 0 | 11.8 | 1 | 0 | 11.1 |
| Molybdenum | 1 | 0 | 4.35 | 1 | 0 | 6.35 |
| Nickel | 1 | 0 | 4.69 | 1 | 0 | 3.95 |
| Potassium | 1 | 0 | 4.23 | 1 | 0 | 4.38 |
| Selenium | 1 | 0 | 11.8 | 1 | 0 | 6.93 |
| Silver | 1 | 1 | 28.6 | 1 | 0 | 12.4 |
| Sodium | 1 | 1 | 27.2 | 1 | 0 | 22.5 |
| Thallium | 1 | 0 | 0 | 1 | 0 | 0 |
| Vanadium | 1 | 0 | 1.59 | 1 | 0 | 1.1 |
| Zinc | 1 | 0 | 5.32 | 1 | 0 | 1.33 |
| Marcus Flats East - Conventional Parameters |  |  |  |  |  |  |
| CEC | 1 | 1 | 30.4 | 1 | 0 | 33.9 |
| Organic carbon | 1 | 0 | 10.8 | 1 | 0 | 28 |
| Solids | 1 | 0 | 0.105 | 1 | 0 | 0.26 |
| Marcus Flats East - Metals/Metalloids |  |  |  |  |  |  |
| Aluminum | 1 | 0 | 0.627 | 1 | 0 | 2.25 |
| Antimony | 1 | 0 | 7.63 | 1 | 0 | 13 |
| Arsenic | 1 | 0 | 5.78 | 1 | 0 | 12.1 |
| Barium | 1 | 0 | 2.68 | 1 | 0 | 4.11 |
| Beryllium | 1 | 0 | 0 | 1 | 0 | 3.94 |
| Cadmium | 1 | 0 | 5.99 | 1 | 0 | 20.2 |
| Calcium | 1 | 0 | 2.6 | 1 | 0 | 1.95 |
| Chromium | 1 | 0 | 7.69 | 1 | 0 | 3.38 |
| Cobalt | 1 | 0 | 4.47 | 1 | 0 | 1.11 |
| Copper | 1 | 0 | 6.16 | 1 | 0 | 4.09 |
| Iron | 1 | 0 | 0.702 | 1 | 0 | 3.05 |
| Lead | 1 | 0 | 7.22 | 1 | 0 | 23.4 |
| Magnesium | 1 | 0 | 1.79 | 1 | 0 | 3.62 |
| Manganese | 1 | 0 | 2.26 | 1 | 0 | 3.46 |
| Mercury | 1 | 0 | 9.52 | 1 | 0 | 10.8 |
| Molybdenum | 1 |  | 25.5 | 1 | 0 | 4.68 |
| Nickel | 1 | 0 | 1.75 | 1 | 0 | 1.14 |
| Potassium | 1 | 0 | 0 | 1 | 0 | 3.86 |
| Selenium | 1 | 1 | 21.4 | 1 | 0 | 6.54 |
| Silver | 1 | 0 | 6.9 | 1 | 0 | 0.00000134 |
| Sodium | 1 | 0 | 15.1 | 1 | 0 | 6.57 |
| Thallium | 1 | 0 | 0 | 1 | 0 | 9.52 |
| Vanadium | 1 | 0 | 1.16 | 1 | 0 | 5.59 |
| Zinc | 1 | 0 | 1.94 | 1 | 0 | 12.8 |

## Notes:

Highlighted cells identify where relative percent differences (RPDs) and relative standard deviations (RSDs) are greater than the control limit.
Control limits specified in the quality assurance project plan (QAPP) (Exponent et al. 2014) are 20\% for analytical RPDs (i.e., metals, mercury, total organic carbon [TOC], cation exchange capacity [CEC], and pH ) and $35 \%$ for field triplicate RSDs. The QAPP did not specify a quality objective for grain size RPDs.

Table 5-10. Comparison of Actual Method Reporting Limits with Analytical Concentration Goals for Nondetected Samples

| Analyte | Soil Fraction | ACG | MRL | Minimum MRL | Maximum MRL | Units | Number of 1X ACG Exceedances / Total Nondetected Results | Number of 10X ACG Exceedances / Total Nondetected Results |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ADA - High-density |  |  |  |  |  |  |  |  |
| Sodium | < 149- $\mu \mathrm{m}$ | 40 | 40 | 106 | 124 | mg/kg | $2 / 2$ | $0 / 2$ |
| Sodium | <2-mm | 40 | 40 | 91.5 | 125 | mg/kg | $3 / 3$ | $0 / 3$ |
| ADA - Primary |  |  |  |  |  |  |  |  |
| Sodium | < 149- $\mu \mathrm{m}$ | 40 | 40 | 97.9 | 127 | mg/kg | $9 / 9$ | $0 / 9$ |
| Sodium | <2-mm | 40 | 40 | 51.4 | 125 | mg/kg | $13 / 13$ | $0 / 13$ |
| WSDA - Columbia Beach South |  |  |  |  |  |  |  |  |
| Selenium | < 2-mm | 0.3 | 0.2 | 0.19 | 0.19 | mg/kg | $0 / 1$ | $0 / 1$ |
| Notes: |  |  |  |  |  |  |  |  |
| ACG - analytical concentration goal |  |  |  |  |  |  |  |  |
| ADA - Aerial deposition area |  |  |  |  |  |  |  |  |
| MRL - method reporting limit |  |  |  |  |  |  |  |  |
| WSDA - windb | sediment depo | ition ar |  |  |  |  |  |  |

Table 5-11a. Summary of Metals Data Compared with Available Eco-SSLs

| Analyte | Soil Fraction | $\begin{aligned} & \text { Eco-SSL } \\ & (\mathrm{mg} / \mathrm{kg})^{\mathrm{a}} \end{aligned}$ | ADA |  | RFDA |  | WSDA |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Number of DUs | Number of DUs > Eco-SSL | Number of DUs | Number of DUs > Eco-SSL | Number of DUs | Number of DUs > Eco-SSL |
| Antimony | < 2-mm | 0.27 | 142 | 142 | 16 | 16 | 13 | 8 |
| Arsenic | <2-mm | 18 | 142 | 41 | 16 | 5 | 13 | 0 |
| Barium | < 2-mm | 330 | 142 | 59 | 16 | 5 | 13 | 0 |
| Beryllium | < 2-mm | 21 | 142 | 0 | 16 | 0 | 13 | 0 |
| Cadmium | < 2-mm | 0.36 | 142 | 142 | 16 | 16 | 13 | 7 |
| Chromium | <2-mm | 26 | 142 | 26 | 16 | 9 | 13 | 0 |
| Cobalt | <2-mm | 13 | 142 | 5 | 16 | 5 | 13 | 0 |
| Copper | <2-mm | 28 | 142 | 21 | 16 | 13 | 13 | 0 |
| Lead | <2-mm | 11 | 142 | 142 | 16 | 16 | 13 | 8 |
| Manganese ${ }^{\text {b }}$ | <2-mm | 220 | 142 | 141 | 16 | 16 | 13 | 13 |
| Nickel | <2-mm | 38 | 142 | 10 | 16 | 0 | 13 | 0 |
| Selenium | <2-mm | 0.52 | 142 | 19 | 16 | 9 | 13 | 0 |
| Silver | <2-mm | 4.2 | 142 | 0 | 16 | 0 | 13 | 0 |
| Vanadium | <2-mm | 7.8 | 142 | 142 | 16 | 16 | 13 | 13 |
| Zinc | <2-mm | 46 | 142 | 142 | 16 | 16 | 13 | 11 |

## Notes:

For decision units (DUs) with field split and triplicate samples, summary statistics are based on the average of results for the DU. Nondetected values (NDs) are included as half the reporting limits (RLs).
${ }^{\text {a }}$ Ecolgocial soil screening level (Eco-SSL) values are presented in the quality assurance project plan (QAPP) (Exponent et al. 2014), except as noted, and are the lowest of the screening levels adopted by EPA for plants, soil invertebrates, birds, and mammals (USEPA 2010a).
${ }^{\mathrm{b}}$ The Eco-SSL for manganese was not presented in Table A7-2 of the QAPP but is referenced in USEPA (2007b).
ADA - aerial deposition area
$\mathrm{mg} / \mathrm{kg}$ - milligram per kilogram
RFDA - relict flood plain deposition area
WSDA - wind blown sediment deposition area










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dw - dry weight
J - estimated value
$\mathrm{mg} / \mathrm{kg}$ - milligram per kilogram ADA - aerial deposition area For decision units (DUs) with field spit and triplicate samples, summary statistics are based on the average of results for the DU. Nondetected values (NDs) are inctiad (2014) and are the lowest of the screening levels adopted by EPA for plants, soil invertebrates, birds, and mammals (USEPA 2010a). Bold and shaded cells indicate concentrations greater than the ecolgocial soil screening level (Eco-SSL)
Averaged results have three significant figures applied.

a pue＇כ ‘a＇$\forall$ seaxe feuo！！！ For decision units（DUs）with field split and triplicate samples，summary statistics are based on the average of results for the DU．Nondetected values（NDs）are included as half the reporting limits（RLS）．
${ }^{\circ}$ Eco－SSL values are presented in the quality assurance project plan（QAPP）（Exponent et al．2014）and are the lowest of the screening levels adopted by EPA for plants，soil invertebrates，birds，and mammals（USEPA 2010a）．
dw－dry weight
－estimated value


| 089 | †z |  | $\downarrow \underbrace{\circ} 0$ |  | LOE＇0 | 8．11 | 082 |  | 8 E ¢ | $6 \cdot 2 \varepsilon$ | SL＇t | と＇ャレ |  | SL＇t |  | ع9て＇0 | $r$ | عاट | ガヤレ | $\Gamma$ | โ6＇$\varepsilon$ | 800－0．4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0092 | でてZ |  | G9\％ |  | $29^{\circ}$ | ع．1 | LOS |  | ع $๕$ | $08 \tau$ | 66.9 | 9.91 |  | L8＇t |  | ャで0 |  | てぃて | 811 | $\Gamma$ | LL＇$\varepsilon$ | 200－미ํ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | व̇צ্ |
| $97 \varepsilon$ | †＇98 |  | 9 ${ }^{\circ} 0$ |  | s\％ | $1 \cdot 61$ | 922 |  | S0S | 2ヵT | 2L＇8 | ع＇g |  | T＇0才 |  | 99＇0 |  | 891 | $\varepsilon z$ | r | $9 \varepsilon^{\prime} \mathrm{Z}$ | 800－つゝy |
| T9\％ | 6.88 |  | 9200 |  | 90 | $\varepsilon \cdot L$ | LLZ |  | 695 | SIT | عL＇8 | $\varepsilon \cdot 92$ |  | ع9＇8 | $r$ | $\angle 9^{\circ}$ |  | 092 | ¢＇81 | $r$ | 96.2 | L00－О प्र |
| 乙โ๕ |  |  | LL＇0 |  | 切0 | $9 \cdot 6$ | 982 |  | 29t | 9ZT | 95＇8 | T＇92 |  | $68 \cdot 8$ | $r$ | sco |  | 2＜1 | て＇sz | $r$ | 6T＇Z | 900－Оэy |
| てعt | 6.28 |  | \＆960 |  | ع2900 | 902 | $68 \varepsilon$ |  | 909 | tSI | 21．6 | 6.92 |  | 68.6 |  | $\angle 89^{\circ} 0$ |  | S02 | †＇ṫ | $r$ | โع＇£ | 900－Оృу |
| $00 \varepsilon$ | て＇も |  | $99^{\circ} 0$ |  | で「0 | ع＇81 | Lヵ¢ |  | $\varepsilon\llcorner\varepsilon$ | †＇69 | 61．8 | $\varepsilon 乙$ |  | 92＇8 |  | ぐ0 |  | GSt | でく1 | r | โع＇乙 | ャ00－つ」у |
| 9 99 | $\varepsilon ' z \varepsilon$ |  | LO＇L |  | S9L＇0 | て＇61 | 9 tr |  | $08 L$ | S\＆I | L68 | ぐして |  | $\varepsilon$ ¢ $\dagger$ ¢ |  | $8{ }^{\circ} 0$ |  | てL乙 | $\varepsilon \cdot 62$ | $r$ | $\varepsilon て ' G$ | 800－О建 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | ОЈצ্ |
| tot | †＇92 | $\Gamma$ | $81^{\circ}$ | $\Gamma$ | 21.0 | 9.81 | zzع |  | L＇08 | $\downarrow$ ¢ $\downarrow$ | ع1＇9 | 9．81 | $r$ | ${ }^{\text {88＇L }}$ |  | $2 \varepsilon^{\circ} 0$ |  | 981 | $L$ | $r$ | ${ }^{1} 60$ | $800-8$ पy |
| $\downarrow$ ャะ | $\downarrow \varepsilon$ | $r$ | Lで0 |  | $8 t^{\circ} 0$ | 6.61 | GZt | $r$ | 08T | LOZ | $60 \cdot 8$ | $\dagger \vdash$ ¢ | $r$ | โヵ゙と |  | ع8t\％ |  | † 21 | 811 | $r$ | $9 \varepsilon$＇z | $800-8$ पу |
| 022 | でしદ |  | $88^{\circ}$ |  | ts＇0 | く＇もて | 867 |  | 102 | ع＇92 | 168 | て\＆ |  | L＇E | $r$ | $19^{\circ}$ |  | $\angle \varepsilon 乙$ | 6 Zl | $r$ | Ls＇z | 200－8ํㅣ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 8コצ |
| OTLL | s＇zع | $\Gamma$ | 19．1 |  | 6T＇T | 9 ZL | 0szt |  | $0 \angle 9$ | LZS | $\varepsilon$ ¢＇si | †゙L¢ | $\Gamma$ | S9＇E |  | 98\％ |  | 809 | 6 6t | r | T＇TI | S00－ ¢ $^{\text {¢ }}$ |
| $0+98$ | でも |  | เて＇Z |  | โ9＇ธ | でて1 | 0881 |  | 822 | 892 | S．6I | $9 \cdot \varepsilon \varsigma$ |  | ع8＇乙 | $r$ | で0 |  | 059 | 21 | r | $6 . \downarrow \tau$ | ャ00－$\ddagger$ ¢ |
| 0LZs | でโ |  | くが |  | โT＇I | じカ｜ | 0 OT |  | 29¢ | 29t | ¢＇EL | て＇s¢ |  | S0＇s | $r$ | ャ¢0 |  | †t9 | c゙ゅt | $r$ | て＇ZI |  |
| 0092 | 6．1દ | $r$ |  |  | $9 \mathrm{t}^{\text {＇T }}$ | s．1t | OOSI |  | ¢82 | 995 | SI | 9•88 | $r$ | $85^{1} \varepsilon$ |  | ¢980 |  | 9 99 | 186 | $r$ | $9.0 \tau$ | 200－ 7 ¢y |
| 0089 | 8＇† |  | $89^{\circ}$ |  | S2＇I | 8 ＇tr | OZII |  | LSt | 9TS | $8 \cdot \varepsilon \tau$ | て＇LE |  | L6＇t | $r$ | $98^{\circ}$ |  | 809 | 8.91 | $r$ | L＇t | 100－$\forall \pm$ ¢ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | $\forall \pm$ ¢ |
| 97 | $8 \stackrel{1}{ }$ |  | Z＇t |  | $29^{\circ} 0$ | $8 \varepsilon$ | $0 Z 2$ |  | II | 82 | $\varepsilon I$ | 92 |  | $98^{\circ} 0$ |  | $I Z$ |  | $0 \varepsilon \varepsilon$ | 81 |  | LZ＇O | ${ }_{q}($（бy／6u） 7 7S－0才ヨ |
| งu！${ }^{\text {a }}$ | un！peue $\wedge$ |  | 1ə＾！！ |  | un！uәəอS | ｜อYग！ N | әsəue6uew |  | ${ }^{\text {peor }}$ | ${ }^{\text {addo }}$ | Heqo | un！шохч |  | un！upes |  | un！｜｜Kん20g |  | un！ueg | ग！ |  | Kuou！̣uy | l！un uolsivod |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

$\mathrm{Jg} / \mathrm{kg}$－milligram per kilogram ${ }^{6}$ Eco－SSL values are presented in the quality assurance project plan（QAPP）（Exponent et al．2014）and are the lowest of the screening levels adopted by EPA for plants，soil invertebrates，birds，and mammals（USEPA 2010a）．
dw－dry weight
 Bold and shaded cells indicate concentrations greater than the ecolgocial soil screening level（Eco－SSL）

| $0 \downarrow$ T | †＇\＆z | $80^{\circ} 0$ | $r$ | LLL＇0 | r | 1.01 | STt | $r$ | 8 t T | 1－21 | $9{ }^{\prime} \downarrow$ | $r$ | ع＇21 | โt＇$¢$ |  | カガ0 | $6 \downarrow 1$ | 9．1t | $\Gamma$ | عL＇z | L00－ヨコW |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LSI | 6＇z2 | $90^{\circ} 0$ | r | 610 | $r$ | ع01 | Lで | $r$ | ¢ ¢ | L＇21 | \＆t＇t | r | 8．1． | $9 \underbrace{\prime} \varepsilon$ |  | Lヵ0 | LSt | $9 \cdot 1$ | r | $10 \%$ | $900-\ni \exists \mathrm{W}$ |
| LSI | と＇もて | 100 |  | 120 | $r$ | 801 | でt | $r$ | $68 \tau$ | ぐゅ！ | ＋0＇s | $r$ | †で | ¢ $\varepsilon^{\prime} \varepsilon$ |  | 190 | 691 | ＜＇E1 | r | 96.1 | S00－3コW |
| SLI | ぐわて | $80^{\circ} 0$ |  | ャで0 | $r$ | 1．ド | 08t | $r$ | ZLI | g．st | $61 \cdot 9$ | $r$ | でヤト | LL＇$\varepsilon$ |  | $99^{\circ} 0$ | $6 \angle 1$ | 91 | $r$ | $69^{\prime} \mathrm{Z}$ | เ00－3コW |
| toz | 8 8て | $1{ }^{\circ} \mathrm{O}$ |  | 920 | $r$ | 8.01 | †¢¢ | $r$ | $96 \pm$ | 6 Sb | ャで¢ |  | 8．11 | $68 \cdot \downarrow$ |  | Ls 0 | 681 | $9 \times 1$ | $r$ | 26.2 | 800－ЭコW |
| $66 \tau$ | く＇sz | $1{ }^{\circ} \mathrm{O}$ |  | てで0 | $\Gamma$ | H | LtG | $r$ | 697 | ع．91 | $61 \cdot 9$ | r | く．1 | 6T＇t |  | ¢ $9^{\circ} 0$ | L81 | 9 Gl | $r$ | s＇z | 200－ヨコW |
| 8 SZ | 6.58 | Stro |  | $8 \mathrm{C}^{\circ} 0$ | $r$ | ガト | 乙\＆ऽ | $r$ | $9 \varepsilon 乙$ | 6.21 | St＇s | － | $\varepsilon 1$ | ［0＇s |  | $6 \downarrow^{\circ} 0$ | L81 | $\varepsilon \cdot L$ | $r$ | 82＇$\varepsilon$ | 100－3 ${ }^{\text {W }}$ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 Peg stely snorew |
| †＇St | ＋＇8L | 200 | ก | $61^{\circ} 0$ |  | ع＇01 | L0E |  | 16.4 | 1．1！ | $88^{\prime} \downarrow$ |  | 6.1 | $81^{\circ}$ | $\Gamma$ | ャ®0 | $1 \cdot 96$ | $89^{\circ} \mathrm{S}$ | r | $\angle 10$ | 200－s80 |
| St | 6＇zz | 800 | $r$ | $80^{\circ} 0$ |  | 9 Zl | 908 |  | 91.8 | L＇21 | 8＜＇9 |  | St | $81^{\circ}$ | r | $\angle \varepsilon^{\circ} 0$ | StI | $80^{\circ} 9$ | $\Gamma$ | $61^{\circ}$ | 100－S83 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | －s yrerg e！qumios |
| 69 | $L 2$ | 900 | $r$ | $80^{\circ} 0$ |  | 9＇91 | L8E |  | て＇zT | 91 | 91． |  | 8．81 | $\varepsilon{ }^{\circ}$ | $r$ | $68^{\circ} 0$ | 8\＆1 | 97＇8 | r | 920 | 700－NGO |
| 8＇zs | でわて | L9t0 0 | $r$ | ع $880{ }^{\circ}$ |  | L＇z1 | $62 \varepsilon$ |  | 8.01 | 8 － | St＇9 |  | stb | \＆とで0 |  | $\angle ⿻^{\circ} 0$ | カカレ | 28.1 | $r$ | 2820 | 800－N8O |
| LOS | L．sz | ＋0．0 | $r$ | 600 |  | \＆ | $08 \varepsilon$ |  | ع01 | $8 \cdot \varepsilon$ | 29．9 |  | g＇st | \＆で0 | $\Gamma$ | $\angle \checkmark^{\circ} 0$ | 891 | 29.9 | $r$ | Lで0 | 200－N8つ |
| $6.8 t$ | て＇sz | S800 | $r$ | S80\％ |  | 8 Zl | 67 ¢ |  | $6 \mathrm{t}^{\circ} 6$ | 6.21 | 91－9 |  | 6 ¢\％ | 9020 | $r$ | $\vdash^{\circ} 0$ | 221 | ع8＇9 | $r$ | 120 | 100－N8O |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | N पэeәg e！qumioj |
| 9 t | $8{ }^{\circ}$ | 2＇t |  | 2s＇0 |  | $8 \varepsilon$ | 0 Oz |  | IT | 82 | $\varepsilon \tau$ |  | 92 | $98^{\circ} 0$ |  | $\tau Z$ | $0 \varepsilon \varepsilon$ | 81 |  | LZ＇O | ${ }_{9}(6 y / 6 \mathrm{~L}) 7$ 7S－003 |
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|  |  |  |  |  |  |  |  |  | （mp 6y／6 | reuv 亻q uo！ | uoj ！ 10 S |  |  |  |  |  |  |  |  |  |  |



Table 5-12a. Summary of Metals Data from < 149- $\mu \mathrm{m}$ Fraction Compared with Available Human Health Screening Levels

| Analyte | Human Health Soil Screening Level ${ }^{\text {a }}$ $(\mathrm{mg} / \mathrm{kg})^{\text {a }}$ | ADA |  | RFDA |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total Number of DUs | Number of DUs > Screening Level | Total Number of DUs | Number of DUs > Screening Level |
| Aluminum | 77,400 | 142 | 0 | 16 | 0 |
| Antimony | $31.3^{\text {b }}$ | 142 | 0 | 16 | 0 |
| Arsenic ${ }^{\text {c }}$ | $9.33^{\text {b,d }}$ | 142 | 68 | 16 | 8 |
| Barium | 15,300 | 142 | 0 | 16 | 0 |
| Beryllium | 156 | 142 | 0 | 16 | 0 |
| Cadmium | 70.3 | 142 | 0 | 16 | 0 |
| Chromium | 117,000 | 142 | 0 | 16 | 0 |
| Cobalt | 23.4 | 142 | 0 | 16 | 0 |
| Copper | 3,130 | 142 | 0 | 16 | 0 |
| Iron | 54,800 | 142 | 0 | 16 | 0 |
| Lead ${ }^{\text {e }}$ | 400 | 142 | 21 | 16 | 2 |
| Manganese | 1,830 | 142 | 0 | 16 | 0 |
| Mercury | $24^{\text {b }}$ | 142 | 0 | 16 | 0 |
| Molybdenum | 390 | 142 | 0 | 16 | 0 |
| Nickel | 1,550 | 142 | 0 | 16 | 0 |
| Selenium | 391 | 142 | 0 | 16 | 0 |
| Silver | 391 | 142 | 0 | 16 | 0 |
| Thallium | 0.782 | 142 | 0 | 16 | 0 |
| Vanadium | 394 | 142 | 0 | 16 | 0 |
| Zinc | 23,500 | 142 | 0 | 16 | 0 |

## Notes:

For decision units (DUs) with field split and triplicate samples, summary statistics are based on the average of results for the DU. Nondetected values (NDs) are included as half the reporting limits (RLs).
${ }^{\text {a }}$ Screening level values are from Syracuse Research Corporation (SRC) (2013) and presented in the quality assurance project plan (QAPP) (Exponent et al. 2014).
${ }^{\mathrm{b}}$ The screening levels for antimony, arsenic, and mercury were adjusted to reflect changes to the default values for those metals as discussed by SRC when developing screening levels for use in EPA's subsurface sediment screen (SRC 2013).
${ }^{c}$ Arsenic concentrations adjusted for EPA's default relative bioavailability (RBA) of 60 percent arsenic in soil (USEPA 2012b).
${ }^{d}$ The human health screening level for arsenic is based on the 2012 default residential soil screening level for a 1 in 1 million risk level (USEPA 2012b) plus an estimate of the concentration of arsenic in natural background ( $9 \mathrm{mg} / \mathrm{kg}$ ).
${ }^{e}$ Lead concentrations adjusted for the ratio of site-specific RBA to EPA's default RBA, see Table 5-5.
ADA - aerial deposition area
$\mathrm{mg} / \mathrm{kg}$ - milligram per kilogram
RFDA - relict flood plain deposition area


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| 009¢ $8 乙$ | t68 | 28.10 | ${ }^{168}$ | ${ }^{168}$ | oss' $\downarrow$ | 068 | ャ | 088 ' | 00t | 008'ts | $08+$ ' $\varepsilon$ | †¢ | 000'LL | $\varepsilon 0<$ | 991 | 008'st | $68^{6}$ | $\varepsilon$ ' $\varepsilon$ | 00t'Ll | $\left.{ }_{\mathrm{q}}{ }^{(\text {( }} \mathrm{y} / \mathrm{\sigma} \mathrm{\omega}\right)$ <br>  ц„еән иешин |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
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$\mathrm{mg} / \mathrm{kg}$ - milligram per kilogram
RFA, RFB, RFC, RFD - relict flood plain depositional areas A, B, C, and D dThe human health screening level for arsenic is based on the 2012 default residential soil screening level for a 1 in 1 million risk level (USEPA 2012b) plus an estimate of the concentration of arsenic in natural background ( $9 \mathrm{mg} / \mathrm{kg}$ )
e Lead concentrations adjusted for the ratio of site-specific RBA to EPA's default RBA, see Table $5-5$.
dw - dry weight
 Bold and shaded cells indicate concentrations greater than the human health soil screening level.
Averaged results have three significant figures applied.
${ }^{\text {a For decision units (DUs) with field split and triplicate samples, summary statistics are based on the }}$




[^0]:    ${ }^{4}$ Screens of chemicals of interest will identify COPCs to be carried forward into the BERA and HHRA.
    ${ }^{5}$ As specified in the QAPP (Exponent et al. 2014), a No. 100 sieve (mesh size of $149 \mu \mathrm{~m}$ ) was used to collect the soil fraction for analysis of data for the HHRA. The laboratory reports and validation reports refer to this fraction as the $150-\mu \mathrm{m}$ fraction.
    ${ }^{6}$ WSDAs were focused on evaluating risks to ecological receptors (USEPA 2012a). Prior sampling showed that beaches sampled nearest the WSDAs (i.e., Summer Island and Marcus Flats Island for the Marcus Flats WSDAs and Seven Bays for the Columbia Beach WSDAs) had no lead or arsenic concentrations above human health soil screening levels (USEPA 2012c).

[^1]:    ${ }^{7} \mathrm{pH}$ was measured in the bulk soil sample rather than in the $<2-\mathrm{mm}$ soil fraction as specified in the QAPP (Exponent et al. 2014) so that the measurement would not be influenced (i.e., altered) by the soil drying and sieving process.

[^2]:    ${ }^{11}$ A simplified Excel-based calculator for conducting the site-specific bioavailability adjustments is available at the following website: http://www.arche-consulting.be/metals-csa-toolbox/soil-pnec-calculator.
    ${ }^{12}$ At the direction of EPA, IVBA was expanded to include TAL metals and molybdenum rather than just lead as specified in the QAPP. The usability of the IVBA data to assess RBA for metals in the $<2-\mathrm{mm}$ fraction will be determined in conjunction with EPA.
    ${ }^{13}$ Maps have only been prepared for metals with concentrations exceeding human health RBCs.

[^3]:    ${ }^{14}$ An agreement on soil background concentrations for use in the risk assessments has not been reached. Therefore, the comparison with background will be conducted as part of the risk assessments.
    ${ }^{15}$ According to the QAPP, field quality control includes the use of trip blanks. Trip blanks are used to assess the contamination of volatile compounds during sample transport; however, because volatile compounds were not being assessed in this study, trip blanks were deemed unnecessary and, therefore, not included.

[^4]:    ${ }^{16}$ IVBA data for all metals were obtained at the request of EPA and may be used to estimate the RBA in the $<2-\mathrm{mm}$-fraction for use in the BERA.

[^5]:    ${ }^{17}$ A 500-m no-sample buffer zone was established for mine sites within the ADAs that were sampled as part of the assessment detailed in the START-2 report (START-2 2002). A 100-m nosample buffer was established around the other known mine sites in the study area, including those identified as "producer," "past producer," "occurrence," "prospect," or "unknown."

[^6]:    ${ }^{18}$ Archaeological monitoring was conducted by professional archaeologists meeting the Secretary of Interior's Professional Qualification Standards, as outlined in 36 Code of Federal Regulations Part 61.
    ${ }^{19}$ Only 15 increments were collected from ADA-101 because the terrain was too steep to collect all 30 increments; see Appendix A.

[^7]:    ${ }^{20}$ The total number of samples collected using ICS methods does not include the field split samples, which were prepared in the laboratory.

[^8]:    ${ }^{21}$ A simplified Excel-based calculator for conducting the site-specific bioavailability adjustments is available at the following website:
    http://www.arche-consulting.be/metals-csa-toolbox/soil-pnec-calculator.
    ${ }^{22}$ EPA default RBA $=60 \%$ (USEPA 2007a); empirical lead soil concentrations are multiplied by this ratio before comparison to the human health soil screening value to account for differences in bioavailability relative to the screening value.

[^9]:    ${ }^{23}$ The numbers of qualified samples presented in the tables (obtained from the project database) do not include laboratory QC samples and are not always consistent with the numbers of qualified samples presented in text (obtained from ESI), which do include laboratory QC samples.

[^10]:    ${ }^{24}$ Lead data were adjusted for site-specific relative bioavailability (RBA) and arsenic data were adjusted for 60 percent soil arsenic oral RBA prior to comparing the data to screening levels.

[^11]:    ${ }^{25}$ The QAPP did not specify a quality objective for grain size RPDs. Therefore, field QC summaries for grain size are based on triplicate sample RSDs only.

[^12]:    ${ }^{26}$ Eco-SSLs exist for metals that are typically present as cations and can form complexes with inorganic material in soil.
    ${ }^{27}$ Metals with Eco-SSLs are antimony, arsenic, barium, beryllium, cadmium, chromium, cobalt, copper, lead, manganese, nickel, selenium, silver, vanadium, and zinc.

[^13]:    ${ }^{28}$ The following metals were used for the comparisons of screening levels with the $<149-\mu \mathrm{m}$ fraction of soil: aluminum, antimony, arsenic, barium, beryllium, cadmium, chromium, cobalt, copper, iron, lead, manganese, molybdenum, nickel, selenium, silver, thallium, vanadium, and zinc.
    ${ }^{29}$ The human health screening level for lead includes a default RBA adjustment of 60 percent. To ensure appropriate comparison of upland soil lead concentrations to the lead screening level, soil

[^14]:    ${ }^{30}$ The WSDAs were not evaluated for human health because WSDA sampling was focused on evaluating risks to ecological receptors (USEPA 2012a). Prior sampling showed that the beaches sampled nearest the WSDAs (i.e., Summer Island and Marcus Island for the Marcus Flats WSDAs and Seven Bays for the Columbia Beach WSDAs) had no lead or arsenic concentrations above human health soil screening levels (USEPA 2012c).
    ${ }^{31}$ Note that the QAPP specified only IVBA analysis for lead. However, IVBA was performed for all TAL metals and molybdenum at EPA's request (see Appendix B).

[^15]:    ${ }^{\text {a }}$ Coordinates were calculated as the mean of the increment coordinates. Coordinates for decision units sampled in triplicate are from triplicate ' A ' ADA - Aerial deposition area

    EPA - U. S. Environmental Protection Agency - not sampled

    QC - quality control

[^16]:    
    the analysis of lead. However, per EPA request, IVBA analysis was later expanded to include all TAL metals.
    ${ }^{\mathrm{b}}$ No $<149-\mu \mathrm{m}$ fraction was prepared for samples collected from the wind blown sediment deposition areas (WSDAs).
    ADA - Aerial deposition area
    CEC - cation exchange capacity - incring
    ns - not sampled
    RFDA - relict flood plain deposition area
    RFA, RFB, RFC, RFD - relict flood plain depositional areas A, B, C, and D
    TOC - total organic carbon

